Freshwater Macroinvertebrate Fauna Biological Monitoring Programme Annual State of the Environment Monitoring Report 2012-2013

Technical Report 2013-48
(and Report CF557)

ISSN: 0114-8184 (Print)
ISSN: 1178-1467 (Online)
Document No: 1283508 (Word)
Document No: 1317198 (Pdf)

Taranaki Regional Council
Private Bag 713
STRATFORD

February 2014

Executive summary

Section 35 of the Resource Management Act requires local authorities to undertake monitoring of the region's environment, including land, air, and fresh and marine water quality. The Taranaki Regional Council initiated the freshwater biological component of the State of Environment Monitoring (SEM) programme for Taranaki in the 1995-96 monitoring year. The macroinvertebrate component was separated from the microfloral component in the 2002-03 year. The latter programme was broadened to incorporate recently-developed techniques and is reported separately.

This report covers the 2012-2013 monitoring year. Biological surveys were performed in spring (October 2012 through to November 2012) and summer (February to early March 2013), during a late summer low flow period. Each seasonal survey assessed the macroinvertebrate communities at 57 sites in 25 rivers and streams. The Hangatahua (Stony) River was selected as a river with high conservation value and the Maketawa Stream was identified in the Regional Freshwater Plan for its regionally important recreational value. The Waitara, Manganui, Patea, Waiwhakaiho and the Mangaehu Rivers were chosen as examples of waterways with large catchments and multiple human impacts. The Waingongoro River was included in the programme as a river under intensive usage with more recent wastes diversions out of the river, and the Waiongana Stream as a stream from which there is a major water abstraction. The Timaru, Mangaoraka, Waiokura (added in 2007) and Punehu Streams were included as streams within primary agricultural catchments. The Kaupokonui River, Mangorei Stream and Waimoku Stream were selected to monitor the progress of riparian planting in these catchments. These catchments had been targeted in management policies for riparian planting initiatives. The Katikara and Kapoaiaia streams are western Taranaki streams also targeted for riparian planting initiatives, which have been part of the monitoring programme for the thirteenth time this year. The Tangahoe River was included in 2007 to monitor land use changes in an eastern hill country catchment. The Kurapete Stream was added to the programme as an example of a small seepage ringplain stream where significant improvements to a major point source discharge have been implemented. The Waiau Stream is an example of a northern lowland catchment. The Mangawhero and Mangati Streams were selected as examples of small, degraded streams. The Huatoki Stream was selected as an example of a stream influenced by urbanisation and also in part by riparian vegetation while the Herekawe Stream, on the western outskirts of the New Plymouth urban area (with a lengthy consent monitoring record), has been added in order to monitor the impact of recent community walkway planting initiatives.

For sites located lower in catchments the proportion of 'sensitive' taxa in the macroinvertebrate communities generally have been lower in summer than in spring, coincident with lower flows and some increase in smothering of habitats by more widespread algal growth within rivers and streams in summer. The proportion of 'sensitive' taxa in the macroinvertebrate communities has declined down the length of the waterways which was reflected in the deterioration in generic stream 'health' from 'very good' in the upper reaches though 'good' in mid-reaches to 'fair' to 'good' in the lower reaches. Generally, sites in 2012-2013 exhibited the typical summer trend of decreased scores, more particularly at mid and lower reach sites, where long term data have indicated lower median summer scores by 4 and 5 units respectively. Overall, median seasonal MCI scores were 10 (spring) to 3 (summer) units higher than historical median scores during the 2012-

2013 period with spring scores significantly higher than both summer and historical median scores.

A number of sites (10 in mid-reaches and 10 in lower reaches) recorded new historical maximum MCI scores, while no decreases in minimum scores were recorded in the 2012-2013 period. These improvements were particularly apparent within the Punehu, Kaupokonui, Kapoaiaia, and Waiokura Streams and in the Waingongoro River. Increases in overall (eighteen-year) median scores resulted at 18 sites (mainly in mid-reaches) while small decreases were recorded at two upper reach and one mid reach sites.

The macroinvertebrate communities at all ringplain sites have also been assessed in terms of predictive relationships recently established for ringplain streams taking into account altitude and/or distance from the National Park. Evaluations of generic and predictive stream 'health' have also been performed and assessments made for all sites in relation to River Environment Classification (REC) predictions.

The trends through time have been evaluated and will continue to be assessed on an annual basis as the SEM programme continues. Only eight sites have shown temporal deterioration over the eighteen year period (several as a result of headwater erosion effects) but no statistically significant very strong temporal deterioration in MCI scores has been found at any site. Twenty-one sites have shown very strong improvements and a further five sites, strong improvement, all but one of which were of ecological significance. Proportionately fewer of these sites were located in the lower reaches of ringplain catchments where the macroinvertebrate communities are very 'tolerant' of the cumulative impacts of organic enrichment. Significant temporal enhancement of (predominantly 'fair') biological stream 'health' at the lowest sites is unlikely to be detected until habitat improvements occur by way of substantial catchment-wide initiatives such as riparian planting and diversion of point source surface water dairy treatment ponds systems wastes discharges to land irrigation.

Substrate instability and sedimentation caused by extensive headwater erosion events in recent years have affected the macroinvertebrate communities at upper sites in the Stony River (in particular), Maketawa Stream, Waiwhakaiho River, Katikara Stream and Timaru Stream on occasions in past years. Most of these sites continued to show recovery from these impacts over the 2012–2013 period.

The recommendations for the 2013-2014 monitoring year provide for the freshwater biological component of the SEM monitoring to be maintained by way of a similar macroinvertebrate faunal programme and for temporal trend reporting to be performed annually.

Table of contents

List of tables

List of figures

1. Introduction

The Resource Management Act 1991 (RMA) established new requirements for local authorities to undertake environmental monitoring. Section 35 of the RMA requires local authorities to monitor, among other things, the state of the environment of their region or district, to the extent that is appropriate to enable them to effectively carry out their functions under the Act.

To this effect, the Taranaki Regional Council ('the Council') has established a state of the environment monitoring (SEM) programme for the region. This programme is outlined in the Council's 'State of the Environment Monitoring Procedures Document', which was prepared in 1997. The monitoring programme is based on the significant resource management issues that were identified in the Council's Regional Policy Statement for Taranaki (1994).

The SEM programme is made up of a number of individual monitoring activities, many of which are undertaken and managed on an annual basis (from 1 July to 30 June). For these annual monitoring activities, summary reports are produced following the end of each monitoring year (i.e., after 30 June). Where possible, individual consent monitoring programmes have been integrated within the SEM programme to save duplication of effort and minimise costs. The purpose of annual SEM reports is to summarise monitoring activity results for the year, provide an interpretation of these results, together with an update of trends in the data.

Annual SEM reports act as 'building blocks' towards the preparation of regular regional state of the environment reports. The Council's first, or baseline, state of the environment report was prepared in 1996 (TRC, 1996c), summarising the region's progress in improving environmental quality in Taranaki over the past two decades. The second regional state of the environment report was published (TRC, 2003) and discussed the data gathered over the inaugural five year monitoring period. With the completion of the first ten years of the programme in mid 2005, a report on trends (at 60 sites) in biological stream 'health' was completed (Stark and Fowles, 2006), with a subsequent report focusing on the interpretation of significant trends (TRC, 2006). The third regional state of the environment report published in 2009 (TRC, 2009a) encompassed data from 1995 to 2007 and included trending (at 53 sites) for the twelve year period. Subsequent Annual SEM reports consider trends in stream health for all sites as the data record for each monitoring activity increases with time.

This report summarises the results for the sites surveyed in the freshwater biological SEM programme over the 2012-2013 monitoring year, the eighteenth year of this programme.

2. Monitoring activity

2.1 Introduction

The Council commenced the freshwater biological SEM programme in spring 1995. The 2012-2013 monitoring year was therefore the eighteenth year in which this SEM programme was undertaken. This report presents the results from the sites surveyed in the 2012-2013 monitoring year. The methodology for the programme is described in TRC (1997b) and summarised below.

2.2 Monitoring methodology

The standard '400 ml kick-sampling' technique was used to collect streambed (benthic) macroinvertebrates from various sampling sites in selected catchments in the Taranaki region (detailed in section 2.4 and TRC, 1997b). This 'kick-sampling' technique is very similar to Protocol C1 (hard-bottomed, semi-quantitative) of the New Zealand Macroinvertebrate Working Group (NZMWG) protocols for macroinvertebrate samples in wadeable streams (Stark et al, 2001). Surveys of all sites are normally performed twice during the monitoring year, once during spring (October to December) and once during summer (February and March). Sampling dates are detailed in Table 3.

Samples were preserved with Kahle's Fluid for later sorting and identification under a stereomicroscope according to Taranaki Regional Council methodology using protocol P1 of NZMWG protocols for sampling macroinvertebrates in wadeable streams (Stark et al. 2001). Macroinvertebrate taxa found in each sample were recorded as:

R (rare)	= less than 5 individuals;
C (common)	= 5-19 individuals;
A (abundant)	= estimated 20-99 individuals;
VA (very abundant)	= estimated 100-499 individuals;
XA (extremely abundant)	= estimated 500 individuals or more.

2.3 Environmental parameters and indicators

2.3.1 Taxonomic richness

The number of macroinvertebrate taxa found in each sample is used as an indicator of the richness of the community at each site.

2.3.2 Macroinvertebrate Community Index (MCI)

Stark (1985) developed a scoring system for macroinvertebrate taxa according to their sensitivity to organic pollution in stony New Zealand streams. Highly 'sensitive' taxa were assigned the highest scores of 9 or 10, while the most 'tolerant' forms scored 1. Sensitivity scores for certain taxa have been modified in accordance with Taranaki experience (see TRC, 1997b). By averaging the scores obtained from a list of taxa taken from one site and multiplying by a scaling factor of 20, a Macroinvertebrate Community Index (MCI) value was obtained. The MCI is a measure of the overall sensitivity of macroinvertebrate communities to the effects of organic pollution. More 'sensitive' communities inhabit less polluted waterways.

Ringplain rivers and streams sites' data have also been compared with relevant historical survey data which have been used to establish relationships between MCI scores and altitude and distance from stream/river source (National Park boundary) on the ringplain (Stark and Fowles, 2009). These generic relationships for predicting MCI in ringplain streams/rivers are:

$MCI = 84.43 + 0.102A$ [where A= altitude (masl)]; and
$MCI = 131.72 - 25.83 \log_{10} D$ [where D = distance from source (km)]

and have been based upon more than 2400 TRC surveys of about 300 ringplain 'control' sites over the period from 1980 to 2008.

Rates of MCI change have been calculated for each river or stream, where more than a single site has been surveyed, based upon updated G.I.S. river/stream length information incorporated into the Council's 'SITES' database system. [Note: these MCI rates (units/km) of downstream change may differ from earlier reported calculations due to more recent refinements to the accurate determination of river length measurements].

2.3.3 Gradations of biological 'health'

A refinement of Stark's classification (Stark, 1985, Boothroyd and Stark, 2000; and Stark and Maxted, 2007) has been made in order to grade the biological 'health' based upon MCI ranges. This gradation is presented in Table 1.

Table 1 Generic MCI gradation of biological water quality conditions adapted for Taranaki streams and rivers

Grading	MCI	Code	Stark's classification
Excellent	>140		Excellent
Very Good	120-140		
Good	100-119		Good
Fair	80-99		Fair
Poor	60-79		Poor
Very Poor	<60		

This generic adaption is considered to provide more resolution of stream 'health' in the context of more precise upper and lower MCI score bands, than the earlier grading classification. Despite the acknowledgement that the boundaries between gradings may be fuzzy (Stark and Maxted, 2007) these gradings can assist with the assessment of trends in long term temporal data.

Following the establishment of relationships between MCI scores and ringplain stream altitude and distance from source (Stark and Fowles, 2009), biological 'health' may also be graded against predictive values recognising the degree of degradation between the National Park and the coast. These 'predictive' gradings (Table 2) may be applied throughout the length of ringplain streams and range from 'better than expected' through 'expected' to 'worse than expected'.

Table 2 MCI gradation of biological 'health' categories adapted for Taranaki ringplain streams/rivers (based on the relationships for ringplain streams of Stark and Fowles, 2009)

Grading	Differences in MCI scores from predicted values
Better than expected	plus more than 10 units
Expected	+/- 0 to 10 units
Worse than expected	minus more than 10 units

Both systems of grading sites' biological 'health' have been utilised for presentation (see Appendix II) and discussion of data in this report.

2.3.4 Semi Quantitative MCI ($SQMCI_S$)

A semi-quantitative MCI value ($SQMCI_s$) (Stark 1998 & 1999) has also been calculated for the taxa present at each site by multiplying each taxon score by a loading factor (related to its abundance), totalling these products, and dividing by the sum of the loading factors (Stark, 1998, 1999). The loading factors were 1 for rare (R), 5 for common (C), 20 for abundant (A), 100 for very abundant (VA) and 500 for extremely abundant (XA). Unlike the MCI, the $SQMCI_s$ is not multiplied by a scaling factor of 20, so that its corresponding range of values is 20x lower. In this report, the index is used to emphasize the numerical dominance of certain taxa where this is relevant to the interpretation of community structure.

2.4 Trend analysis

State of the environment (SEM) macroinvertebrate data, collected at SEM sites in the region over the eighteen year period (1995-2013) under standard TRC programme protocols, have been statistically analysed for trends using documented methodology (Stark and Fowles, 2006). The significance of any (positive or negative) trends found has been provided for the principal index of stream 'health', the MCI, which was selected as the most appropriate index for use in the assessment of temporal trends in the macroinvertebrate biological quality of regional surface freshwaters (see Stark and Maxted, 2007).

The significance of any site's trend (i.e. the strength of the trend) can be ranked (eg from strongest to weakest) according to the statistical probability of occurrence (p-value), as long as similar numbers of samples were collected for analysis (G McBride, NIWA pers comm) which has been the case with the TRC programme. Following LOWESS [Locally Weighted Scatterplot] smoothing of the temporal MCI data for each site (Stark and Fowles, 2006), the statistical significance of all sites' trends can be ordered (from strongest to weakest) for all sites, with a short-list of sites provided in terms of the strengths of significant trends. The 'cut-off' point has been chosen as those sites' trends with probabilities significant at the 5% level (followed by false discovery rate (FDR) analysis). Those sites can be ranked in order beginning at the lowest p-value. This approach is statistically defensible and should identify sites having trends with valid ecological significance. However, a trend may be statistically significant but have no ecological significance or vice versa. The consideration of ecological significance requires the best professional judgment (BPJ) of a freshwater ecologist with knowledge of the region's rivers and streams. It is likely that the strongest trends (lowest p-values) would have the greatest ecological significance.

In relation to the indicator of stream 'health', the MCI, the estimation error for this index is 10.8 units (Stark 1998) for the sampling protocols used by TRC. Therefore although a statistically significant temporal trend may be found for a site's data, if the LOWESS range of MCI scores is less than 11 units, the best professional judgment may eliminate this from a list of significant results. Also, to place these trends in perspective, each site may be assessed against graduations (bands of MCI values) of stream health. In this instance, Stark's (1985) categories have been modified (using BPJ) as illustrated in Section 2.3.2 above.

2.5 Assessment of significant differences

When the same number of replicate samples are collected per site, the detectable difference method may be used to assess the significance of MCI score differences. Stark (1998) provides statistically significant detectable differences for the protocols used by TRC (10.8 MCI units). Between season and long term median MCI scores and/or taxa richnesses may also be compared using t-tests (Stark and Maxted, 2007).

2.6 Site locations

All sites in the freshwater biological SEM programme for the Taranaki region are illustrated in Figure 1 and described in Table 3. The biological programme for the 2012-2013 period involved the continuation of a riparian vegetation monitoring component incorporating five sites in the Kaupokonui River (see Table 3) and five sites in western Taranaki ring plain streams (Katikara Stream and Kapoaiaia Stream). Evaluations of the effects of, and recovery from, extensive erosion in the headwaters of the Waiaua River had been included in this programme. These surveys commenced in December 1998 and the two sites on the Waiaua River were incorporated into the SEM biological monitoring programmes since the initial documentation of the effects and recovery was established. This river has continued to be affected by headwater erosion in recent years. Therefore, the programme was reviewed in 2006 and the Waiaua River excluded from the SEM programme. The Kurapete Stream (upstream and 5.5km downstream of the Inglewood oxidation pond system) has been monitored throughout the SEM period, using the appropriate SEM protocols, and has been included in the programme. Two additional sites in the Waiwhakaiho River catchment were included in 2002-2003 in recognition of the importance of this major catchment.

Table 3 Freshwater biological monitoring sites in the State of the Environment Monitoring programme

Type	River/stream	Site	Site code	GPS location		Distance from Nat Park (km)	Altitude (m asl)	Spring sampling date	Summer sampling date
				E	N				
Conservation	Hangatahua (Stony) R	Mangatete Road	STY000300	1677460	5657823	7.3	160	5 Oct 13	26 Feb 13
	Hangatahua (Stony) R	SH45	STY000400	1674632	5661558	12.5	70	5 Oct 13	26 Feb 13
	Maketawa S	Opp Derby Road	MKW000200	1702192	5656304	2.3	380	14 Nov 12	28 Feb 13
	Maketawa S	Tarata Road	MKW000300	1708784	5665231	15.5	150	14 Nov 12	28 Feb 13
Large catchment/ multiple impacts	Waiwhakaiho R	National Park	WKH000100	1696096	5658351	0	460	28 Nov 12	1 Feb 13
	Waiwhakaiho R	SH3 (Egmont Village)	WKH000500	1698297	5666893	10.6	175	28 Nov 12	1 Feb 13
	Waiwhakaiho R	Constance St (NP)	WKH000920	1695827	2677271	26.6	20	28 Nov 12	1 Feb 13
	Waiwhakaiho R	Adjacent to L Rotomanu	WKH000950	1696587	2678336	28.4	2	28 Nov 12	1 Feb 13
	Mangorei S	SH3	MGE000970	1696094	5671500	21.6	90	28 Nov 12	1 Feb 13
	Manganui R	SH3	MGN000195	1708871	5651282	8.7	330	14 Nov 12	28 Feb 13
	Manganui R	Bristol Road	MGN000427	1711210	5667887	37.9	140	14 Nov 12	28 Feb 13
	Waitara R	Mamaku Road	WTR000850	1708384	5678739	N/A	15	3 Oct 12	28 Feb 13
	Patea R	Barclay Rd	PAT000200	1702620	5646598	1.9	500	1 Oct 12	28 Feb 13
	Patea R	Swansea Rd	PAT000315	17118011	5644382	12.9	300	1 Oct 12	28 Feb 13
	Patea R	Skinner Rd	PAT000360	715919	5644681	19.2	240	1 Oct 12	28 Feb 13
Intensive usage	Waingongoro R	700m d/s Nat Park	WGG000115	1700835	5645086	0.7	540	31 Oct 12	25 Feb 13
	Waingongoro R	Opunake Rd	WGG000150	1705692	5642523	7.2	380	31 Oct 12	25 Feb 13
	Waingongoro R	Eltham Rd	WGG000500	1710576	5634824	23.0	200	31 Oct 12	25 Feb 13
	Waingongoro R	Stuart Rd	WGG000665	1709784	5632049	29.6	180	31 Oct 12	25 Feb 13
	Waingongoro R	SH45	WGG000895	1704042	5618667	63.0	40	31 Oct 12	25 Feb 13
	Waingongoro R	Ohawe Beach	WGG000995	1702531	5617624	66.6	10	31 Oct 12	25 Feb 13
Primary agricultural	Timaru S	Carrington Road	TMR000150	1684423	5659634	0	420	5 Oct 12	26 Feb 13
	Timaru S	SH45	TMR000375	1679509	5665554	10.9	100	5 Oct 12	26 Feb 13
	Mangaoraka S	Corbett Road	MRK000420	1702538	5676320	N/A	60	5 Oct 12	12 Feb 13
	Punehu S	Wiremu Rd	PNH000200	1687323	5637020	4.4	270	4 Oct 12	15 Feb 13
	Punehu S	SH45	PNH000900	1677946	5627786	20.9	20	4 Oct 12	15 Feb 13
	Waiokura S	Skeet Rd	WKR000500	1698807	5628892	N/A	150	4 Oct 12	15 Feb 13
	Waiokura S	Manaia Golf Course	WKR000700	1697636	5622019	N/A	70	4 Oct 12	15 Feb 13
Eastern hill country	Tangahoe R	Upper Valley	TNH000090	1725340	5626101	N/A	85	30 Oct 12	7 Mar 13
	Tangahoe R	Tangahoe Vly Rd bridge	TNH000200	1719126	5622681	N/A	65	30 Oct 12	7 Mar 13
	Tangahoe R	d/s railbridge	TNH000515	1715751	5612470	N/A	15	30 Oct 12	7 Mar 13
	Mangaehu R	Raupuha Rd	MGH000950	1726300	5639062	N/A	120	1 Oct 12	28 Feb 13
Riparian	Waimoku S	Lucy's Gully	WMK000100	1681324	5666240	0	160	5 Oct 12	26 Feb 13
	Waimoku S	Beach	WMK000298	1681725	5669851	4.0	1	5 Oct 12	26 Feb 13
Western	Katikara S	Carrington Road	KTK000150	1683566	5657855	0	420	16 Nov 12	4 Mar 13
	Katikara S	Beach	KTK000248	1676597	5667473	18.1	5	16 Nov 12	4 Mar 13
	Kapoaiaia S	Wiremu Road	KPA000250	1678009	5652025	5.7	240	4 Oct 12	4 Mar 13
	Kapoaiaia S	Wataroa Road	KPA000700	1672739	5652272	13.5	140	4 Oct 12	4 Mar 13
	Kapoaiaia S	Cape Egmont	KPA000950	1665690	5652452	25.2	20	4 Oct 12	4 Mar 13
	Kaupokonui R	Opunake Road	KPK000250	1698088	5639231	3.3	380	4 Oct 12	15 Feb 13
	Kaupokonui R	U/s Kaponga oxi ponds	KPK000500	1698609	5634423	9.2	260	4 Oct 12	15 Feb 13
Southern	Kaupokonui R	U/s Lactose Co.	KPK000660	1697613	5629791	15.5	170	4 Oct 12	15 Feb 13
	Kaupokonui R	Upper Glenn Road	KPK000880	1693026	5622705	25.7	60	4 Oct 12	15 Feb 13
	Kaupokonui R	Near mouth	KPK000990	1691209	5620444	31.1	5	4 Oct 12	15 Feb 13
Small degraded ('poor') catchment	Mangati S	D/s railway line	MGT000488	1700095	5678043	N/A	30	3 Oct 12	12 Feb 13
	Mangati S	Te Rima Pl, Bell Block	MGT000520	1699385	5679103	N/A	20	3 Oct 12	12 Feb 13
	Mangawhero S	u/s Eltham WWT Plant	MWH000380	1712475	5633431	N/A	200	31 Oct 12	25 Feb 13
	Mangawhero S	d/s Mangawharawhara S	MWH000490	1710795	5632738	N/A	190	31 Oct 12	25 Feb 13
	Kurapete S	u/s Inglewood WWT Plant	KRP000300	1705087	5665510	N/A	180	5 Oct 12	1 Feb 13
	Kurapete S	6 km d/s Inglewood WWTP	KRP000660	1709239	5667481	N/A	120	5 Oct 12	1 Feb 13
Urbanisation	Huatoki S	Hadley Drive	HTK000350	1693349	5671486	N/A	60	16 Nov 12	4 Mar 13
	Huatoki S	Huatoki Domain	HTK000425	1693041	5673404	N/A	30	16 Nov 12	4 Mar 13
	Huatoki S	Molesworth St	HTK000745	1692800	5676424	N/A	5	16 Nov 12	4 Mar 13
	Herekawe S	Centennial Drive	HRK000085	1688283	5674972	N/A	5	5 Oct 12	12 Feb 13
Northern lowland catchment	Waiau S	Inland North Road	WAI000110	1714587	5680018	N/A	50	5 Oct 12	12 Feb 13
								5 Oct 12	12 Feb 13
Major abstraction	Waiongana S	SH3a	WGA000260	1705159	5669554	16.1	140	5 Oct 12	6 Mar 13
	Waiongana S	Devon Road	WGA000450	1704063	5680381	31.2	20	5 Oct 12	6 Mar 13

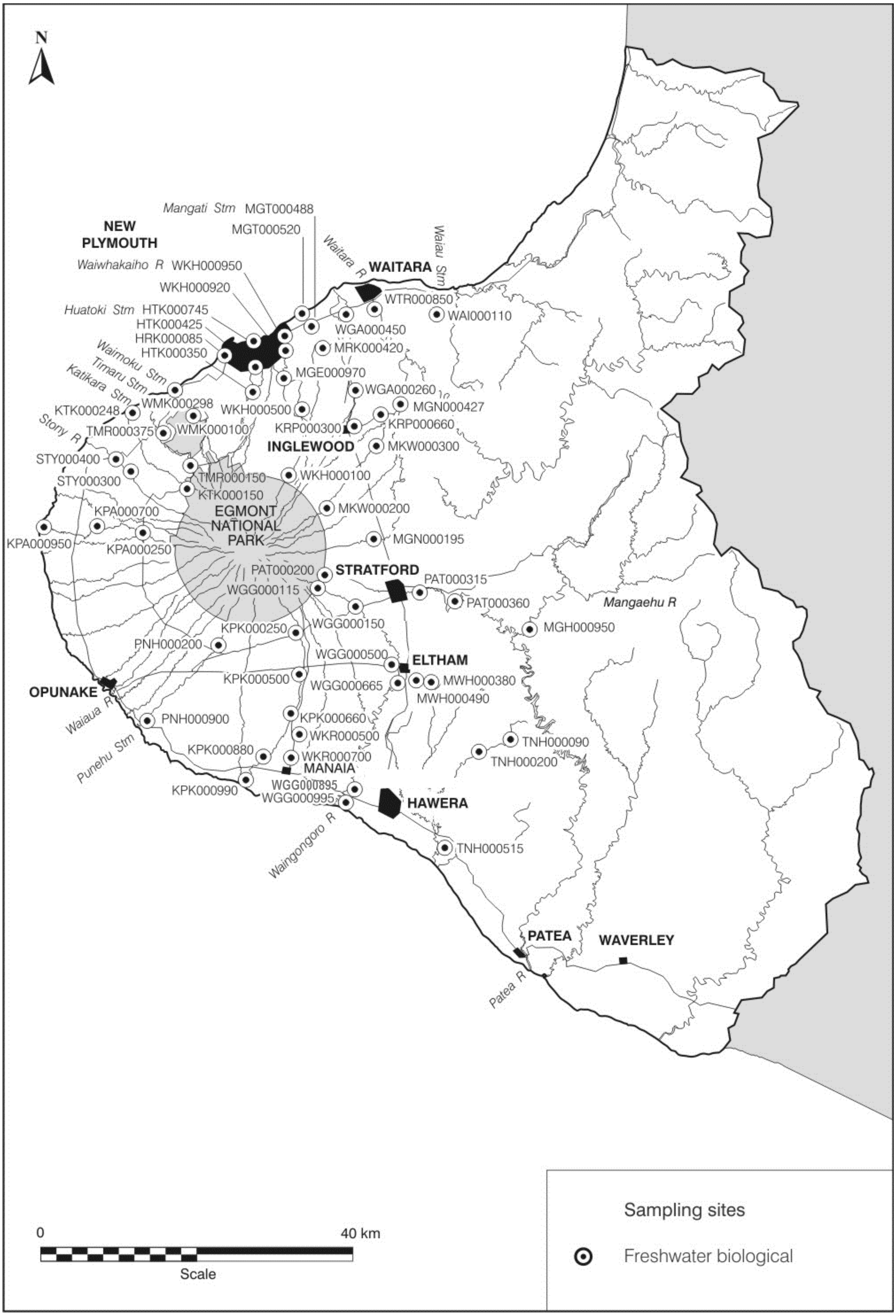


Figure 1 Location of macroinvertebrate fauna sampling sites for the 2011-2012 SEM programme

Two sites in the Maketawa Stream were also added as a result of a commitment to continue the documentation of conditions in this catchment following the investigation of baseline water quality conditions during the 2000-2002 period (Stark, 2003). Three sites in the Tangahoe River were established in the 2007-2008 period for the purposes of monitoring land use changes (aforestation) in an eastern hill country catchment. The two sites in the Waiokura Stream were also added in the 2007-2008 period as a long term monitoring commitment to the collaborative best practice dairying catchment project. One site in the Herekawe Stream (a long term consent monitoring site) was incorporated into the programme in the 2008-2009 period for the purpose of monitoring the local initiatives of walkway establishment and riparian planting of this small catchment on the western outskirts of the New Plymouth urban area.

The Hangatahua (Stony) River was selected for the SEM programme as a waterway of high conservation value. The headwaters of the river are the Ahukawakawa swamp within Egmont National Park, and several tributaries that begin above the tree line on the north-west of Mount Taranaki. Once the river leaves the National Park boundary its catchment becomes very narrow so that it receives little water from surrounding farmland before reaching the sea. This factor and the protection order on the catchment maintains good water quality in the river. However, exceptions occur from time to time after headwater erosion events when sedimentation and scouring of the riverbed may be particularly severe. The sites at Mangatete Road and State Highway 45 are approximately seven kilometres and twelve kilometres downstream of the National Park boundary respectively.

The Timaru and Mangaoraka Streams were chosen for the SEM programme as examples of streams within primary agricultural catchments. The Timaru Stream arises within the National Park boundary, near the peak of Pouakai, in the Pouakai Range. Upon leaving this range, the stream flows along the edge of the Kaitake Range (also part of the National Park) and receives several tributaries that flow through adjacent agricultural land. From the edge of the Kaitake Range, the stream flows north through agricultural land to the sea. Carrington Road crosses the stream within the National Park boundary and State Highway 45 is six kilometres downstream of the confluence with the first farmland tributary. The Mangaoraka Stream rises below the National Park boundary near Egmont Road and flows north through farmland for its entire length before joining the Waiongana Stream near the coast. Corbett Road is 26 kilometres downstream of the source.

The Waiongana Stream was included in the SEM programme as an example of a stream with a major water abstraction. The stream originates within the National Park, near the North Egmont visitor's centre. After crossing the park boundary, it flows north-east through agricultural land to the sea. State Highway 3a crosses the stream fifteen kilometres downstream of the National Park boundary, and the intake for the Waitara industrial water supply is a further five kilometres downstream of that. Devon Road is 30 kilometres downstream of the National Park boundary.

The Waiwhakaiho, Manganui, Waitara, and Mangaehu Rivers were selected for the SEM programme as examples of waterways with large catchments and multiple impacts from human land uses including plantation forestry, rural, urban and industrial.

The Waiwhakaiho River and its headwater tributaries arise above the tree line on the north face of Mount Taranaki. Upon leaving the National Park, the river flows north through agricultural and industrial land for 27 kilometres to the sea. The river passes under State Highway 3 near Egmont Village, nine kilometres downstream of the National Park boundary. The sites at Constance Street and adjacent to Lake Rotomanu are included in the lower Waiwhakaiho River industrial monitoring programme. The site adjacent to Lake Rotomanu has replaced the site immediately downstream of the Mangaone Stream that was used in the 1995-96 State of the Environment monitoring survey. This allows the State of the Environment monitoring programme to better synchronise with the industrial monitoring programme. The Mangorei Stream is the principal tributary catchment in the lower reaches, downstream of the major abstraction of water for hydroelectric and community supply purposes.

The source of the Manganui River is situated above the tree line on the eastern slopes of Mount Taranaki. After leaving the National Park, the river flows east and then north through agricultural land for 44 kilometres before joining the Waitara River. State Highway 3 is eight kilometres downstream of the National Park boundary. At Tariki Road, much of the flow of the Manganui River is diverted through the Motukawa hydroelectric power scheme to the Waitara River. Therefore, except when the Tariki weir is overtopping, most of the water in the Manganui River at Bristol Road (14 kilometres downstream of the diversion) comes from tributaries such as the Mangamawhete, Waitepuke, Maketawa, and Ngatoro Streams. Like the Manganui River, these streams originate high on the eastern slopes of Mount Taranaki. They flow through agricultural land before joining the river. The Maketawa Stream provides a valued trout and native fish habitat. Sites were included in the upper and lower reaches of the stream.

The small Kurapete Stream, which rises as seepage to the west of Inglewood, was included to monitor trends in relation to the removal of the town’s Wastewater Treatment Plant’s discharge from this tributary of the lower Manganui River in 2000. Sites were included upstream and nearly six km downstream of where the discharge was located.

The Waitara River flows south-west and then north-west out of the eastern hill country through a mix of agricultural land and native forest before passing through the town of Waitara and out to sea. It has a different character from the steep ring plain rivers and carries a high silt load. The Mamaku Road site is located six kilometres upstream of the coast above any tidal influence. This site is part of the monitoring programme for the stormwater discharge from the Waitara Valley Methanex plant to the Waitara River.

The Mangaehu River originates in the eastern hill country and flows south-west through agricultural land for most of its length before joining the Patea River, ten kilometres upstream of Lake Rotorangi. Raupuha Road crosses the river less than one kilometre upstream of the confluence with the Patea River.

The Tangahoe River is a smaller eastern hill country catchment which flows through agricultural land, some of which has undergone afforestation in the upper reaches. Fonterra extracts dairy company processing waters in the lower reaches near the coast, south of Hawera township.

The Mangati Stream was chosen for the SEM programme as an example of a small, degraded stream. Only five kilometres in length, the stream rises in farmland and flows north through the Bell Block industrial area and suburbs to the sea. The site downstream of the railway line is upstream of all industrial discharges to the stream. The site at Te Rima Place is located within a suburban park, downstream of all Bell Block industrial discharges. Both sites are part of the Mangati Stream industrial monitoring programme.

The Waimoku Stream originates in Egmont National Park where it flows down Lucy's Gully in the Kaitake Ranges. Once the stream leaves the park it flows through farmland for three and a half kilometres, and through the coastal township of Oakura for about 200 metres, before entering the sea. It was included in the SEM programme in the 1999-2000 monitoring year to monitor the effects of a riparian planting programme in the catchment. Sampling sites are located in Lucy's Gully under native forest, and in Oakura township, about 100 metres upstream of the sea.

The Waiau Stream originates in farmland near Tikorangi, near the Waitara River. It flows for 12.5 km to the sea. The stream was included in the SEM programme in the 1999-2000 monitoring year as an example of a northern lowland catchment. The sampling site at Inland North Road is located in a pasture setting.

The Punehu Stream is representative of a south-western Taranaki catchment subject primarily to intensive agricultural land use with water quality affected by diffuse source run-off and point source discharges from dairy shed treatment pond effluents particularly in the Mangatawa Stream, a small lower reach tributary. No industrial discharges to the stream system are known to occur. Both sites were Taranaki ring plain survey sites (TCC, 1984) and the lower site near the coast remains a NIWA hydrological recording station as a representative basin. The upstream site is representative of relatively unimpacted stream water quality although it lies approximately 2 km below the National Park boundary.

The small seepage fed, ringplain Waiokura Stream drains an intensively dairy-farmed catchment. The Fonterra, Kapuni factory irrigates wastewater within the mid reaches of this catchment. The catchment is the subject of a (five region) collaborative long term study of best practice dairying catchments (Wilcock et al, 2009).

The Patea River rises on the eastern slopes of Mt Taranaki, within the National Park and is a trout fishery of regional significance, particularly upstream of Lake Rotorangi (formed by the Patea dam) in its mid reaches. Site 1 (at Barclay Road) is representative of the upper catchment adjacent to the National Park above agricultural impacts. Site 2 (at Swansea Road), which is integrated with special order consent monitoring programmes, was also a ring plain survey site, and is representative of developed farmland drainage and is downstream of Stratford township (urban run-off, but upstream of the rubbish tip and oxidation pond discharges and the combined cycle power station discharge). Site 3 (at Skinner Road) is an established hydrological recorder station downstream of these discharges and the partly industrialised Kahouri Stream catchment.

The Waingongoro River rises on the south-eastern slopes of Mount Taranaki within the National Park and is one of the longest of the ring plain rivers, with a meandering 67 km of river length from the National Park boundary prior to entering the Tasman Sea at Ohawe Beach. The river is the principal trout fishery in Taranaki and is also utilised for water abstraction purposes and up until mid 2010, received treated industrial and municipal wastes discharges in mid-catchment at Eltham. Site 1 (near the National Park boundary) is representative of high water quality conditions with minimal agricultural impacts. Site 2, six km further downstream (at Opunake Road) represents agricultural impacts, still in the upper reaches of the river while site 3 (at Eltham Road) a further 16 km downstream remains representative of the impacts of farmland drainage and some water abstraction while upstream of the major Eltham point source discharges from a meatworks and the municipal wastewater treatment plant. The meatworks wastewaters were diverted to spring and summer land irrigation in the mid 2000's and treated wastewater subsequently has continued to be irrigated onto farmland in this manner. The Eltham municipal wastes were permanently diverted by pipeline to Hawera in June 2010. The Stuart Road site, a further six km downstream is located below these discharges with a major portion of the meatworks discharge diverted to land irrigation (spring through late summer) since the early 2000's and the Eltham WWTP discharge diverted out fo the catchment by pipeline to the Hawera WWTP in July 2010. A further two sites (SH45 and Ohawe Beach) located 33 km and 37 km downstream of Stuart Road in the intensively developed farmland lower reaches of the catchment. River flow recording sites are located at Eltham Road and SH45.

The Mangawhero Stream is a relatively small, swamp-fed catchment rising to the east of Eltham in the Ngaere Swamp and draining developed farmland. The upper site is located in the mid reaches of the stream upstream of the point source discharge from the Eltham municipal wastewater treatment plant while the lower site is located a further three km downstream, below the Mangawharawhara Stream confluence, near the confluence with the Waingongoro River. Apart from the municipal point source discharge, which was diverted out of the stream in July 2010 (see above), the catchment is predominantly developed farmland.

The Huatoki Stream was sampled as part of the State of the Environment monitoring programme for the first time in the 1997-98 monitoring year. The stream rises one kilometre outside the National Park boundary on the foothills of the Pouakai Range. It flows through agricultural land for 12.5 km to the outskirts of New Plymouth where it enters native forest reserve. The stream flows for four and a half kilometres alongside walkways and beneath the central business district of New Plymouth before entering the sea next to Puke Ariki Landing. Within New Plymouth it flows through a culvert in a flood retention dam and over a small weir in the Huatoki Reserve prior to the business section of the city. Beautification works adjacent to 'Centre City' near the stream mouth (in 2010) involved the creation of a weir and fishpass immediately upstream of the lowest site which subsequently has altered the flow regime at this site and created a run-like habitat rather than the previous riffle habitat.

The Herekawe Stream is a small seepage stream on the western boundary of New Plymouth. It drains a mainly urban catchment and receives stormwater discharges particularly in its lower reaches. Recent completion of a walkway and riparian planting community project now warrants the inclusion of the consent monitoring 'control' site at Centennial Drive for monitoring the effectiveness of these initiatives.

The Kaupokonui River rises on the southern slopes of Mt Taranaki within the National Park. It drains an intensively farmed dairy catchment. The principal point source discharges to the river occur in the mid-reaches from the Kaponga oxidation pond system, and cooling water from NZMP (Kapuni) Ltd. The river has patchy riparian vegetation cover and has been targeted for intensive riparian management initiatives. Site 1 is two and a half kilometres downstream of the National Park boundary and has high water quality, with minor agricultural impacts. Toward the mid-reaches, site 2 (six kilometres further downstream) is subject to some agricultural impacts, but is a short distance upstream of the Kaponga oxidation ponds' system discharge. A further six kilometres downstream, site 3 is upstream of wastes irrigation, cooling water discharges and factory abstraction. The Upper Glenn Road (site 4) is a further 10 km downstream, below all of the factory's activities and is a river flow hydrological recording site. The final site 5, is located near the mouth of the river, 5 km below site 4, upstream of any tidal influence at Kaupokonui beach domain camping ground.

Two western catchments, the Katikara Stream and Kapoaiaia Stream, were included in the programme to monitor trends in relation to riparian planting. Such riparian planting initiatives have been concentrated in certain catchments where current riparian vegetation is poor. The Katikara Stream rises on the western slopes of Mt Taranaki, passing through primarily agricultural land in the relatively short distance to the sea. The Kapoaiaia Stream also rises from Mt Taranaki on the western side but south of the Katikara Stream. The Kapoaiaia Stream drains agricultural land throughout its entire catchment below the National Park boundary, passing through Pungarehu township at SH45 before entering the sea at Cape Egmont. A hydrological telemetry recorder is located at Cape Egmont.

3. Results and discussion

3.1 Flows and water temperature

Hydrological flow recorders continuously monitor water levels in the Mangaoraka, Waiongana, Punehu, and Kapoaiaia Streams, and the Waiwhakaiho, Manganui, Stony, Patea, Mangaehu, Waingongoro, Kaupokonui and Waitara Rivers. Flow conditions can therefore be determined in these watercourses for the period prior to the collection of biological samples. The proximity of previous freshes for each site surveyed, are summarised in Table 4, with flow assessments extrapolated from nearby catchments for sites where flow recorders were not available.

Table 4 Duration since freshes at sampling sites in the 2012-2013 SEM biomonitoring programme

River/stream	Site	Spring survey (days after flow above)		Summer survey (days after flow above)	
		3 x median	7 x median	3 x median	7 x median
Hangatahua (Stony) R	Mangatete Road	17	18	21	21
Hangatahua (Stony) R	SH45	17	18	21	21
Timaru S	Carrington Road	(17)	(18)	(21)	(21)
Timaru S	SH45	(17)	(18)	(21)	(21)
Mangaoraka S	Corbett Road	18	27	7	87
Waiongana S	SH3a	18	18	29	29
Waiongana S	Devon Road	18	18	29	29
Waiwhakaiho R	National Park	11	11	17	17
Waiwhakaiho R	SH3 (Egmont Village)	11	11	17	17
Waiwhakaiho R	Constance St (NP)	11	11	17	17
Waiwhakaiho R	Adjacent Lake Rotomanu	11	11	17	17
Mangorei S	SH3	(11)	(11)	(17)	(17)
Manganui R	SH3	11	11	22	23
Manganui R	Bristol Road	11	11	22	23
Maketawa S	opp Derby Road	(11)	(11)	(22)	(23)
Maketawa S	Tarata Road	(11)	(11)	(22)	(23)
Waitara R	Mamaku Road	15	16	22	23
Mangati S	D/s railway line	(16)	(25)	(7)	(87)
Mangati S	Te Rima Pl, Bell Block	(16)	(25)	(7)	(87)
Waimoku S	Lucy's Gully	(17)	(18)	(21)	(21)
Waimoku S	Beach	(17)	(18)	(21)	(21)
Waiau S	Inland North Road	(18)	(27)	(7)	(87)
Punehu S	Wiremu Rd	17	17	10	10
Punehu S	SH45	17	17	10	10
Patea R	Barclay Rd	14	14	23	23
Patea R	Swansea Rd	14	14	23	23
Patea R	Skinner Rd	14	14	23	23
Mangaehu R	Raupuha Road	13	20	22	132
Mangawhero S	u/s Eltham WWT Plant	(17)	(-)	(105)	(-)
Mangawhero S	d/s Mangawharawhara S	(10)	(18)	(20)	(20)
Waingongoro R	900m d/s Nat Park	10	18	20	20
Waingongoro R	Opunake Rd	10	18	20	20
Waingongoro R	Eltham Rd	10	18	20	20
Waingongoro R	Stuart Rd	10	18	20	20
Waingongoro R	SH45	10	18	20	20
Waingongoro R	Ohawe Beach	10	18	20	20
Huatoki S	Hadley Drive	(13)	(13)	(27)	(107)
Huatoki S	Huatoki Domain	(13)	(13)	(27)	(107)
Huatoki S	Molesworth St	(13)	(13)	(27)	(107)
Kaupokonui R	Opunake Rd	16	18	9	10
Kaupokonui R	U/s Kaponga oxi ponds	16	18	9	10
Kaupokonui R	U/s Lactose Co.	16	18	9	10
Kaupokonui R	Glenn Rd	16	18	9	10
Kaupokonui R	Beach	16	18	9	10
Katikara S	Carrington Road	(13)	(13)	(27)	(27)
Katikara S	Near mouth	(13)	(13)	(27)	(27)
Kapoaiaia S	Wiremu Road	16	17	27	27
Kapoaiaia S	Wataroa Road	16	17	27	27
Kapoaiaia S	Near coast	16	17	27	27
Kurapete S	u/s Inglewood WWTP	(18)	(27)	(18)	(75)
Kurapete S	6km d/s Inglewood WWTP	(18)	(27)	(18)	(75)
Tangahoe R	Upper Valley	(10)	(11)	(29)	(30)
Tangahoe R	Tangahoe Valley Road	(10)	(11)	(29)	(30)
Tangahoe R	d/s railbridge	(10)	(11)	(29)	(30)
Waiokura S	Skeet Road	(16)	(18)	(9)	(10)
Waiokura S	Manaia Golf-Course	(16)	(18)	(9)	(10)
Herekawe S	Centennial Drive	(8)	(18)	(7)	(87)
Herekawe S	Centennial Drive	(8)	(18)	(7)	(87)

NB: ( ) = extrapolation from nearby catchment

Spot water temperatures recorded at each site at the time of sampling during spring 2012 and summer 2013 SEM biomonitoring surveys are summarised in Table 5.

Table 5 Water temperature recorded at the times of SEM biological monitoring surveys

Watercourse	Spring 2012	Summer 2013
Hangatahua (Stony) River	10.1-10.2	14.6-15.6
Timaru Stream	8.1-10.6	14.6-16.4
Mangaoraka Stream	11.6	17.7
Waiongana Stream	10.1-12.1	15.7-19.0
Waiwhakaiho River	10.3-18.7	12.0-21.3
Mangorei Stream	16.2	18.5
Manganui River	9.1-12.4	14.4-18.9
Maketawa Stream	11.0-11.7	12.2-16.5
Waitara River	14.6	21.6
Mangati Stream	14.0-14.7	17.8-18.4
Waimoku Stream	10.4-12.3	13.9-16.1
Waiau Stream	11.3	17.4
Punehu Stream	9.6-12.2	15.6-17.3
Patea River	8.3-11.7	10.9-16.1
Mangaehu River	13.8	20.7
Mangawhero Stream	13.6-15.2	15.3-16.7
Waingongoro River	8.5-16.2	11.3-20.3
Huatoki Stream	12.7-13.8	16.0-16.7
Kaupokonui River	8.3-13.6	12.2-17.7
Katikara Stream	9.9-15.8	13.7-17.1
Kapoaiaia Stream	10.2-13.0	15.3-18.9
Kurapete Stream	10.6-11.2	16.2-17.0
Tangahoe River	13.7-14.7	15.0-15.9
Waiokura Stream	12.1-12.2	14.6-15.9
Herekawe Stream	12.6	17.7

(Note: N/R = not recorded)

3.1.1 Water temperature

Spring 2012

The spring 2012 surveys were mainly undertaken in early spring following relatively short to moderate recessions of one to two weeks after freshes, or delayed toward later spring due to mid-spring freshes. Spring surveys in nearly all streams were conducted from 10 to 18 days after moderate freshes while none were more than 18 days after freshes. Water temperatures ranged from 8.1˚C to 10.4˚C in the upper reaches; 9.1˚C to 11.0˚C in the middle reaches; and from 10.2˚C to 18.7˚C in the lower reaches of streams and rivers at the time of the surveys (Table 5).

Summer 2013

Generally, rivers and streams were in relatively low recession flow following a few January 2013 freshes with drier conditions occurring between early February and mid-March 2013 when all of the surveys were performed. All but four surveys were performed no less than 10 days after significant freshes, with most surveys three weeks or longer after a significant (3x median) fresh.

Water temperatures ranged from 10.9°C to 14.6°C in the upper reaches, 12.2°C to 16.2 °C in the mid reaches, and from 15.6°C to 21.6°C in the lower reaches of streams and rivers at the time of the surveys (Table 5). These ranges tended to be typical of most past summer surveys.

3.2 Macroinvertebrate communities

Lists of the taxa found during spring 2012 and summer 2013 surveys, together with taxa richness, MCI scores and other appropriate indices for each site are tabulated and attached as Appendix I. These results are discussed on a stream by stream basis for the sites and seasons (spring and summer) in which the surveys were conducted. Data from previous surveys are also presented for each site and results to date are illustrated as appropriate.

3.2.1 Hangatahua (Stony) River

Prior to the commencement of the SEM programme (in 1995), three samples had been collected from the site at State Highway 45. During the 1999-2000 monitoring year, an extra survey was performed in July 1999, and an extra site (STY000260, near the end of Saunders Road) was included in all three surveys, in order to closely monitor the recovery of the Stony River following massive sand drifts in the channel. This extra monitoring was not performed in subsequent monitoring years until 2004, following the very heavy rainfall events in late summer.

In the winter of 1996 a massive drift of sand moved down the Hangatahua River and devastated macroinvertebrate communities, following a major erosion event in the headwaters of the river. Few macroinvertebrate taxa were found in the river in the spring of 1996 (Figure 2 and Figure 4). Since then sand has continued to affect the macroinvertebrate communities of the river, although some recovery was observed in the communities in March and November 1997, January and February 1999, late 2000, and again in 2002-2003. At these times greater numbers and varieties of macroinvertebrates were recorded on the riverbed. The very high MCI score of 160 recorded at SH45 in November 1998 (Figure 2) was the result of a community consisting of only one taxon (and just a single individual) which was highly sensitive to pollution. The MCI is not a good indicator of water quality when only a small number of taxa are present and is not typically the index used to assess the impacts of sedimentation in stony streams. However, the MCI has some value in the assessment of recovery of the faunal community with time and has some value in trend evaluation.

A further massive sand drift moved down the river following very heavy February 2004 rainfall and significant flood flows in late February, some three weeks prior to the summer 2004 survey. An additional survey was performed in late winter 2004 to document the continuing effects of sand/sediment drift (see Figures 2 and 3), some 3 months prior to the late spring survey. Further erosion effects occurred in late 2006 delaying the spring 2006 survey and during the latter months of 2007 while significant sand and scoria bed-scouring and sedimentation occurred down the river in mid year and again in spring 2008 delaying the 'spring' survey until early in 2009. No significant headwater erosion events were recorded in 2009-2010, 2010-2011, or 2011-2012 but bed-scouring and sedimentation effects continued to impact through this period. The results of spring (2012) and summer (2012-2013) surveys are presented in Table 126 and Table 127, Appendix I.

3.2.1.1 Mangatete Road site (STY000300)

3.2.1.1.1 Taxa richness and MCI

Thirty-seven surveys have been undertaken in the Stony River at this mid-reach site between October 1995 and March 2012. These results are summarised in Table 6, together with results from the current period, and illustrated in Figure 2.

Table 6 Results from SEM surveys performed in the Stony River at Mangatete Road together with spring 2012 and summer 2013 results

Site code	SEM data (1995 Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
STY000300	37	1-21	10	64-160	113	12	112	10	106

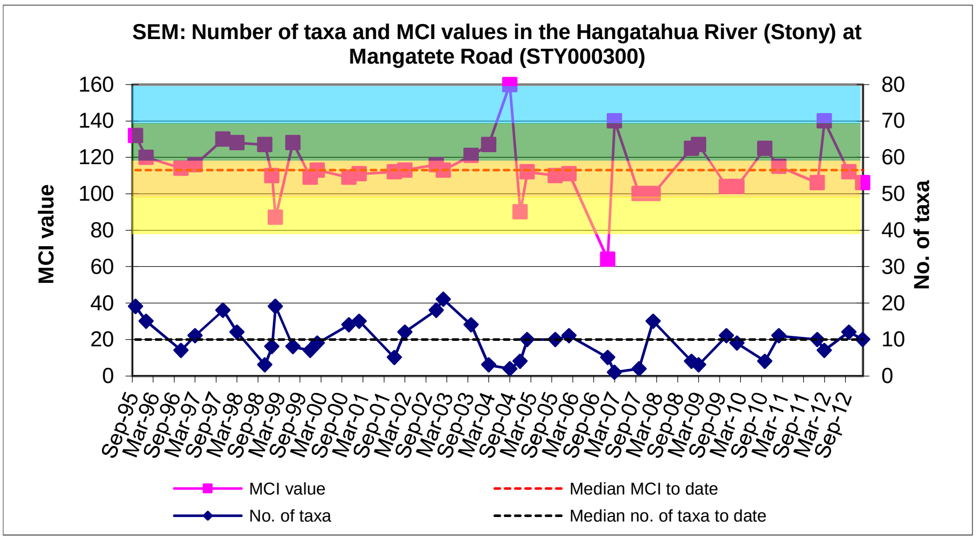


Figure 2 Numbers of taxa and MCI values in the Hangatahua (Stony) River at Mangatete Road

A wide range of richnesses (1 to 21 taxa) has been recorded as a consequence of extensive headwater erosion impacts on the river's communities with a median richness of only 10 taxa, far fewer than might be expected for a ringplain river site at this altitude (160 m asl). In the 2012-2013 period, richness was slightly higher than this median in spring and equal with this median on the summer sampling occasion, indicative of continuing erosion impacts of scouring, finer sediment deposition, and bed movement.

While it is recognised that there may be significant limitations to the appropriateness of the MCI when community compositions are affected by sedimentation and erosion events (e.g. scores show considerable variability when relatively few taxa are present), values at this site have ranged widely between 64 and 160 units with a median MCI value of 113 units. The 2012-2013 scores (113 and 106 units) were relatively similar with the 'spring' score one unit lower, and the summer score seven units lower, than this historical median. Spring and summer scores respectively categorised this site as having 'good' health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' (spring and summer) health for the mid-reaches of a ringplain river at the times of these surveys but the paucity of the communities in terms of richnesses in particular must be taken into account at the site, where headwater erosion effects have been very pronounced. The historical median score (113 units) also placed this site's river health in the 'good' and 'expected' categories.

3.2.1.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 7.

Table 7 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Stony River at Mangatete Road between 1995 and March 2012 [37 surveys] and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	1	3		
EPHEMEROPTERA	*Deleatidium*	8	26	70	A	VA
PLECOPTERA	*Zelandoperla*	8	12	32		A
COLEOPTERA	Elmidae	6	11	30		
TRICHOPTERA	*Aoteapsyche*	4	3	8		
	Costachorema	7	5	14		
	Hydrobiosis	5	1	3		
	Oxyethira	2	1	3		
DIPTERA	*Aphrophila*	5	1	3		
	Eriopterini	5	4	11		
	Maoridiamesa	3	3	8		
	Orthocladiinae	2	7	19		

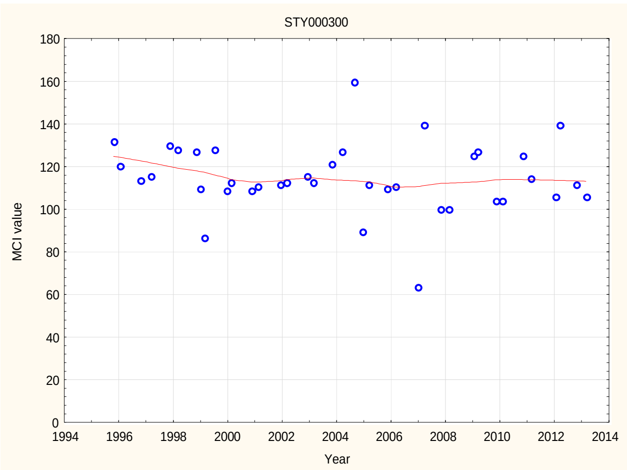
Prior to the current 2012-2013 period, twelve taxa have characterised this site's communities on occasions. These are comprised of two 'highly sensitive', five 'moderately sensitive', and five 'tolerant' taxa. The only predominant taxon has been the 'highly sensitive' taxon [ubiquitous mayfly (*Deleatidium*)]. This taxon and elmid beetles are often present (frequently in large numbers) on unstable shingle-cobble substrates (Death, 2000) and during recovery from erosion/siltation events (Fowles, 1987). Two of these characteristic taxa were dominant in the summer community ('highly sensitive' stonefly, *Zelandoperla* and mayfly, *Deleatidium*) and only one of these taxa was dominant in the summer community; both these results indicative of the significant reduction in diversity of characteristic taxa due to headwater erosion impacts and unstable substrate. The lack of abundances of orthoclad midges on both occasions was coincident with the presence of minimal periphyton mats cover on the cobble-boulder substrate; an indication of limited recovery from scouring/erosion events. The similarity of the characteristic taxa on the two occasions was reflected in the relatively small difference in $SQMCI_s$ scores of 0.8 unit between seasons (Tables 126 and 127), with the higher summer value due to an increased abundance of the mayfly, *Deleatidium.*

3.2.1.1.3 Predicted river 'health'

The Stony River at Mangatete Road is 7.3 km downstream of the National Park boundary at an altitude of 160 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 101 (altitude) and 109 (distance) for this site. The historical site median (113 units) is significantly higher (Stark, 1998) than the altitude prediction and 4 units above the distance predictive value while the spring, 2012 score was slightly higher than the distance and significantly higher than the altitude predictive values, while the summer, 2013 survey score was within 5 units of both predictive values. Of the 39 surveys to date at this site, only 13% of MCI scores have been less than 101 units while 49% have been greater than 109 units.

3.2.1.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Stony River at Mangatete Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes (with the proviso noted earlier for this Stony River site). A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 3.



N = 39
Kendall tau = - 0.159
p level = 0.155 [>FDR, p = 0.239]
N/S at p <0.05

Figure 3 LOWESS trend plot of MCI data at Mangatete Road site

Although a slight decreasing trend in MCI scores is illustrated, this has not been statistically significant. This site has a LOWESS-smoothed range of MCI scores of about 14 units indicative of some significant ecological variability over the period, not surprising given the erosion effect documented earlier and further emphasised by the wide range of individual scores, particularly since 2004. Overall this smoothed trend line shows generic river 'health' (Table 1) deteriorating slightly from 'very good' to 'good' while, in terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain river, 'health' has also deteriorated slightly but from 'better than expected' to 'expected'. However, the majority of the variability has been caused by severe headwater erosion events at varying intervals over the period.

3.2.1.2 SH 45 site (STY000400)

3.2.1.2.1 Taxa richness and MCI

Thirty-five surveys have been undertaken in the Stony River at this lower reach site between October 1995 and March 2012. These results are summarised in Table 6, together with results from the current period, and illustrated in Figure 6.

Table 8 Results from SEM surveys performed in the Stony River at SH 45 together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
STY000400	37	0-18	8	0-160	108	8	105	9	127

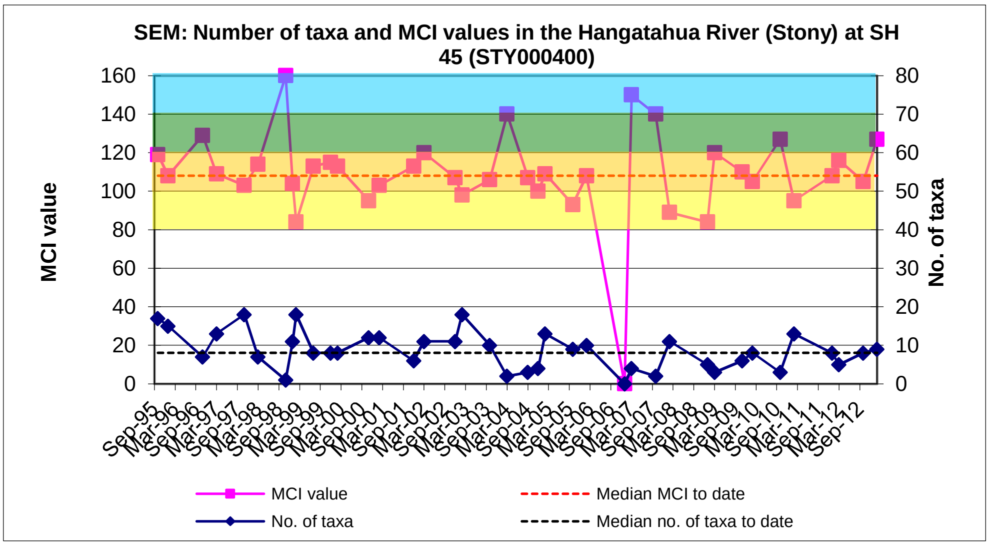


Figure 4 Numbers of taxa and MCI values in the Hangatahua (Stony) River at SH 45

A wide range of richnesses (0 to 18 taxa) has been recorded mainly as a consequence of extensive headwater erosion impacts on the river's communities with a median richness of only 8 taxa, far fewer than would be expected for a ringplain river site at this altitude (70 m asl). In the 2012-2013 period richnesses were equivalent with, or slightly above, this median at the time of the two sampling occasions, indicative of continuing erosion impacts of scouring, finer sediment deposition, and bed movement at this site.

While it is recognised that there may be significant limitations to the appropriateness of the MCI when community compositions are affected by sedimentation and erosion events (e.g. scores show considerable variability when relatively few taxa are present), values at this site have ranged widely between 0 and 160 units with a median MCI value of 108 units. The MCI scores in spring, 2012 (105 units) and summer, 2013 (127 units) were significantly different and in summer, significantly different from the historical median (Figure 4). They categorised this site as having 'good' (spring) and 'very good' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' health in spring, 2012 and 'better than expected' in summer, 2013 for the lower reaches of a ringplain river. The historical median score (108 units) categorised this site as having 'good' generic health, but 'better than expected' predictive health, for a lower river reach.

However, the paucity of numbers and richnesses (in both seasons) should be recognised in this assessment given the historical impacts of headwater erosion effects along the length of the river channel.

3.2.1.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 9.

Table 9 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Stony River at SH 45 between 1995 and March 2012 [37 surveys] and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	1	3		
EPHEMEROPTERA	*Deleatidium*	8	25	68	VA	XA
PLECOPTERA	*Zelandoperla*	8	8	22		A
COLEOPTERA	Elmidae	6	5	14		
TRICHOPTERA	*Aoteapsyche*	4	5	14		A
	Costachorema	7	4	11		
	Hydrobiosis	5	4	11		
	Oxyethira	2	1	3		
DIPTERA	*Aphrophila*	5	1	3		
	Eriopterini	5	1	3		
	Maoridiamesa	3	3	8		
	Orthocladiinae	2	8	22	A	

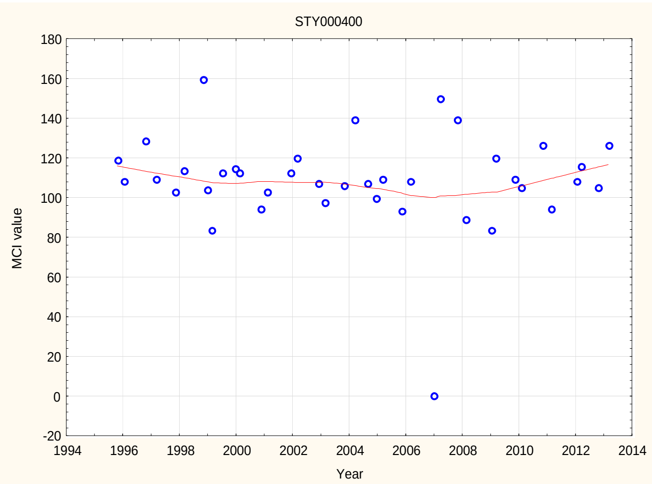
Prior to the current 2012-2013 period, twelve taxa have characterised this site's communities on occasions. These are comprised of two 'highly sensitive', five 'moderately sensitive', and five 'tolerant' taxa. Only one taxon has been predominant; a 'highly sensitive' taxon [the ubiquitous mayfly (mayfly, *Deleatidium*)]. This taxon is often present on unstable shingle-cobble substrates (Death, 2000) and during recovery from erosion/siltation events (Fowles, 1987). Only two of the characteristic taxa were dominant in the spring community [mayfly, (*Deleatidium*) and orthoclad midges] and the mayfly taxon (*Deleatidium*) was also (extremely) abundant in the summer community along with one other 'highly sensitive' taxon [stonefly (*Zelandoperla*)] and one 'tolerant' taxon [free-living caddisfly (*Aoteapsyche*)]. Both these results were indicative of a paucity of characteristic taxa due to preceding headwater erosion impacts and/or substrate instability. An abundance of midges was recorded in spring despite the presence of only thin periphyton mats on the cobble-boulder substrate. The greater abundance of 'highly sensitive' taxa was reflected in the higher summer $SQMCI_s$ score (increase of 1.0 unit).

3.2.1.2.3 Predicted river 'health'

The Stony River at SH 45 is 12.5 km downstream of the National Park boundary at an altitude of 70 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 92 (altitude) and 103 (distance) for this site. The historical site median (108 units) is significantly higher (Stark, 1998) than the altitude prediction and 5 units above the distance predictive value while the spring, 2012 and summer 2013 surveys' scores were both significantly higher than the altitude predictive value and also higher than the distance predictive value. Of the 39 surveys to date at this site, only 8% of MCI scores have been less than 92 units while 72% have been greater than 103 units.

3.2.1.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Stony River at SH 45. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes (with the proviso noted earlier for this Stony River site). A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 5.



N = 39
Kendall tau = - 0.057
p level = 0.608 [>FDR, p = 0.702]
N/S at $p < 0.05$

Figure 5 LOWESS trend plot of MCI data at SH 45 site

An overall slightly decreasing trend in MCI scores has not been statistically significant. The site has a LOWESS-smoothed MCI range of about 17 units indicative of some significant ecological variability over the period for the same reasons as those responsible for variability at the upstream site (Mangatete Rd). Overall, smoothed scores have shown generic river 'health' (Table 1) within the 'good' grade while, in terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain river, 'health' has deteriorated slightly from 'better than expected' to 'expected' before a more recent gradual improvement. This was a similar trend to that found at the upstream mid-reach (Mangatete Road) site. Greater variability in scores has been apparent since 2004 with the majority of the variability in MCI scores associated with headwater erosion events.

3.2.1.3 Discussion

Due to the major influence of historic and relatively frequent headwater erosion events, scouring, and instability of the river bed; seasonal and spatial differences in macroinvertebrate communities in the Stony River often have not been as pronounced as elsewhere in ringplain streams. Although seasonal MCI values at each site showed variable differences between spring and in summer, with a 22 unit increase in scores at the downstream site under summer conditions, the paucity of the communities at both sites should be noted on both seasonal occasions.

MCI scores typically (slightly) decreased in a downstream direction in spring over a distance of 5.2 km, equating to a rate of decline of 1.3 units/km which was relatively similar to the predicted rate (1.15 units/km) over the equivalent length of a National Park-sourced river (Stark and Fowles, 2009). A marked atypical increase of 21 units in summer was influenced by the paucity of the communities as a result of prior headwater erosion events and/or substrate instability.

3.2.2 Timaru Stream

In the 2008-2009 period severe headwater erosion events had impacted upon the macroinvertebrate communities of the upper reaches of this stream in particular (TRC, 2009). The results found in the 2012-2013 surveys are presented in Table 128 and 129, Appendix I.

3.2.2.1 Carrington Road site (TMR000150)

3.2.2.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this upper reach site in the Timaru Stream inside the National Park boundary at Carrington Road between October 1995 and March 2012. These results are summarised in Table 10, together with the results from the current period, and illustrated in Figure 6.

Table 10 Results of previous surveys performed in the Timaru Stream at Carrington Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
TMR000150	34	8-32	25	119-144	136	27	132	28	138

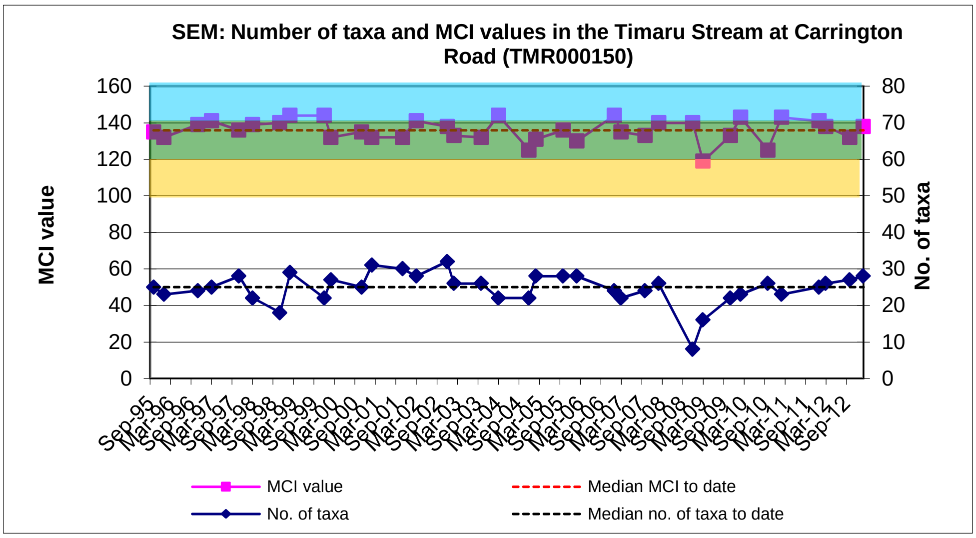


Figure 6 Numbers of taxa and MCI values in the Timaru Stream at Carrington Road

A wide range of richnesses (8 to 32 taxa) has been found; wider than might be expected, due to headwater erosion effects over the 2008-2009 period which markedly reduced richness, with a median richness of 25 taxa (slightly below that representative of typical richnesses in ringplain streams and rivers near the National Park boundary). During the 2012-2013 period, spring (27 taxa) and summer (28 taxa) richnesses were slightly above this median richness and indicative of continuing recovery from earlier headwater erosion events.

MCI values have had a slightly wider range (25 units) at this site than typical of a site near the National Park boundary, due in part to an atypically low value after the 2008-2009 headwater erosion period. The median value (136 units) has been typical of upper reach sites elsewhere on the ringplain however. The spring, 2012 score (132 units) was slightly lower than this median, while the summer, 2013 score (138 units) was typical for such a site, and close to the historical median. These scores categorised this site as having 'very good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' health for the upper reaches of a ringplain stream on both of these occasions. The historical median score (136 units) placed this site in the 'very good' category for the generic, and 'expected' category for the predictive methods of assessment.

3.2.2.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 11.

Table 11 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Timaru Stream at Carrington Road between 1995 and March 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
EPHEMEROPTERA	*Austroclima*	7	3	9		
	Coloburiscus	7	25	74	A	A
	Deleatidium	8	34	100	XA	VA
	Nesameletus	9	30	88	A	A
PLECOPTERA	*Acroperla*	5	4	12		
	Stenoperla	10	2	6		
	Zelandobius	5	25	74	VA	A
	Zelandoperla	8	22	65	A	A
COLEOPTERA	Elmidae	6	13	38		A
MEGALOPTERA	*Archichauliodes*	7	1	3		
TRICHOPTERA	*Costachorema*	7	2	6		
	Hydrobiosis	5	1	3		
	Hydrobiosella	9	3	9		
	Orthopsyche	9	2	6		
	Beraeoptera	8	2	6		A
	Helicopsyche	10	4	12		
DIPTERA	*Aphrophila*	5	10	29		A
	Maoridiamesa	3	2	6		
	Orthocladiinae	2	19	56		A

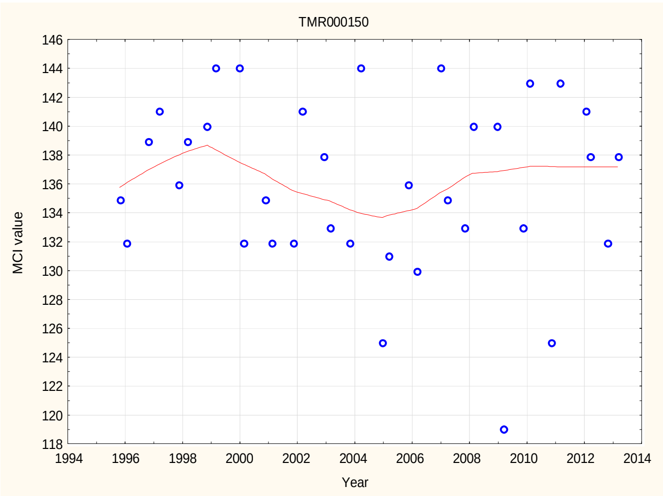
Prior to the current 2012-2013 period, 19 taxa had characterised the community at this site on occasions. These have comprised eight 'highly sensitive', nine 'moderately sensitive', and two 'tolerant' taxa i.e. a majority of 'sensitive' taxa as would be expected near the National Park boundary of a ringplain stream. Predominant taxa have included three 'highly sensitive' taxa [mayflies ((*Deleatidium* on every sampling occasion) and *Nesameletus*) and stonefly (*Zelandoperla*)]; two 'moderately sensitive' taxa [mayfly (*Coloburiscus*) and stonefly (*Zelandobius*)], and one 'tolerant' taxon (orthoclad midges). Five of these taxa (all 'sensitive' taxa) were dominant in the spring, 2012 community. All of these taxa were again dominant in the summer, 2013 community together with one of the 'highly sensitive' taxa [cased caddisfly (*Beraeoptera*)], one of the 'moderately sensitive' taxa (elmid beetles), and one of the 'tolerant' taxa [net-building caddisfly (*Aoteapsyche)*]. Some decreases in abundances within certain 'sensitive' characteristic taxa in summer community composition were reflected in the small drop in seasonal $SQMCI_s$ values of 0.5 unit (Tables 128 and 129).

3.2.2.1.3 Predicted stream 'health'

The Timaru Stream at Carrington Road is within the National Park boundary at an altitude of 420 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 127 (altitude) and 132 (distance) for this site. The historical site median (136 units) is 9 units higher than the altitude prediction and 4 units higher than the distance predictive value. The spring 2012 score (132 units) was equal with and 5 units higher than these predictive values and the summer score (138 units) was significantly higher (Stark, 1998) than the predictive altitude value. Of the 36 surveys to date at this site, only 8% of MCI scores have been less than 127 units while 69% have been greater than 132 units.

3.2.2.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Timaru Stream at Carrington Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 7.



N = 36
Kendall tau = -0.056
p level = 0.632 [>FDR, p = 0.709]
N/S at p <0.05

Figure 7 LOWESS trend plot of MCI data at the Carrington Road site

MCI scores have trended very slightly downwards in general, with a more recent improvement, but the trend has not been statistically significant over the period. The LOWESS-smoothed MCI scores have ranged over 6 units which have not been ecologically significant. Smoothed scores have been indicative of 'very good' generic stream health (Table 1) throughout the period and, in terms of predictive relationships (Table 2) for a site in the upper reaches of a ringplain stream at the boundary of the National Park, stream health has remained within the 'expected' category throughout the eighteen year period.

3.2.2.2 SH45 site (TMR000375)

3.2.2.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Timaru Stream at this lower, mid-reach site at SH45 between October 1995 and March 2012. These results are summarised in Table 12, together with the results from the current period, and illustrated in Figure 8.

Table 12 Results of previous surveys performed in the Timaru Stream at SH45, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
TMR000375	34	13-34	26	89-120	102	28	114	27	103

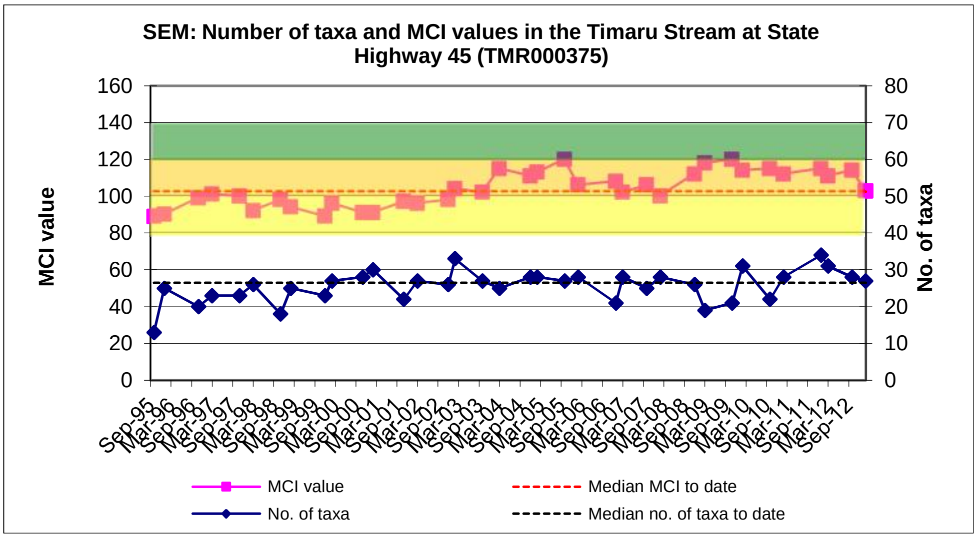


Figure 8 Numbers of taxa and MCI values in the Timaru Stream at State Highway 45

A wide range of richnesses (13 to 34 taxa) has been found; wider than might be expected, with a median richness of 26 taxa (more representative of typical richnesses in the mid reaches of ringplain streams and rivers). During the 2012-2013 period spring (28 taxa) and summer (27 taxa) richnesses were very similar and only slightly higher than the median taxa number in both spring and in summer, when substrate periphyton cover (mats and filamentous algae) was patchy in spring and more widespread in summer.

MCI values have had a slightly wider range (31 units) at this site than typical of sites in the mid reaches of ringplain streams. The median value (102 units) has been relatively typical of mid reach sites elsewhere on the ringplain however, but the spring 2012 (114 units) score was well above those typical for such a site and significantly (Stark, 1998) higher than the historical median by 12 units in spring but insignificantly higher (by 1 unit) in summer. These scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' health in spring and 'expected' health in summer for the lower mid reaches of a ringplain stream. The historical median score (102 units) placed this site in the 'good' category for the generic and 'expected' category for the predictive methods of assessment.

3.2.2.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 13.

Table 13 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Timaru Stream at SH45 between 1995 and March 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	3		
ANNELIDA	Oligochaeta	1	13	38		
MOLLUSCA	*Potamopyrgus*	4	4	12		
EPHEMEROPTERA	*Austroclima*	7	11	32		A
	Coloburiscus	7	21	62	VA	A
	Deleatidium	8	14	41	XA	A
	Rallidens	9	2	6		
PLECOPTERA	*Acroperla*	5	5	15		
	Zelandobius	5	2	6	VA	
	Zelandoperla	8	15	44	A	A
COLEOPTERA	Elmidae	6	17	50	A	VA
MEGALOPTERA	*Archichauliodes*	7	14	41	A	A
TRICHOPTERA	*Aoteapsyche*	4	31	91	A	VA
	Costachorema	7	11	32		
	Hydrobiosis	5	7	21		
	Neurochorema	6	6	18		A
	Beraeoptera	8	5	15	VA	
	Confluens	5	1	3		
	Oxyethira	2	7	21		
	Pycnocentrodes	5	15	44	VA	
DIPTERA	*Aphrophila*	5	32	94	A	VA
	Maoridiamesa	3	26	76		A
	Orthocladiinae	2	33	97	A	A
	Tanytarsini	3	6	18		VA
	Empididae	3	5	15		
	Muscidae	3	4	12		
	Austrosimulium	3	13	38		

Prior to the current 2012-2013 period, 27 taxa had characterised the community at this site on occasions. These have comprised four 'highly sensitive', twelve 'moderately sensitive', and eleven 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and an increase in the proportion of 'tolerant' taxa as would be expected in the mid reaches compared with the upper reaches of a ringplain stream. Predominant taxa have included no 'highly sensitive' taxa, three 'moderately sensitive' taxa [mayfly (*Coloburiscus*), elmid beetles, and cranefly (*Aphrophila*)], and three 'tolerant' taxa [net-building caddisfly (*Aoteapsyche*)] and midges (*Maoridiamesa* and orthoclads)]. Eleven of the historically characteristic taxa were dominant in the spring 2012 community. These comprised three 'highly sensitive', six 'moderately sensitive', and two 'tolerant' taxa, whereas two 'highly sensitive', six 'moderately sensitive', and four 'tolerant' taxa comprised the dominant taxa of the summer community when 'tolerant' taxa proportionally were more dominant. Eight of these 15 taxa were dominant in both spring and summer communities (Table 13) but a reduction in numerical dominance of two 'highly sensitive' taxa and increased numerical dominance in two 'tolerant' taxa in particular were reflected in the lower summer seasonal $SQMCI_s$ score (Table 128 and 129) which decreased by 2.0 units.

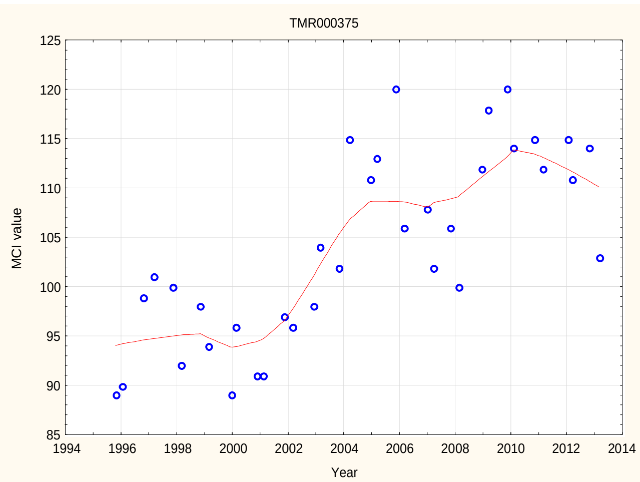
Of note, the 'highly sensitive' flare-cased caddisfly (*Beraeoptera*) and 'moderately sensitive' stonefly (*Zelandobius*), which have seldom characterised this site's communities on past survey occasions, were very dominant in the spring survey but not among the dominant taxa at the time of the summer survey.

3.2.2.2.3 Predicted stream 'health'

The Timaru Stream at SH45 is 10.9 km downstream of the National Park boundary at an altitude of 100 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 95 (altitude) and 105 (distance) for this site. The historical site median (102) is 7 units higher than the altitude prediction and 5 units lower than the distance predictive value. The spring survey score (114 units) was higher than both predictive values while the summer score (103 units) was higher than the predictive altitude value but lower than predictive distance value. Of the 36 surveys to date at this site, 20% of MCI scores have been less than 95 units while 44% have been greater than 105 units.

3.2.2.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Timaru Stream at SH45. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 9.



N = 36
Kendall tau = +0.552
p level < 0.0001 [>FDR, p <0.001]
Significant at p <0.05 and p <0.01 after FDR

Figure 9 LOWESS trend plot at the SH45 site

MCI scores have shown a strong temporal trend of improvement (statistically significant), particularly since 2001, with most of the more recent scores (since 2004) well above scores recorded toward the start of the monitoring period. The LOWESS-smoothed scores have ranged over 19 units, an ecologically significant range. No obvious explanations have been apparent for the trend but a possible reason may be related to improved management of dairy shed wastes disposal in the catchment above this SH45 site. Smoothed MCI scores have indicated an improvement in generic stream 'health' (Table 1) from 'fair' to 'good', where it currently remains. In terms of predictive relationships (Table 2) for a site toward the lower end of the mid reaches of a ringplain stream, stream health has improved from 'expected' to 'better than expected' where it has remained since 2003.

3.2.2.3 Discussion

Seasonal MCI values typically remained relatively similar between spring and summer at the National Park boundary site where historical median scores have been within two units (Appendix II), whereas a significant summer decrease of 11 units was found at the lower mid reach site where a difference of one unit in seasonal historical median scores has been found (Appendix II). The percentage composition of 'tolerant' taxa increased (by 12%) in the summer community. Seasonal communities at the upper site shared 20 common taxa (57% of the 35 taxa found at this site in 2012-2013), a lower percentage than normally the case at an upper reach site. This compares with 21 shared common taxa (62% of the 34 taxa found in 2012-2013) at the lower mid reaches site (SH45), an atypically less pronounced seasonal change in community structure at the further downstream site. The two sites shared 20 common taxa (57% of the 35 taxa at upper and mid reach sites) in spring and 15 common taxa (38% of 40 taxa) in summer, indicative of the dissimilarity in spatial community structures in both spring and typically more so in summer.

MCI score typically fell in a downstream direction in both spring (by 18 units) and in summer (by 35 units), over a stream distance of 10.9 km downstream from the National Park boundary. These equated to rates of decline of 1.7 units/km in spring increasing to 3.2 units/km in summer, compared with a predicted rate of 2.4 units/km over the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009).

Using the long-term median SEM MCI scores for both sites (Appendix II), the rate of decline between the upper and lower reach sites has been 3.0 units/km over the surveyed length. Therefore rates of decline were much lower in both spring and similar in summer of the 2012–2013 period to the average rate for the period to date.

3.2.3 Mangaoraka Stream

The results found by the 2012-2013 surveys are presented in Tables 130 and 131, Appendix I.

3.2.3.1 Corbett Road site (MRK000420)

3.2.3.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this lower reach site in the Mangaoraka Stream between October 1995 and March 2012. These results are summarised in Table 14, together with the results from the current period, and illustrated in Figure 10.

Table 14 Results of previous surveys performed in Mangaoraka Stream at Corbett Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MRK000420	34	11-30	25	75-105	90	24	101	24	92

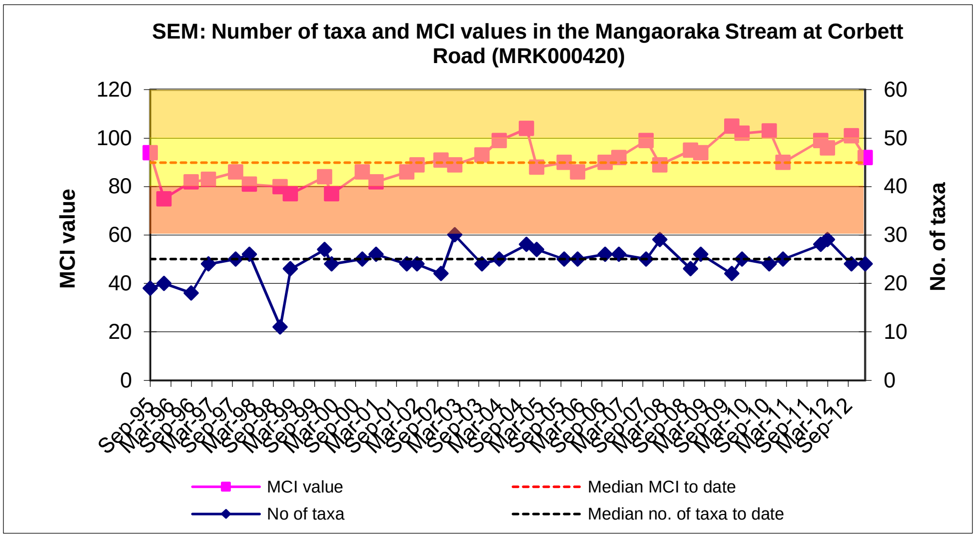


Figure 10 Numbers of taxa and MCI values in the Mangaoraka Stream at Corbett Road

A wide range of richnesses (11 to 30 taxa) has been found, with a median richness of 25 taxa (more representative of typical richnesses in the lower reaches of ringplain streams rising outside the National Park boundary). During the 2012-2013 period spring (24 taxa) and summer (24 taxa) richnesses were identical and slightly lower than this median richness.

MCI values have also had a relatively wide range (30 units) at this site. The median value (90 units) has been typical of lower reach sites elsewhere on the ringplain however, but the spring, 2012 (101 units) score was higher than typical for such a site and a significant (Stark, 1998) 11 units above the historical median, while the summer value (92 units) was 2 units higher than the historical median. These scores categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' (spring and summer) health for the lower reaches of a ringplain stream on these occasions. The historical median score (90 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.3.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 15.

Table 15 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Mangaoraka Stream at Corbett Road, between 1995 and March 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
PLATYHELMINTHES	*Cura*	3	1	3		
NEMERTEA	Nemertea	3	4	12		A
ANNELIDA	Oligochaeta	1	26	76	A	
MOLLUSCA	*Latia*	5	2	6		
	Physa	3	1	3		
	Potamopyrgus	4	30	88		A
CRUSTACEA	Ostracoda	1	1	3		
EPHEMEROPTERA	*Austroclima*	7	19	56	A	VA
	Coloburiscus	7	4	12		
	Deleatidium	8	5	15	VA	
	Zephlebia group	7	3	9		
PLECOPTERA	*Zelandobius*	5	12	35	A	
COLEOPTERA	Elmidae	6	22	65	VA	XA
MEGALOPTERA	*Archichauliodes*	7	14	41	A	A
TRICHOPTERA	*Aoteapsyche*	4	29	85	A	XA
	Costachorema	7	2	6	A	
	Hydrobiosis	5	25	74		A
	Neurochorema	6	2	6		
	Oxyethira	2	6	18		
	Pycnocentria	7	2	6		
	Pycnocentrodes	5	25	74	A	
DIPTERA	*Aphrophila*	5	19	56	VA	
	Maoridiamesa	3	9	26		
	Orthocladiinae	2	28	82	A	
	Tanytarsini	3	8	24		A
	Empididae	3	5	15		
	Muscidae	3	2	6		
	Austrosimulium	3	11	32		

Prior to the current 2012-2013 period, 28 taxa had characterised the community at this site on occasions. These have comprised only one 'highly sensitive', 13 'moderately sensitive', and 14 'tolerant' taxa i.e. a high proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain stream.

Predominant taxa have included five 'moderately sensitive' taxa [mayfly (*Austroclima*), elmid beetles, free-living caddisfly (*Hydrobiosis*), stony-cased caddisfly (*Pycnocentrodes*), and cranefly (*Aphrophila*)], and four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and orthoclad midges].

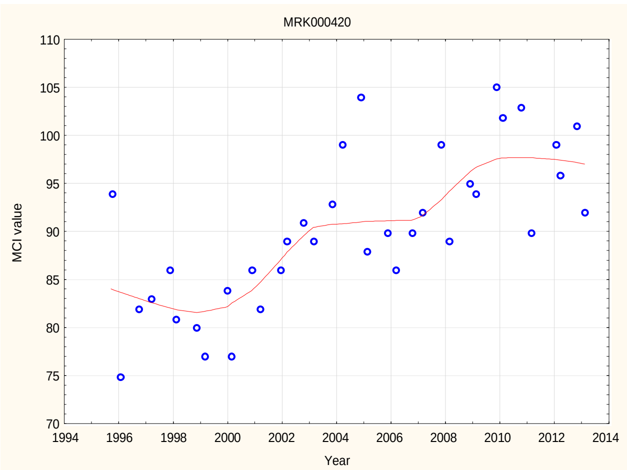
Eleven of the historically characteristic taxa were dominant in the spring, 2012 community and were comprised of seven of the predominant taxa (above) together with another one 'highly sensitive' and three 'moderately sensitive' taxa. The summer, 2013 community was characterised by only four of the taxa dominant in spring, together with an additional two 'moderately sensitive' and one 'tolerant' (sandfly) taxa, all of which previously had been characteristic of this site's communities (Table 15). The small decrease in 'sensitive' summer dominant taxa and absence of the dominant 'highly sensitive' taxon were reflected in the increase in $SQMCI_s$ scores of 0.6 unit (Tables 130 and 131). With one exception [mayfly (*Deleatidium*)], the taxa which were recorded as very or extremely abundant during spring and/ or summer had characterised this site's communities on 56% to 85% of past surveys.

3.2.3.1.3 Predicted stream 'health'

The Mangaoraka Stream rises below the National Park boundary and the site at Corbett Road is in the lower reaches at an altitude of 60 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 91 units for this site. The historical site median (90 units) is only one unit below this altitude prediction while the spring survey score (101 units) and the summer score (92 units) were both above this predictive value. Of the 36 surveys to date at this site, 67% of MCI scores have been less than 91 units, indicating that the current spring and summer MCI scores were better than typical historical conditions.

3.2.3.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Mangaoraka Stream at Corbett Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 11.



N = 36
Kendall tau = + 0.563
p level < 0.0001 [>FDR, p <0.001]
Significant at p <0.05 and p <0.01 after FDR application

Figure 11 LOWESS trend plot at the Corbett Road site

This site's MCI scores have shown a strong, statistically significant improvement ($p < 0.01$ after FDR), particularly since 1999 with the trend tending to plateau between 2003 and 2007 before improving strongly more recently. These latest scores remain above scores recorded prior to 2000. The trend was statistically significant after FDR application. The LOWESS-smoothed scores have varied over an ecologically significant range of 16 units during the period. SEM physicochemical monitoring at this site had illustrated significant improvements in aspects of organic loadings at this site in the lower reaches of the stream prior to mid 2008. This was coincident with more rigorous surveillance monitoring of nearby quarrying and waste disposal activities and good dairy shed wastewater disposal compliance performance during that period although more recently, aspects of poorer overall water quality (i.e. increased bacteriological numbers and increasing trends in certain nutrient species) have been recorded despite the apparent improvement in biological communities.

LOWESS-smoothed MCI scores have been indicative of 'fair' generic stream health (Table 1) throughout the period improving toward 'good' very recently and, in terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, health has remained in the 'expected' category (which has been maintained since 2003) over the eighteen year period.

3.2.3.2 Discussion

Seasonal MCI values typically decreased between spring and summer (by 9 units) at this lower reach site, a larger decrease than the historical median summer decrease (4 units) in scores (Appendix II). The percentage composition of 'tolerant' taxa increased by 8% in the summer community although periphyton mats and filamentous algal substrate cover were very similar. Seasonal communities at this site shared a relatively high number of common taxa (20 taxa; 71% of the 28 taxa found at this site in 2012-2013), and the small increase in the proportion of 'tolerant' taxa in summer resulted in a decrease of 9 units in MCI values between seasons.

3.2.4 Waiongana Stream

The results found by the 2012-2013 surveys are presented in Table 132 and Table 133, Appendix I.

3.2.4.1 State Highway 3a site (WGA000260)

3.2.4.1.1 Taxa richness and MCI

Thirty-five surveys have been undertaken at this mid reach site in the Waiongana Stream between October 1995 and April 2012. These results are summarised in Table 16, together with the results from the current period, and illustrated in Figure 12.

Table 16 Results of previous surveys performed in the Waiongana Stream at SH3a together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to April 2012))					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGA000260	35	12-30	24	82-112	96	24	110	26	98

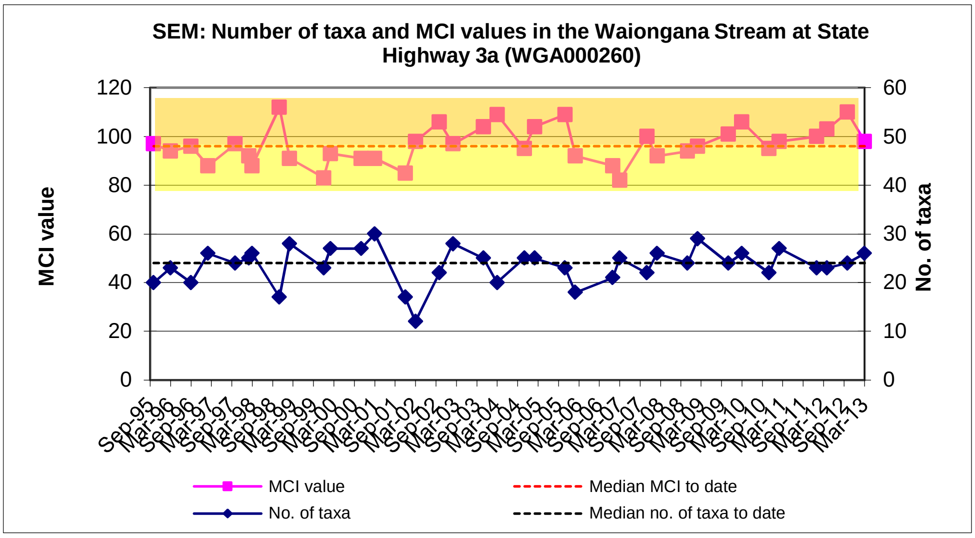


Figure 12 Numbers of taxa and MCI values in the Waiongana Stream at State Highway 3A

A wide range of richnesses (12 to 30 taxa) has been found; with a median richness of 24 taxa (more representative of typical richnesses in the mid-reaches of ringplain streams and rivers. During the 2012-2013 period, spring (24 taxa) and summer (26 taxa) richnesses were similar and very close to this median richness.

MCI values have also had a relatively wide range (30 units) at this site, relatively typical of a site in the mid reaches of a ringplain stream. The median value (96 units) also has been typical of mid-reach sites elsewhere on the ringplain. The spring, 2012 (110 units) and summer, 2013 (98 units) scores were significantly 12 units different and 14 and 2 units above the historical median respectively. These scores categorised this site as having 'fair' (summer) and 'good' (spring) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' health for the mid reaches of a ringplain stream in spring and in summer. The historical median score (96 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.4.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 17.

Table 17 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiongana Stream at SH3a between 1995 and April 2012 [35 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	2	6		
ANNELIDA	Oligochaeta	1	18	51		A
MOLLUSCA	*Potamopyrgus*	4	14	40		
CRUSTACEA	*Paracalliope*	5	1	3		
EPHEMEROPTERA	*Austroclima*	7	3	9		A
	Coloburiscus	7	3	9	A	
	Deleatidium	8	16	46	XA	XA
COLEOPTERA	Elmidae	6	26	74	VA	XA
MEGALOPTERA	*Archichauliodes*	7	9	26		A
TRICHOPTERA	*Aoteapsyche*	4	22	63		VA
	Costachorema	7	9	26	A	A
	Hydrobiosis	5	16	46		A
	Neurochorema	6	2	6		
	Oxyethira	2	9	26		
	Pycnocentrodes	5	9	26	A	A
DIPTERA	*Aphrophila*	5	29	83	VA	VA
	Maoridiamesa	3	22	63	A	VA
	Orthocladiinae	2	31	89	VA	A
	Tanytarsini	3	12	34		A
	Empididae	3	6	17		
	Muscidae	3	6	17		
	Austrosimulium	3	3	9		

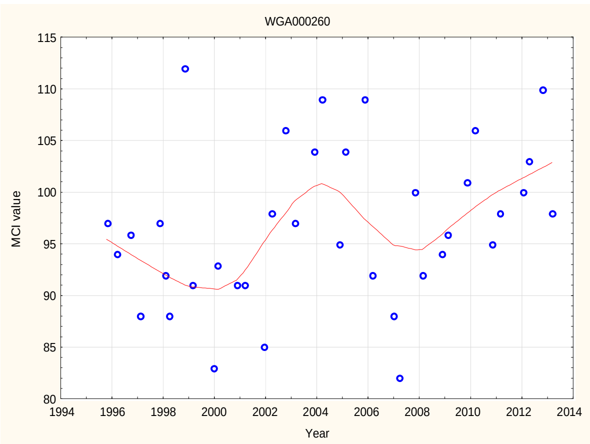
Prior to the current 2012-2013 period, 22 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', ten 'moderately sensitive', and eleven 'tolerant' taxa i.e. a relatively even balance of 'sensitive' and 'tolerant' taxa as would be expected in the mid-reaches of a ringplain stream. Predominant taxa have included no 'highly sensitive' taxa, two 'moderately sensitive' taxa [elmid beetles and cranefly (*Aphrophila*)], and four 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]. Four of these predominant taxa were dominant in the spring, 2012 community together with four of the other historically characteristic taxa. The summer, 2013 community was characterised by seven of the taxa dominant in spring, together with an additional three 'moderately sensitive' and three 'tolerant' taxa, all of which previously had been characteristic of this site's communities on occasions (Table 17). A marked increase in the numerical abundance of the 'tolerant' net-building midge (*Aoteapsyche*) in particular was reflected in the small summer decrease of 0.4 unit in $SQMCI_s$ scores (Tables 132 and 133). Despite the 'highly sensitive' taxa not being amongst the historically predominant taxa, it was extremely abundant on both occasions. The six taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 46% to 89% of past survey occasions.

3.2.4.1.3 Predicted stream 'health'

The Waiongana Stream site at SH3a is 16.1 km downstream of the National Park boundary at an altitude of 140 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 99 (altitude) and 100 (distance) for this site. The historical site median (96 units) is 3 units lower than the altitude prediction and 4 units below the distance predictive value, while the spring, 2012 survey score (110 units) was 10 to 11 units above both predictive values while the summer, 2013 score (98 units) was slightly lower than both predictive values. Of the 37 surveys to date at this site, 68% of MCI scores have been less than 99 units while only 24% have been greater than 100 units.

3.2.4.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waiongana Stream at SH3a. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 13.



N = 37
Kendall tau = + 0.238
p level = 0.037 [>FDR, p = 0.071]
Significant at $p < 0.05$
N/S after FDR

Figure 13 LOWESS trend plot of MCI data at the SH3a site

There was a positive overall trend in the MCI scores identified, which had no statistical significance ($p > 0.05$ after FDR). There has been a steady improvement in scores between 2001 and 2004 followed by a decline in scores until 2008, and a steady increase to date. This site's scores have had a LOWESS-smoothed range of 12 units indicative of marginal ecologically significant variability over the period.

Overall, smoothed scores remained indicative of 'fair' generic stream health (Table 1) for the majority of the period, improving to 'good' 'health' briefly over 2003 to 2005 and again since 2011. In terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain stream, stream health has been in the 'expected' category almost throughout the eighteen year period, bordering on 'worse than expected' for a short period in 1998-2001.

3.2.4.2 Devon Road site (WGA000450)

3.2.4.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this lower reach site at SH45 in the Waiongana Stream between October 1995 and April 2012. These results are summarised in Table 18, together with the results from the current period, and illustrated in Figure 14.

Table 18 Results of previous surveys performed in the Waiongana Stream at Devon Road together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to April 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGA000450	34	12-29	22	72-102	88	19	98	22	87

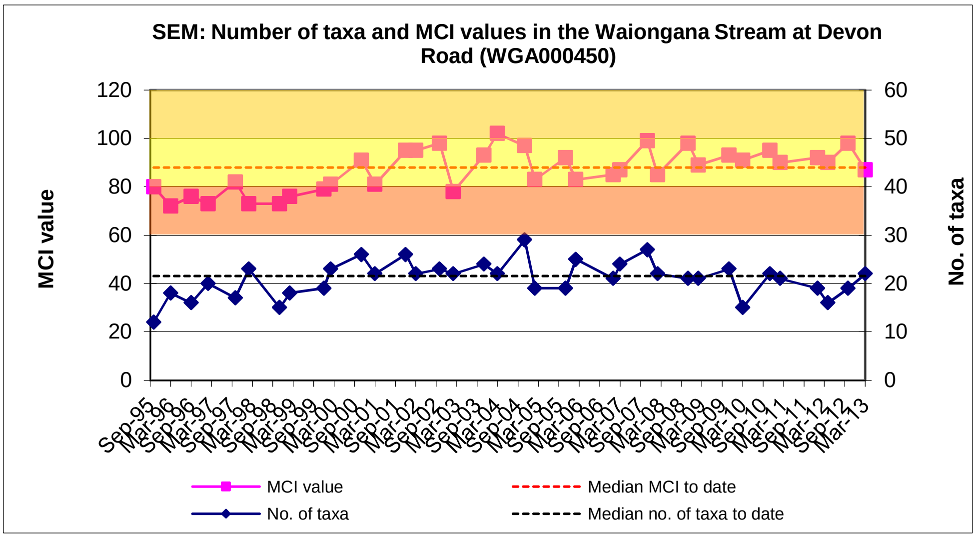


Figure 14 Numbers of taxa and MCI values in the Waiongana Stream at Devon Road

A wide range of richness (12 to 29 taxa) has been found; wider than might be expected with a median richness of 22 taxa, more representative of typical richnesses in ringplain streams and rivers in the lower reaches. During the 2012-2013 period, spring (19 taxa) and summer (22 taxa) richnesses were relatively similar and within 3 taxa of the median taxa number.

MCI scores have had a relatively wide range (30 units) at this site more typical of sites in the lower reaches of ringplain streams. The median value (88 units) has been relatively typical of lower reach sites elsewhere on the ringplain however, with the spring, 2012 (98 units) and summer, 2013 (87 units) scores within the range typical for such a site but higher than the historical median by 10 units in spring and within one unit in summer. These scores categorized this site as having 'fair' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' health on both occasions for the lower reaches of a ringplain stream. The historical median score (88 units) placed this site in the 'fair' category for generic and 'expected' category for predictive methods of assessment.

3.2.4.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 19.

Table 19 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiongana Stream at Devon Road between 1995 and April 2012 [34 surveys], by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	9		
ANNELIDA	Oligochaeta	1	27	79		
MOLLUSCA	*Ferrissia*	3	1	3		
	Latia	5	2	6		
	Potamopyrgus	4	22	65		XA
CRUSTACEA	*Paracalliope*	5	2	6		
	Paratya	3	1	3		
EPHEMEROPTERA	*Austroclima*	7	4	12		
	Deleatidium	8	6	18	VA	
PLECOPTERA	*Zelandobius*	5	1	3		
COLEOPTERA	Elmidae	6	15	44	A	VA
MEGALOPTERA	*Archichauliodes*	7	6	18		A
TRICHOPTERA	*Aoteapsyche*	4	22	65		XA
	Costachorema	7	2	6		
	Hydrobiosis	5	12	35		A
	Oxyethira	2	8	24		
	Pycnocentrodes	5	14	41	A	
DIPTERA	*Aphrophila*	5	14	41	A	
	Maoridiamesa	3	14	41	A	
	Orthocladiinae	2	29	85	A	
	Tanytarsini	3	11	32		A
	Empididae	3	1	3		
	Muscidae	3	3	9		
	Austrosimulium	3	6	18		

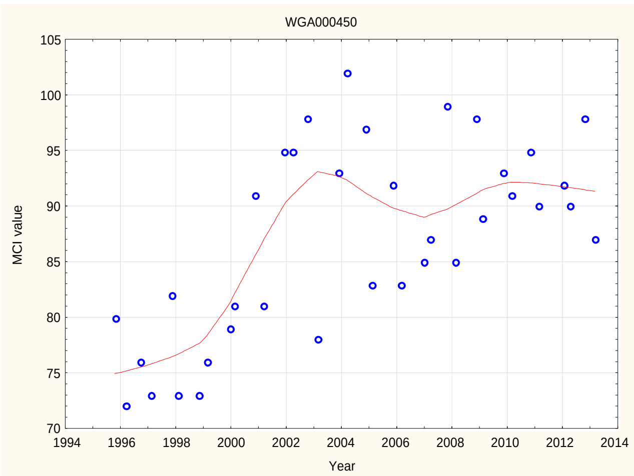
Prior to the current 2012-2013 period, 24 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', ten 'moderately sensitive', and 13 'tolerant' taxa i.e. a majority of 'tolerant' taxa as would be expected in the lower reaches of a ringplain stream. Predominant taxa have included no 'highly sensitive' taxa; no 'moderately sensitive' taxa; but four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and orthoclad midges]. Only six of the historically characteristic taxa were dominant in the spring 2012 community. These comprised one 'highly sensitive', three 'moderately sensitive', and two 'tolerant' taxa, whereas no 'highly sensitive', three 'moderately sensitive', and three 'tolerant' taxa comprised the dominant taxa in the summer community. Only one of these eleven taxa was dominant in both spring and summer communities (Table 19). The decrease in abundance of the one 'highly sensitive' taxon and the marked increase in numerical abundance within two 'tolerant' taxa in particular in summer were reflected in the lower (by 1.7 units) $SQMCI_s$ score at that time (Tables 132 and 133). All taxa recorded as very abundant during spring and /or summer had characterised this site's communities on 18% to 65% of past surveys.

3.2.4.2.3 Predicted stream 'health'

The Waiongana Stream at Devon Road is 31.2 km downstream of the National Park boundary at an altitude of 20 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 86 (altitude) and 93 (distance) for this site. The historical site median (88 units) is two units above the altitude prediction and 5 units lower than the predictive distance value, while the spring, 2012 survey score (98 units) was a significant (Stark, 1998) 12 units higher than the altitude predictive value. The summer, 2013 score (87 units) was very similar to the predictive altitude value and 6 units less than the predictive distance value. Of the 36 surveys to date at this site, 44% of MCI scores have been less than 86 units while only 28% have been greater than 93 units.

3.2.4.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site, in the Waiongana Stream at Devon Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 15.



N = 36
Kendall tau = + 0.438
p level = 0.0002 [>FDR, p = 0.001]
Significant at $p < 0.05$ and $p < 0.01$
after FDR application

Figure 15 LOWESS trend plot at the Devon Road site

MCI scores at this site have shown a statistically significant, strong temporal improvement over the period, despite some relatively low scores between 2003 and 2008. However, the more recent scores remain well above those recorded over the first five years of the period. The LOWESS-smoothed scores have varied over an ecologically significant range of 18 units. Improvement has been coincident with a reduction in consented NPDC water abstraction and more rigorous control of an upstream large piggery's wastes disposal loadings to the stream. This trend of improvement in stream 'health' at this site is much more pronounced than the trend at the site some 15 km upstream, particularly since 1999, indicating that activities in the catchment between these two sites have had a significant influence.

Overall smoothed MCI scores have indicated an improvement in generic stream 'health' (Table 1) from consistently 'poor' prior to 2000 to 'fair' where it has remained over the last twelve years. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, the stream health has improved from 'worse than expected' to 'expected' where it has remained since 2000.

3.2.4.3 Discussion

Seasonal MCI values decreased (by 12 units) between spring and summer at the mid-reach (SH3a) site where the historical median summer decrease has been 3 units (Appendix II). A typical decrease (of 11 units) was found at the lower reach site where a larger historical median summer decrease of 9 units has been recorded (Appendix II). The percentage compositions of 'tolerant' taxa were higher (by 17% and 18%) in the summer mid and lower reach communities. Seasonal communities at the mid-reach site (SH3a) shared 19 common taxa (61% of the 31 taxa found at this site in 2012-2013) compared with 15 shared common taxa (58% of the 26 taxa found in 2012-2013) at the lower reach site (Devon Road), a slightly more pronounced seasonal change in community structure at the lower-reach site. The two sites shared 17 common taxa (65% of the 26 taxa) in spring and 18 common taxa (60% of 30 taxa) in summer, indicative of slightly greater dissimilarity in spatial community structures in summer.

MCI score typically decreased (by 12 units) in a downstream direction in spring and similarly in summer (by 11 units), over a stream distance of 15.1 km downstream from the National Park boundary. These falls in MCI scores equated to rates of decline of 0.8 unit/km in spring increasing to 0.7 unit/km in summer, compared with a predicted rate of 0.5 unit/km over the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009).

Using the long-term median SEM MCI scores for both sites (Appendix II), the rate of overall decline has been 0.6 MCI unit/km over the surveyed length. Therefore rates of decline over the 2012–2013 period were slightly higher in both spring and summer than the average historical rate.

3.2.5 Waiwhakaiho River

An additional site was established in the upper reaches of the Waiwhakaiho River for the 2002-2003 SEM programme, to complement the three sites in the central to lower reaches of this large ringplain river, in recognition of its importance as a water resource and particularly its proximity to New Plymouth city. The site was established a short distance inside the National Park boundary at an elevation of 460 m asl. The results from the 2012-2013 surveys are presented in Table 134 and Table 135, Appendix I.

3.2.5.1 National Park site (WKH000100)

3.2.5.1.1 Taxa richness and MCI

Twenty surveys have previously been undertaken at this upper reach site just inside the National Park boundary in the Waiwhakaiho River between November 2002 and April 2012. These results are summarised in Table 20, together with the results from the current period, and illustrated in Figure 20.

Table 20 Results of previous surveys performed in the Waiwhakaiho River at National Park together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Apr 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WKH000100	20	4-29	19	115-147	129	21	133	22	125

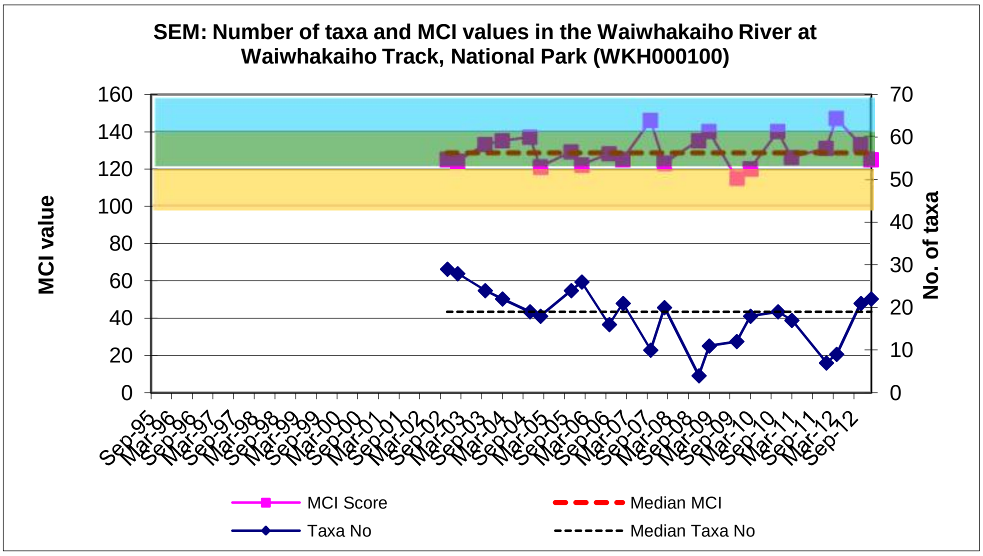


Figure 16 Numbers of taxa and MCI values in the Waiwhakaiho River at Egmont National Park

A wide range of richnesses (4 to 29 taxa) has been found, wider than might be expected, due to headwater erosion effects over the 2008-2009 period with a median richness of 19 taxa, but lower than typical richnesses (28 taxa) in ringplain streams and rivers near the National Park boundary. During the 2012-2013 period spring (21 taxa) and summer (22 taxa) richnesses were slightly above this median richness following continued recent post-headwater erosion recover, with minimal evidence of siltation remaining at this site.

MCI values have had a wider range (32 units) at this site than typical of a National Park boundary site, due in part to an atypically very high value in 2008 following a marked drop in richness and low values after the 2008-2009 headwater erosion period. The median value (129 units) has been slightly lower than typical of upper reach sites elsewhere on the ringplain (TRC, 1999b (updated, 2012)), and the spring, 2012 (133 units) and summer, 2013 (125 units) scores were 4 units higher and 4 units lower than the historical median respectively. They categorised this site as having 'very good' (spring and summer) health generically and, in terms of predictive relationships, 'expected' health for the upper reaches of a ringplain stream on both these occasions with taxa richnesses indicative of continued post-headwater erosion recover. The historical median score (129 units) placed this site in the 'very good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.5.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 21.

Table 21 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiwhakaiho River at the National Park between 1995 and April 2012 [20 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
EPHEMEROPTERA	*Coloburiscus*	7	2	10		
	Deleatidium	8	20	100	XA	XA
	Nesameletus	9	4	20		
PLECOPTERA	*Megaleptoperla*	9	6	30	A	A
	Zelandoperla	8	15	75	VA	VA
COLEOPTERA	Elmidae	6	18	90	VA	XA
TRICHOPTERA	*Aoteapsyche*	4	1	5		
	Costachorema	7	1	5	A	
	Hydrobiosella	9	1	5		
	Beraeoptera	8	5	25		
DIPTERA	*Aphrophila*	5	8	40		A
	Eriopterini	5	3	15		
	Maoridiamesa	3	1	5		
	Orthocladiinae	2	2	10		

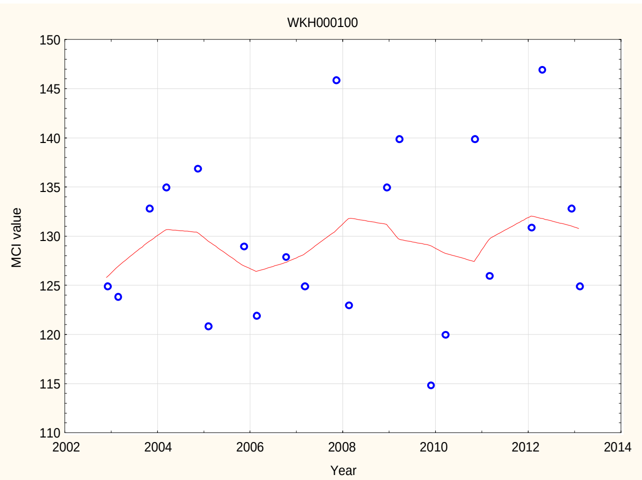
Prior to the current 2012-2013 period, 14 taxa had characterised the community at this site on occasions. These have comprised six 'highly sensitive', five 'moderately sensitive', and three 'tolerant' taxa i.e. a majority of 'sensitive' taxa as would be expected near the National Park boundary of a ringplain stream although numerically fewer dominant taxa than are typical in the upper reaches of a ringplain stream. Predominant taxa have included two 'highly sensitive' taxa [mayfly (*Deleatidium* on every sampling occasion) and stonefly (*Zelandoperla*)]; one 'moderately sensitive' taxon [elmid beetles]; but no 'tolerant' taxa. Five of these taxa were dominant in the spring 2012 community and four of these same taxa were again dominant in the summer 2013 community. No 'tolerant' taxa were dominant on either sampling occasion coincident with minimal periphyton substrate cover at this site. All taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 75% to 100% of past surveys.

3.2.5.1.3 Predicted stream 'health'

The Waiwhakaiho River site at the National Park is just inside the National Park boundary at an altitude of 460 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 131 (altitude) and 132 (distance) for this site. The historical site median (129 units) is only 2 units lower than the altitude prediction and 3 units lower than the distance predictive value, with the spring, 2012 survey score (133 units) was within two units of both predictive values and the summer, 2013 score (125 units) was insignificantly 6 to 7 units lower than both predictive values. Of the 22 surveys to date at this site, 55% of MCI scores have been less than 131 units while 41% have been greater than 132 units.

3.2.5.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eleven years of SEM results collected to date from the site in the Waiwhakaiho River at the National Park. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 17.



N = 22
Kendall tau = +0.101
p level = 0.511 (>FDR, p = 0.604)
N/S at $p < 0.05$

Figure 17 LOWESS trend plot of MCI data at the National Park site

No significant temporal trend in MCI scores has been found over the eleven year monitoring period at this site within the National Park. Smoothed scores consistently have indicated 'very good' generic (Table 1) river health over the period but, in terms of predictive relationships (Table 2) for a site in the upper reaches of a ringplain river at the boundary of the National Park, river health has remained as 'expected' while some individual scores indicative of 'worse than expected' health have followed headwater erosion events during the eleven year period.

3.2.5.2 Egmont Village site (WKH000500)

3.2.5.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Waiwhakaiho River at this mid-reach site at SH 3, Egmont Village (above the Mangorei Power Scheme) between October 1995 and April 2012. These results are summarised in Table 22, together with the results from the current period, and illustrated in Figure 18.

Table 22 Results of previous surveys performed in the Waiwhakaiho River at Egmont Village together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Apr 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WKH000500	34	14-32	23	87-122	109	22	112	27	112

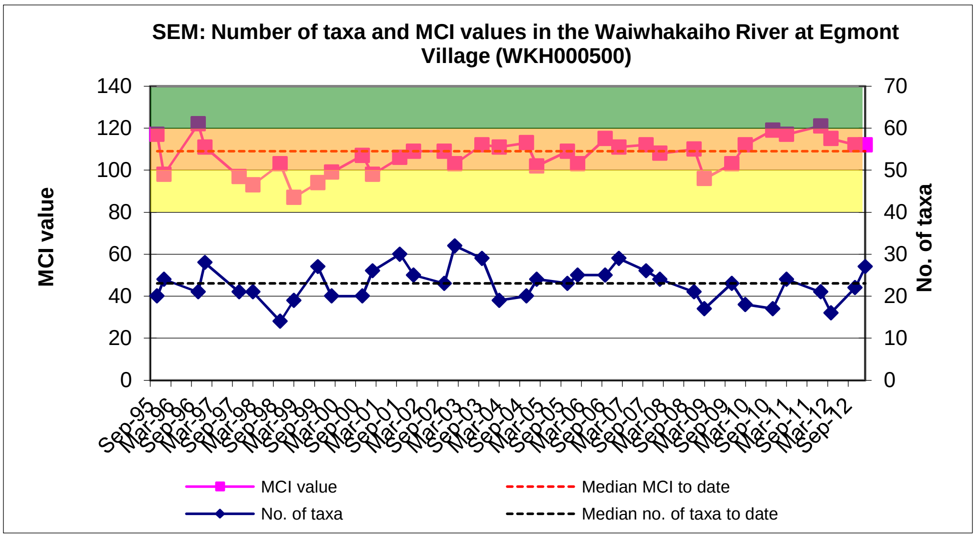


Figure 18 Numbers of taxa and MCI values in the Waiwhakaiho River at Egmont Village

A wide range of richnesses (14 to 32 taxa) has been found; wider than might be expected, with a median richness of 23 taxa (more representative of typical richnesses in the mid reaches of ringplain streams and rivers). During the 2012-2013 period spring (22 taxa) and summer (27 taxa) richnesses were relatively different (by 5 taxa) but near or above the median taxa number to date.

MCI values have had a slightly wider range (35 units) at this site than typical of sites in the mid reaches of ringplain rivers. The median value (109 units) has been relatively typical of mid reach sites elsewhere on the ringplain however, with the spring, 2012 (112 units) and summer, 2013 (112 units) scores typical for such a site and 3 units higher than the historical median. These scores categorised this site as having 'good' (spring and summer) health generically and, in terms of predictive relationships 'expected' (spring and summer) health for the mid reaches of a ringplain river. The historical median score (109 units) placed this site in the 'good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.5.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 23.

Table 23 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiwhakaiho River at Egmont Village between 1995 and April 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMATODA	Nematoda	3	1	3		
ANNELIDA	Oligochaeta	1	8	24		
EPHEMEROPTERA	*Coloburiscus*	7	9	26	A	
	Deleatidium	8	27	79	XA	XA
	Nesameletus	9	3	9		
PLECOPTERA	*Zelandoperla*	8	1	3	A	
COLEOPTERA	Elmidae	6	20	59	A	VA
MEGALOPTERA	*Archichauliodes*	7	2	6		
TRICHOPTERA	*Aoteapsyche*	4	22	65		VA
	Costachorema	7	11	32	A	A
	Hydrobiosis	5	5	15		A
	Neurochorema	6	5	15		
	Beraeoptera	8	1	3		
	Oxyethira	2	8	24		
	Pycnocentrodes	5	3	9	A	
DIPTERA	*Aphrophila*	5	27	79	A	A
	Eriopterini	5	2	6		
	Maoridiamesa	3	29	85	A	VA
	Orthocladiinae	2	32	94		A
	Tanytarsini	3	10	29		
	Empididae	3	2	6		
	Muscidae	3	4	12		
	Austrosimulium	3	1	3		

Prior to the current 2012-2013 period, 23 taxa had characterised the community at this site on occasions. These have comprised four 'highly sensitive', nine 'moderately sensitive', and ten 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and in comparison with the National Park site, a (downstream) increase in 'tolerant' taxa as would be expected in the mid reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; two 'moderately sensitive' taxa [elmid beetles and cranefly (*Aphrophila*)]; and three 'tolerant' taxa [free-living caddisfly (*Aoteapsyche*) and midges (*Maoridiamesa* and orthoclads)]. Eight of the historically characteristic taxa were dominant in the spring, 2012 and summer 2013 communities. These comprised two 'highly sensitive' taxa, five 'moderately sensitive' taxa, and one 'tolerant' taxon in spring, whereas one 'highly sensitive', four 'moderately sensitive', and three 'tolerant' taxa comprised the dominant taxa of the summer community i.e. fewer 'sensitive' and more 'tolerant' taxa in summer. Five of these eleven taxa were dominant in both spring and summer communities (Table 23). The predominant taxon [mayfly (*Deleatidium*)] remained the same in both surveys resulting in relatively high $SQMCI_s$ values (7.4 and 6.3 units).

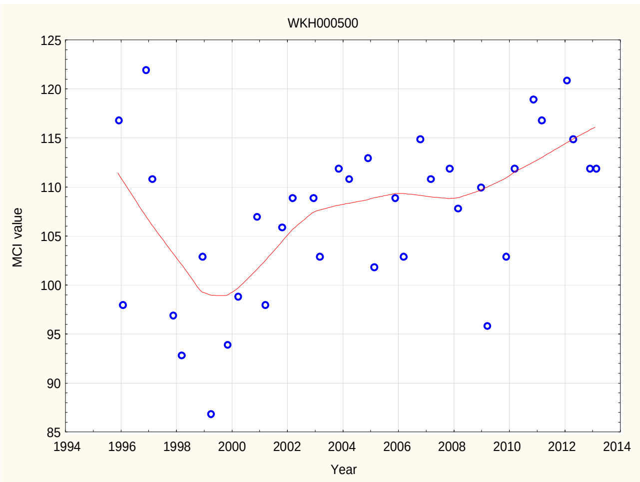
Of the predominant taxa in the 2012-2013 period, the 'highly sensitive' mayfly (*Deleatidium*), 'moderately sensitive' elmid beetles, and 'tolerant' caddisfly (*Aoteapsyche*) and orthoclad midges have characterised this site's communities on 59% to 85% of survey occasions to date.

3.2.5.2.3 Predicted stream 'health'

The Waiwhakaiho River site at Egmont Village is 10.6 km downstream of the National Park boundary at an altitude of 175 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 102 (altitude) and 105 (distance) for this site. The historical site median (109) is 7 units higher than the altitude prediction and 4 units higher than the distance predictive value while the spring, 2012 and summer 2013 survey scores (112 units) were higher than both predictive values, by 7 to 10 units. Of the 37 surveys to date at this site, 22% of MCI scores have been less than 102 units while 64% have been greater than 105 units.

3.2.5.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waiwhakaiho River at Egmont Village. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 19.



N = 36
Kendall tau = +0.348
p level = 0.003 [>FDR, p = 0.007]
Significant at $p < 0.05$ and $p < 0.01$ after FDR application

Figure 19 LOWESS trend plot at the Egmont Village site

An overall significant trend in MCI scores has been found during the eighteen year period. After some initial deterioration in scores, there has been a steady improvement since 1999. The change in the LOWESS-smoothed range (17 MCI units) has been of ecological significance over the period. While the smoothed scores were indicative of 'good' to 'fair' generic river health (Table 1) over the first five years, river health has consistently remained 'good' since 2000. In terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain river, health has been 'expected' for most of the period, improving toward 'better than expected' more recently.

3.2.5.3 Constance Street site (WKH000920)

3.2.5.3.1 Taxa richness and MCI

Thirty-five surveys have been undertaken in the Waiwhakaiho River at this lower reach site at Constance Street, New Plymouth (below the Mangorei Power Scheme) between 1995 and March 2011. These results are summarised in Table 24, together with the results from the current period, and illustrated in Figure 20.

Table 24 Results of previous surveys performed in the Waiwhakaiho River at Constance Street, New Plymouth, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to April 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WKH000920	35	12-29	20	71-110	95	17	108	24	96

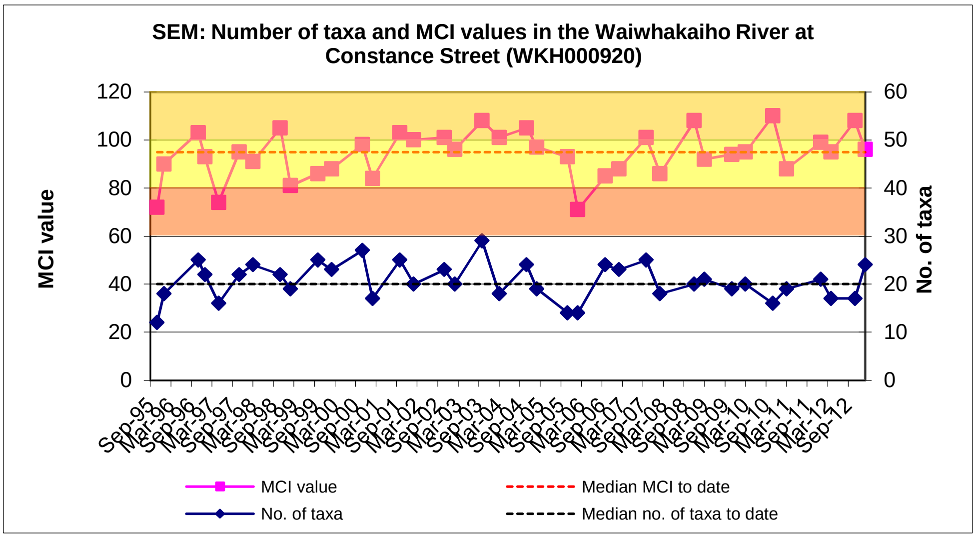


Figure 20 Numbers of taxa and MCI values in the Waiwhakaiho River at Constance Street

A wide range of richnesses (12 to 29 taxa) has been found with a median richness of 20 taxa (more representative of typical richnesses in the lower reaches of ringplain streams and rivers). During the 2012-2013 period spring (17 taxa) and summer (24 taxa) richnesses were relatively different but within four taxa of the median richness on both occasions.

MCI values have had a wide range (39 units) at this site, typical of sites in the lower reaches of ringplain streams and rivers. The median value (95 units) has been relatively typical of scores at lower reach sites elsewhere on the ringplain however (TRC, 1999). The spring, 2012 (108 units) and summer, 2013 (96 units) scores were significantly different (Stark, 1998), although relatively typical of scores for such a site. They were significantly higher than the historical median in spring when the score was within 2 units of the historical maximum and very similar to the median in summer. These scores categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1) and in terms of predictive relationships (Table 2) 'better than expected' health in spring and 'expected' health in summer for the lower reaches of a ringplain river. The historical median score (95 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.5.3.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 25.

Table 25 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiwhakaiho River at Constance Street between 1995 and April 2012 [35 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	3		
ANNELIDA	Oligochaeta	1	20	57		A
MOLLUSCA	*Potamopyrgus*	4	2	6		
CRUSTACEA	*Paratya*	3	1	3		
EPHEMEROPTERA	*Austroclima*	7	1	3		
	Coloburiscus	7	5	14		
	Deleatidium	8	17	49	XA	VA
COLEOPTERA	Elmidae	6	9	26	A	VA
	Staphylinidae	5	1	3		
TRICHOPTERA	*Aoteapsyche*	4	27	77	A	VA
	Costachorema	7	5	14	A	
	Hydrobiosis	5	6	17		A
	Neurochorema	6	1	3		
	Oxyethira	2	10	29		A
DIPTERA	*Aphrophila*	5	8	23		
	Maoridiamesa	3	17	49	A	
	Orthocladiinae	2	34	97	A	VA
	Tanytarsini	3	15	43		A
	Muscidae	3	2	6		
	Austrosimulium	3	4	11		

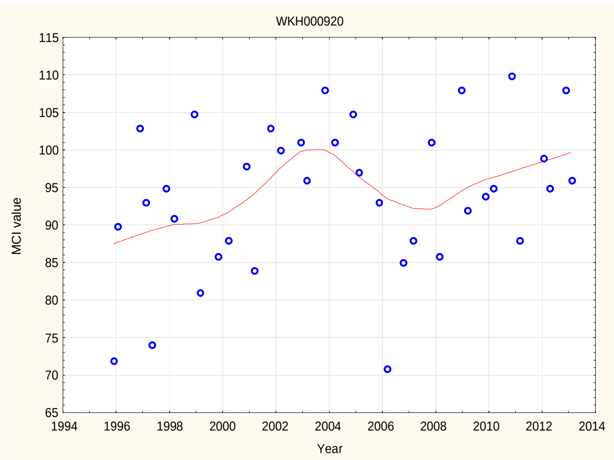
Prior to the current 2012-2013 period, 20 taxa had characterised the community at this site on occasions. These have comprised one 'highly sensitive', eight 'moderately sensitive', and eleven 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and a downstream increased proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain river. Predominant taxa have included no 'highly sensitive' taxa; no 'moderately sensitive' taxa; but three 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), and orthoclad midges]. Only six of the historically characteristic taxa were dominant in the spring 2012 community. These comprised one 'highly sensitive' taxon, two 'moderately sensitive' taxa, and three 'tolerant' taxa, whereas one 'highly sensitive, two 'moderately sensitive', and five 'tolerant' taxa comprised the dominant taxa of the summer, 2013 community. Four of these ten taxa were dominant in both spring and summer communities (Table 25). Increases in summer seasonal dominances by 'tolerant' taxa and a reduction in the abundance of the 'highly sensitive' mayfly (*Deleatidium*) were reflected in the significant decrease (2.7 units) in the summer $SQMCI_s$ score (Tables 134 and 135). The 'highly sensitive' mayfly (*Deleatidium*), 'moderately sensitive' elmid beetles, and 'tolerant' caddisfly (*Aoteapsyche*) and orthoclad midges which were pre-dominant in spring and/or summer surveys, had characterised this site's communities on 26% to 97% of past survey occasions.

3.2.5.3.3 Predicted stream 'health'

The Waiwhakaiho River site at Constance Street, New Plymouth is 26.6 km downstream of the National Park boundary at an altitude of 20 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 86 (altitude) and 95 (distance) for this site. The historical site median (95) is 9 units higher than the altitude prediction and equal with the distance predictive value. The spring, 2012 survey score (108 units) was significantly 22 and 13 units higher than the altitude and distance predictive values respectively while the summer 2013 score (96 units) was 10 units higher than the predictive altitude value and one unit above the predicted distance value. Of the 37 surveys to date at this site, 16% of MCI scores have been less than 86 units while 43% have been greater than 95 units.

3.2.5.3.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waiwhakaiho River at Constance Street. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 21.



N = 37
Kendall tau = +0.187
p level = 0.103 [> FDR, p = 0.168]
N/S at $p < 0.05$

Figure 21 LOWESS trend plot at the Constance Street site

The overall trend in MCI scores has not been statistically significant for the period, due mainly to some decline in scores after 2005. The LOWESS-smoothed range of scores (12 units) indicates variability of some ecological significance. Smoothed MCI scores indicated 'fair' generic river health (Table 1) improving toward 'good' health (after a small increase in summer residual flow releases by the TrustPower Mangorei HEP scheme) before returning to 'fair' health over recent years. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, river health improved from 'expected' early in the period toward 'better than expected' for a brief period (2003-2004) before returning to 'expected', and then a recent trend toward 'better than expected'.

3.2.5.4 Site adjacent to Lake Rotomanu (WKH000950)

3.2.5.4.1 Taxa richness and MCI

Thirty-three surveys have been undertaken in the Waiwhakaiho River at this lower reach site adjacent to Lake Rotomanu between March 1997 and April 2012. These results are summarised in Table 26, together with the results from the current period, and illustrated in Figure 22.

Table 26 Results of previous surveys performed in the Waiwhakaiho River the site adjacent to Lake Rotomanu, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to April 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WKH000950	33	12-28	21	70-111	88	30	96	26	89

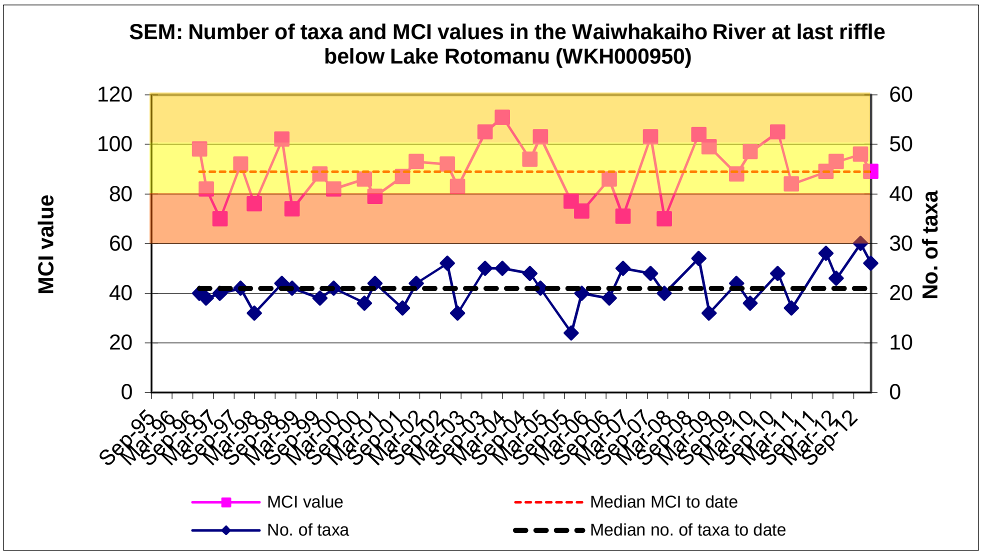


Figure 22 Numbers of taxa and MCI values in the Waiwhakaiho River at Lake Rotomanu

A wide range of richnesses (12 to 28 taxa) has been found; wider than might be expected, with a median richness of 21 taxa. During the 2012-2013 period spring (30 taxa) richness was 4 taxa more than found later in summer and spring richness was seven taxa higher than the median richness and two taxa above the historical maximum.

MCI values have had a wide range (41 units) at this site but typical of sites in the lower reaches of ringplain streams. The median value (88 units) has been relatively typical of lower reach sites elsewhere on the ringplain (TRC, 1999b (updated 2012)). The spring 2012 (96 units) and summer, 2013 (89 units) scores, although typical for such a site, were very similar (in summer) and eight units higher in spring than the historical median. These scores categorised this site as having 'fair' (spring and summer) health generically (Table 1) and in terms of predictive relationships (Table 2) 'expected' health in spring and in summer for the lower reaches of a ringplain river. The historical median score (88 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.5.4.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 27.

Table 27 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiwhakaiho River at the site adjacent to Lake Rotomanu between 1995 and April 2012 [33 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	9		
ANNELIDA	Oligochaeta	1	27	82	A	A
MOLLUSCA	*Physa*	3	1	3		
	Potamopyrgus	4	9	27	A	A
CRUSTACEA	*Paratya*	3	6	18		A
EPHEMEROPTERA	*Coloburiscus*	7	1	3		
	Deleatidium	8	9	27	VA	
COLEOPTERA	Elmidae	6	5	15	A	A
TRICHOPTERA	*Aoteapsyche*	4	22	67	VA	XA
	Costachorema	7	2	6		
	Hydrobiosis	5	3	9		
	Oxyethira	2	13	39	A	A
DIPTERA	*Aphrophila*	5	10	30	VA	VA
	Maoridiamesa	3	16	48	VA	
	Orthocladiinae	2	33	100	VA	VA
	Tanytarsini	3	14	42		VA
	Empididae	3	1	3		
	Muscidae	3	1	3		
	Austrosimulium	3	1	3		

Prior to the current 2012-2013 period, 19 taxa had characterised the community at this site on occasions. These have comprised one 'highly sensitive', five 'moderately sensitive', and thirteen 'tolerant' taxa i.e. a minority of 'sensitive' taxa and a high proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain river. Predominant taxa have included no 'highly sensitive' taxa; no 'moderately sensitive' taxa; but three 'tolerant' taxa (oligochaete worms, net-building caddisfly (*Aoteapsyche*), and orthoclad midges). Nine of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised one 'highly sensitive' taxon, two 'moderately sensitive' taxa, and six 'tolerant' taxa. No 'highly sensitive', two 'moderately sensitive', and seven 'tolerant' taxa comprised the dominant taxa of the summer, 2013 community. Seven of these eleven taxa were dominant in both spring and summer communities (Table 27). The absence of the 'highly sensitive' mayfly taxon from the dominant summer taxa was reflected in the drop of 0.6 unit in $SQMCI_s$ scores between spring and summer (Tables 134 and 135) with this decrease also being due to summer increase in abundance of one 'tolerant' taxon [caddisfly (*Aoteapsyche*)] in particular.

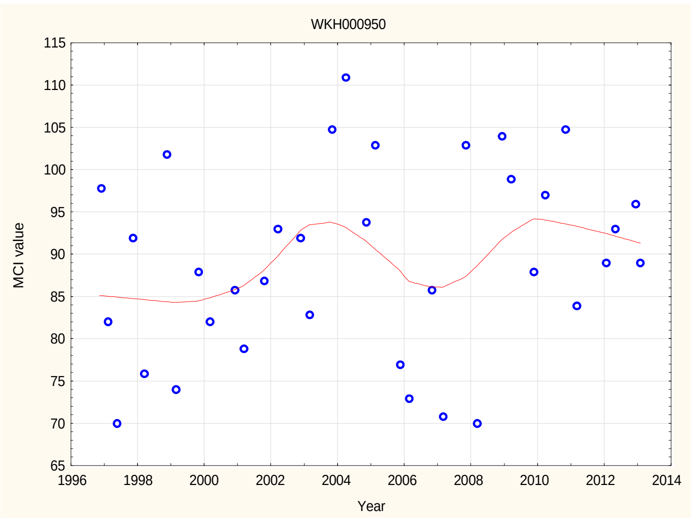
The 'highly sensitive' mayfly (*Deleatidium*), 'moderately sensitive' cranefly (*Aphrophila*), and 'tolerant' net-building caddisfly (*Aoteapsyche*) and *Maoridiamesa*, tanytarsids, and orthoclad midges have characterised this site's communities on 27% to 100% of past survey occasions.

3.2.5.4.3 Predicted stream 'health'

The Waiwhakaiho River at the site adjacent to Lake Rotomanu is 28.4 km downstream of the National Park boundary at an altitude of 2 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 85 (altitude) and 94 (distance) for this site. The historical site median (88) is 3 units higher than the altitude prediction and 7 units lower than the distance predictive value. The spring 2012 survey score (96 units) was slightly above both of these predictive values while the summer score (89 units) was midway between the predictive altitude and the distance values. Of the 35 surveys to date at this site, 31% of MCI scores have been less than 85 units while 29% have been greater than 94 units.

3.2.5.4.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waiwhakaiho River adjacent to Lake Rotomanu. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 23.



N = 35
Kendall tau = +0.166
p level = 0.161 [> FDR, p = 0.239]
N/S at $p < 0.05$

Figure 23 LOWESS trend plot of MCI data at the site adjacent to Lake Rotomanu

Overall, MCI scores have shown no statistically significant trend, despite a significant improvement during the first nine years of the programme (TRC, 2006b). Since 2004, there has been a steady decline in scores toward scores typically found in the first two years of the programme followed by a more recent improvement, a similar trend found at the nearest upstream site (Constance St). The LOWESS-smoothed range of scores (10 units) is not ecologically significant with more marked variability over the 2002 to 2010 period. Smoothed MCI scores have indicated 'fair' generic stream 'health' (Table 1) throughout the period. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain river, health has remained in the 'expected' category over the entire period.

3.2.5.5 Discussion

Seasonal MCI values typically deteriorated between spring and summer at three sites by 8, 12, and 7 units respectively in a downstream direction whereas scores at the mid-catchment site atypically remained identical in summer. Seasonal communities shared 48% of 29 taxa present at the upper site 58% of 31 taxa at the mid reach site, and in the lower reaches, 54% of 28 taxa at the Constance Street site and 56% of 36 taxa at the furthest downstream site. The typical decrease in seasonal faunal similarities in a downstream direction, as might be anticipated given wider variability in seasonal substrate periphyton coverage and water temperature in the lower reaches was not so apparent between seasons in the 2012-2013 period, particularly at the upper site.

The MCI scores fell in a downstream direction between the upper site and the furthest downstream lower reaches site by 37 units in spring and 36 units in summer, over a river distance of 28.7 km. These seasonal falls in MCI scores equated to rates of decline of 1.3 units/km (spring) and 1.3 units/km (summer), compared with a predicted rate of 1.3 unit/km for the equivalent length and reach of a National Park-sourced river (Stark and Fowles, 2009). This was atypical of the trend of past summer seasonal rates of decline which have usually been higher.

Between the upper and mid-reach sites, the spring (2.0 units/km) and summer (1.2 units/km) rates of decline were lower in spring and far lower in summer than the predicted rate (2.6 units/km) for the equivalent river reach. For the mid-reach to lower reach sites, spring (0.9 unit/km) and summer (1.3 units/km) rates of decline were slightly higher (spring) and well above (summer) the predicted rate of 0.6 unit/km.

Using the long-term median SEM MCI scores for each site (Appendix II) the rates of decline between upper and mid catchment and between mid catchment and lower river sites have been about 1.4 and 1.1 units per km respectively with an overall average rate of decline of 1.3 MCI units/km over the river's length. Therefore overall rates of decline over the 2012-2013 period were identical in both spring and summer to the typical rate prior to 2012, but in the upper to mid reaches, spring MCI rate of decline was well above the typical rate.

Community composition varied markedly through the length of the river surveyed. A total of 41 taxa were recorded in spring of which only eight taxa were present at all four sites. These included one 'highly sensitive', six 'moderately sensitive', and one 'tolerant' taxa with only the 'highly sensitive' mayfly (*Deleatidium*) and 'moderately sensitive' elmid beetles abundant at all four sites. One other 'moderately sensitive' taxon was abundant at three sites and one of the 'tolerant' taxa was abundant at three sites (mid and lower reaches of the river. A similar total of 42 taxa was found along the river's length by the summer survey of which only nine taxa were present at all four sites. These were relatively similar to the eight widespread taxa in spring with the addition of one 'highly sensitive' taxon, loss of one 'moderately sensitive' taxon, and addition of one 'tolerant' taxon. Only the one 'moderately sensitive' elmid beetle was abundant at all four sites. These dissimilarities in spatial community structure along the length of the Waiwhakaiho River were less pronounced between seasons than usual.

3.2.6 Mangorei Stream

A site was established in the lower reaches of the Mangorei Stream, near the confluence with the Waiwhakaiho River, for the SEM programme of 2002-2003, in recognition of the importance of this catchment as the only major inflow to the lower reaches of the river below significant HEP and New Plymouth District Council water supply abstractions. The results from the surveys performed in the 2012-2013 monitoring year are presented in Table 136 and Table 137, Appendix I.

3.2.6.1 SH3 site (MGE000970)

3.2.6.1.1 Taxa richness and MCI

Twenty surveys have been undertaken at this lower reach site in the Mangorei Stream between November 2002 and April 2012. These results are summarised in Table 28, together with the results from the current period, and illustrated in Figure 24.

Table 28 Results of previous surveys performed in the Mangorei Stream at SH 3 together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Apr 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MGE000970	20	22-33	29	86-113	104	23	97	25	102

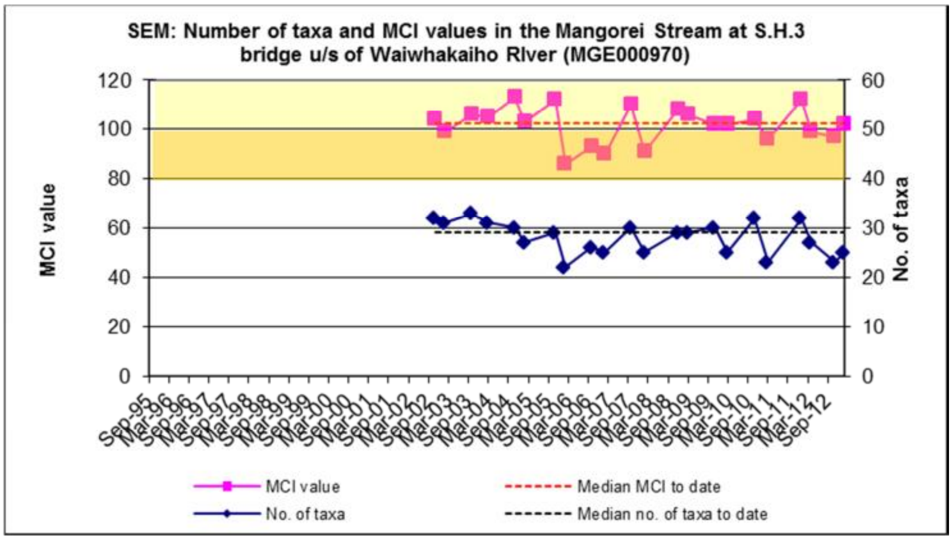


Figure 24 Numbers of taxa and MCI values in the Mangorei Stream at SH3

A moderate range of richnesses (22 to 33 taxa) has been found with a relatively high median richness of 29 taxa (more representative of typical richnesses in upper and middle reaches of ringplain streams and rivers. During the 2012-2013 period, spring (23 taxa) richness was well below this median richness and two taxa fewer than found by the summer survey.

MCI values have had a relatively wide range (27 units) at this site, more typical of a site in the lower reaches of a ringplain stream. However, the median value (104 units) has been more typical of mid-reach sites elsewhere on the ringplain, and the spring, 2012 (97 units) and summer, 2013 (102 units) scores were within 7 units of the historical median. These scores categorised this site as having 'fair' health (spring) and 'good' health generically (Table 1) in summer and, in terms of predictive relationships (Table 2), 'expected' (spring and summer) health for the lower reaches of a ringplain stream. The historical median score (104 units) placed this site in the 'good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.6.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 29.

Table 29 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Mangorei Stream at SH3 between 2002 and April 2012 [20 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	2	10		
ANNELIDA	Oligochaeta	1	14	70		
MOLLUSCA	*Potamopyrgus*	4	4	20		A
EPHEMEROPTERA	*Austroclima*	7	11	55	A	VA
	Coloburiscus	7	9	45	A	A
	Deleatidium	8	13	65	XA	A
PLECOPTERA	*Zelandobius*	5	6	30	A	
	Zelandoperla	8	1	5		A
COLEOPTERA	Elmidae	6	12	60	A	A
MEGALOPTERA	*Archichauliodes*	7	13	65	A	A
TRICHOPTERA	*Aoteapsyche*	4	18	90	VA	XA
	Costachorema	7	3	15		
	Hydrobiosis	5	9	45	A	A
	Neurochorema	6	5	25		
	Confluens	5	3	15		
	Oxyethira	2	7	35		A
	Pycnocentrodes	5	6	30		
DIPTERA	*Aphrophila*	5	15	75	A	VA
	Maoridiamesa	3	8	40	A	
	Orthocladiinae	2	20	100	A	A
	Tanytarsini	3	13	65		A
	Empididae	3	3	15		
	Muscidae	3	1	5		
	Austrosimulium	3	16	80	A	A

Prior to the current 2012-2013 period, 24 taxa have characterised the community at this site on occasions. These have comprised two 'highly sensitive', eleven 'moderately sensitive', and eleven 'tolerant' taxa i.e. a more even balance of 'sensitive' and 'tolerant' taxa as would be expected in the lower reaches of a ringplain stream. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; five 'moderately sensitive' taxa [mayfly (*Austroclima*), elmid beetles, dobsonfly (*Archichauliodes*), free-living caddisfly (*Hydrobiosis*), and cranefly

(*Aphrophila*)]; and five 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), midges (orthoclads and tanytarsids), and sandfly (*Austrosimulium*)].

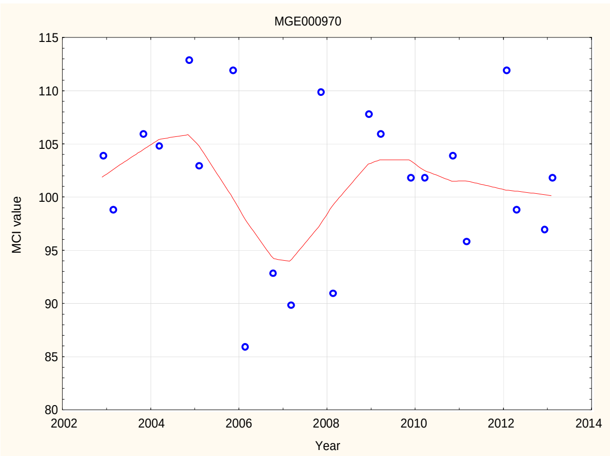
Eight of these predominant taxa were dominant in the spring, 2012 community together with four of the other historically characteristic taxa. The summer, 2013 community was characterised by slightly fewer (ten) of the taxa dominant in spring, together with an additional one 'highly sensitive' and three 'tolerant' taxa (Table 29). In particular, the marked reduction in numerical abundance of the 'highly sensitive' mayfly was reflected in the 1.9 units decrease in $SQMCI_s$ score recorded by the summer survey (Tables 136 and 137). All taxa recorded as very abundant during spring and/or summer had characterised this site's communities on 55% to 90% of past surveys.

3.2.6.1.3 Predicted stream 'health

The Mangorei Stream site at SH3 is 15.6 km downstream of the National Park boundary at an altitude of 90 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 94 (altitude) and 101 (distance) for this site. The historical site median (104 units) is 10 units higher than the altitude prediction and 3 units above the distance predictive value. The spring 2012 score (97 units) was within 4 units of these predictive values while the summer 2013 survey score (102 units) was higher by 8 units than the predictive value for altitude and one unit higher than the predictive value for distance. Of the 22 surveys to date at this site, 18% of MCI scores have been less than 94 units while 64% have been greater than 101 units.

3.2.6.1.4 Temporal trends in 1995 to 2012 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eleven years (2002-2013) of SEM results collected to date from the site in the Mangorei Stream at SH3. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 25.



N = 22
Kendall tau = -0.114
p level = 0.457 [>FDR, p = 0.552]
N/S at p < 0.05

Figure 25 LOWESS trend plot of MCI data at the SH3 site

While MCI scores showed slight initial improvement over the first three years, followed by a steady decline, more recently there has been further improvement in scores towards those recorded earlier in the programme. The very slight overall decline over the eleven year period has not been a statistically significant trend at this site. The LOWESS-smoothed range of scores (12 units) has been indicative of marginal ecological significance in variability. During the period, these smoothed MCI scores have been consistently indicative of 'good' generic stream health (Table 1) with some deterioration to 'fair' health between 2006 and 2008 prior to more recent recovery. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, stream health has been in the 'expected' category almost throughout the eleven year period, bordering on 'better than expected' for a short period in 2004-2005, coincident with good riparian vegetation cover at this site.

3.2.6.2 Discussion

Seasonal MCI values atypically increased between spring and summer (by 5 units) at this lower reach (SH3) site where (more typically) the historical median summer decrease has been 7 units (Appendix II). This was coincident with the percentage composition of 'tolerant' taxa decreasing (by 3%) in the summer community. Seasonal communities at this site shared 21 common taxa (78% of the 27 taxa found at this site in 2012-2013), a higher than usual percentage of common taxa for spring and summer seasonal surveys.

3.2.7 Manganui River

The results found by the 2012-2013 surveys are presented in Tables 138 and 139, Appendix I.

3.2.7.1 State Highway 3 site (MGN000195)

3.2.7.1.1 Taxa richness and MCI

Thirty-six surveys have been undertaken at this mid reach site in the Manganui River between September 1995 and April 2012. These results are summarised in Table 30 together with the results from the current period, and illustrated in Figure 26.

Table 30 Results of previous surveys performed in the Manganui River u/s of railway bridge (SH 3), together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Apr 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MGN000195	36	12-26	21	113-143	126	21	135	18	139

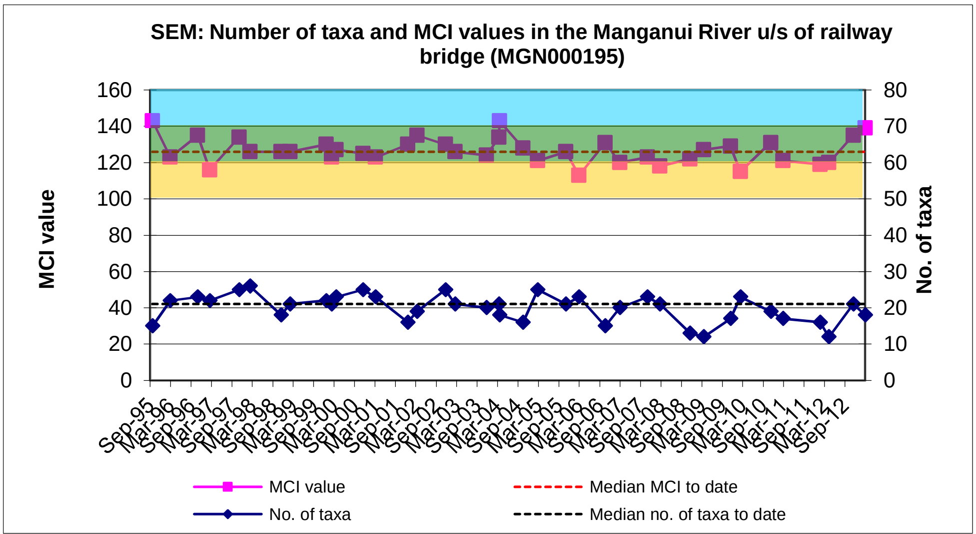


Figure 26 Numbers of taxa and MCI values in the Manganui River above the railway bridge (SH3)

A wide range of richnesses (12 to 26 taxa) has been found, with a median richness of 21 taxa (slightly lower than typical richnesses in the mid-reaches of ringplain streams and rivers). During the 2012-2013 period richnesses decreased between spring (21 taxa) and summer (18 taxa) but were within three taxa of this median richness.

MCI values have had a relatively wide range (30 units) at this site, slightly wider than typical of a site in the mid reaches of a ringplain stream. The median value (126 units) was higher than has been typical of mid-reach sites elsewhere on the ringplain. The spring, 2012 (135 units) and summer, 2013 (139 units) scores were 9 units and a significant (Stark, 1998) 13 units higher respectively than the historical median. These scores categorised this site as having 'very good' health generically (Table 1) in spring and in summer and, in terms of predictive relationships (Table 2), 'better than expected' health in spring and in summer for the mid reaches of a ringplain river. The historical median score (126 units) placed this site in the 'very good' and 'better than expected' categories for generic and predictive methods of assessment respectively.

3.2.7.1.2 Community composition

Characteristic macroinvertebrate taxa (abundant) in the communities at this site prior to the 2012-2013 period are listed in Table 31.

Table 31 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Manganui River at SH3 between 1995 and April 2012 [36 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
EPHEMEROPTERA	*Austroclima*	7	3	8		
	Coloburiscus	7	26	72	A	A
	Deleatidium	8	35	97	XA	XA
	Nesameletus	9	23	64	A	VA
PLECOPTERA	*Acroperla*	5	1	3		
	Megaleptoperla	9	1	3		
	Zelandoperla	8	9	25	A	
COLEOPTERA	Elmidae	6	33	92	XA	VA
MEGALOPTERA	*Archichauliodes*	7	4	11		
TRICHOPTERA	*Aoteapsyche*	4	15	42		A
	Hydrobiosis	5	1	3		
	Beraeoptera	8	8	22		
	Pycnocentrodes	5	1	3		
DIPTERA	*Aphrophila*	5	20	56		A
	Eriopterini	5	3	8		
	Austrosimulium	3	1	3		

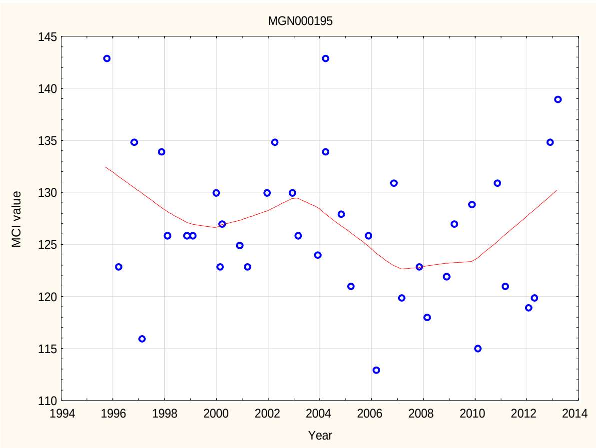
Prior to the current 2012-2013 period, 16 taxa have characterised the community at this site on occasions. These have comprised five 'highly sensitive', nine 'moderately sensitive', and two 'tolerant' taxa i.e. a higher proportion of 'sensitive' taxa than might be expected in the mid-reaches of a ringplain stream. Predominant taxa have included two 'highly sensitive' taxa [mayflies (*Deleatidium* and *Nesameletus*)]; and three 'moderately sensitive' taxa [elmid beetles, mayfly (*Coloburiscus*), and cranefly (*Aphrophila*)], but no 'tolerant' taxa. Four of these predominant taxa were dominant in the spring, 2012 community together with one of the other historically characteristic taxon. The summer, 2013 community was characterised by four of the same taxa dominant in spring, together with one additional 'moderately sensitive' and one 'tolerant' taxa, which previously had been characteristic of this site's communities (Table 31). Only minor seasonal differences in numerically dominant taxa compositions were reflected in the seasonal $SQMCI_s$ values (7.0 and 7.6 units) which were relatively similar (Tables 138 and 139). The three taxa recorded as extremely or very abundant during spring and/or summer had characterised this site's communities on 64% to 97% of past surveys.

3.2.7.1.3 Predicted stream 'health'

The Manganui River site at SH3 is 8.7 km downstream of the National Park boundary at an altitude of 330 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 118 (altitude) and 107 (distance) for this site. The historical site median (126 units) is 7 units higher than the altitude prediction and a significant (Stark, 1998) 19 units above the distance predictive value. The spring, 2012 survey score (135 units) was significantly higher by 17 to 28 units than both predictive values while the summer, 2013 score (139 units) was much higher (by 21 to 32 units) than the predictive values. Of the 38 surveys to date at this site, no MCI scores have been less than 107 units while 89% have been greater than 118 units.

3.2.7.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Manganui River at SH3. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 27.



N = 38
Kendall tau = - 0.137
p level = 0.222 [>FDR, p = 0.306]
Not significant at $p < 0.05$;

Figure 27 LOWESS trend plot of MCI data at the SH3 site

A trend of slight overall decrease in MCI scores was identified particularly over the first twelve years which, however has not been statistically significant for the eighteen year period. The LOWESS-smoothed scores (range of 10 units) represented a marginal ecological significance in terms of variability. These smoothed MCI scores consistently indicated 'very good' generic river health (Table 1) over the entire eighteen year period. In terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain stream, river health has been 'better than expected' throughout the eighteen year period.

3.2.7.2 Bristol Road site (MGN000427)

3.2.7.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this lower reach site at Bristol Road in the Manganui River between October 1995 and April 2012. These results are summarised in Table 32 together with the results from the current period, and illustrated in Figure 28.

Table 32 Results of previous surveys performed in the Manganui River at Bristol Road together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Apr 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MGN000427	34	15-26	20	77-115	98	15	111	19	96

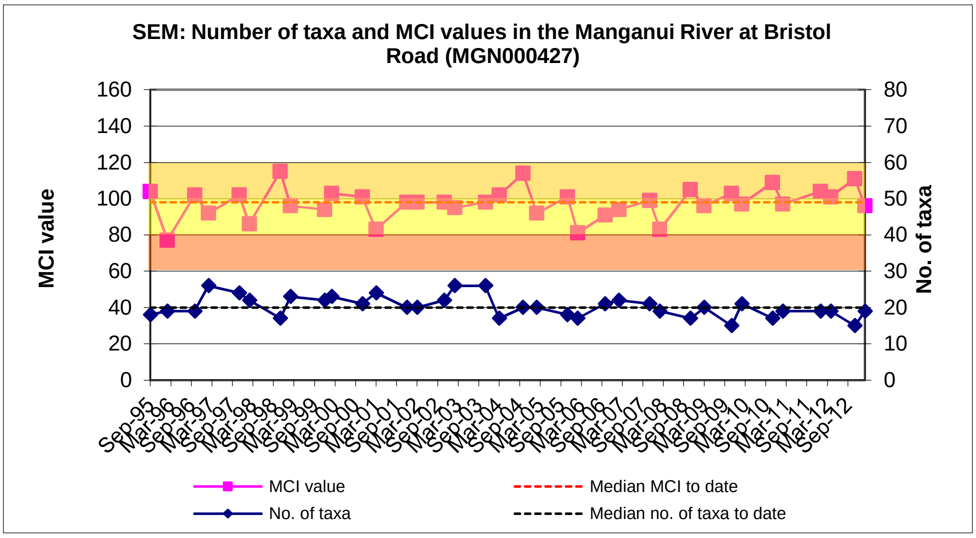


Figure 28 Numbers of taxa and MCI values in the Manganui River at Bristol Road

A moderate range of richnesses (15 to 26 taxa) has been found with a median richness of 20 taxa which is representative of typical richnesses in ringplain streams and rivers in the lower reaches. During the 2012-2013 period, spring (15 taxa) and summer (19 taxa) richnesses were less than the historical median richness in spring, coincident with minimal substrate periphyton cover, and similar to median richness in summer, when periphyton substrate cover was widespread.

MCI scores have had a wide range (38 units) at this site, typical of sites in the lower reaches of streams elsewhere on the ringplain although this site was located at an atypically higher altitude of 140 m asl for a lower reach site more than 37 km downstream from the National Park. The median value (98 units) has been higher than typical of lower reach sites. The spring 2012 (111 units) score was significantly 13 units higher than the historical median while the summer score (96 units) was 2 units lower than the historical median. These scores categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower reaches of a ringplain river. The historical median score (98 units) placed this site in the 'fair' category for generic and 'expected' category for predictive methods of assessment.

3.2.7.2.2 Community composition

Characteristic macroinvertebrate taxa abundant in the communities at this site prior to the 2012-2013 period are listed in Table 33.

Table 33 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Manganui River at Bristol Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	2	6		
ANNELIDA	Oligochaeta	1	16	47		
EPHEMEROPTERA	*Coloburiscus*	7	6	18	A	
	Deleatidium	8	18	53	XA	VA
COLEOPTERA	Elmidae	6	11	32		VA
MEGALOPTERA	*Archichauliodes*	7	2	6		
TRICHOPTERA	*Aoteapsyche*	4	24	71		XA
	Costachorema	7	4	12		
	Hydrobiosis	5	10	29		A
	Neurochorema	6	2	6		
	Oxyethira	2	7	21		
DIPTERA	*Aphrophila*	5	16	47		VA
	Maoridiamesa	3	15	44		VA
	Orthocladiinae	2	34	100		VA
	Tanytarsini	3	11	32		
	Empididae	3	2	6		
	Muscidae	3	6	18		
	Austrosimulium	3	7	21		

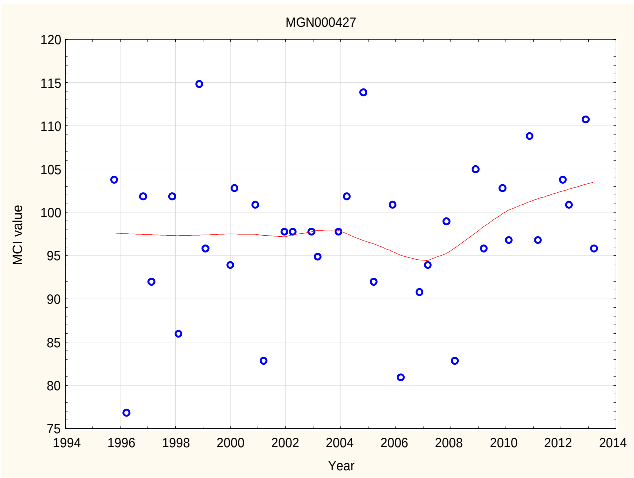
Prior to the current 2012-2013 period 18 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', seven 'moderately sensitive', and ten 'tolerant' taxa i.e. a majority of 'tolerant' taxa but a slightly lower proportion than might be expected in the lower reaches of a ringplain river coincidental with this site's relatively high elevation above sea level. Predominant taxa have included one 'highly sensitive' taxon, no 'moderately' sensitive' taxa, and three 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), and orthoclad midges]. Only two of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised one 'highly sensitive' and one 'moderately sensitive' taxa, whereas seven taxa (one 'highly sensitive', three 'moderately sensitive', and three 'tolerant' taxa) comprised the dominant taxa in the summer, 2013 community. Only one taxon was dominant in both spring and summer communities (Table 33) while the difference of 3.3 units in $SQMCI_s$ scores recorded between seasons (Tables 138 and 139) was due principally to a reduction in abundance of the mayfly, *Deleatidium* in summer and marked increases in abundances of several 'tolerant' taxa. Those taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 32% to 100% of past surveys.

3.2.7.2.3 Predicted stream 'health'

The Manganui River site at Bristol Road is 37.9 km downstream of the National Park boundary at an altitude of 140 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National park boundary (Stark and Fowles, 2009) predict MCI values of 99 (altitude) and 91 (distance) for this site. The historical site median (98 units) is very similar to the altitude prediction and 7 units higher than the predictive distance value, while the spring, 2012 survey score (111 units) was significantly higher than both the predictive values. The summer score (96 units) was within 5 units of these predictive values. Of the 34 surveys to date at this site, only 15% of MCI scores have been less than 91 units while 41% have been greater than 99 units.

3.2.7.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Manganui River at Bristol Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 29.



N = 36
Kendall tau = +0.098
p level = 0.398 [>FDR, p = 0.524]
N/S at $p < 0.05$

Figure 29 LOWESS trend plot at the Bristol Road site

The very slight overall positive trend in MCI scores was not statistically significant. Neither was the ecological variability in LOWESS-smoothed scores of 9 units. The smoothed MCI scores were indicative of 'fair' generic river health at this site almost throughout the eighteen year period improving to 'good' in the last three years. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain river, 'health' has remained in the 'expected' category throughout the period.

3.2.7.3 Discussion

Seasonal MCI values atypically remained very similar between spring and summer at the mid-reach (SH3) site where historical seasonal median scores have decreased in summer by 7 units (Appendix II). A more typical decrease, although relatively high (15 units), was recorded at the lower reach site where the historical median summer score has been 6 units lower (Appendix II). The percentage composition of 'tolerant' taxa decreased in the summer community at the mid reach site but only by 3%, whereas it increased by 20% at the lower reach site in summer. Seasonal communities at the mid-reach site (SH3) shared 14 common taxa (56% of the 25 taxa found at this site in 2012-2013) compared with 11 shared common taxa (48% of the 23 taxa found in 2012-2013) at the lower reach site (Bristol Road), a more pronounced seasonal change in community structure at the lower-reach site. The two sites shared 11 common taxa (44% of the 25 taxa) in spring and only 8 common taxa (29% of 28 taxa) in summer, indicative of the dissimilarity in spatial community structures in spring and more particularly in summer.

MCI score typically fell in a downstream direction in both spring (by 24 units) and much more so in summer (by 43 units), over a stream distance of 29.2 km downstream from the National Park boundary. These falls in MCI scores equated to rates of decline of 0.8 unit/km in spring increasing to 1.5 unit/km in summer, compared with a predicted rate of 0.5 unit/km over the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009).

Using the long-term median SEM MCI scores for both sites (Appendix II), the rates of decline between mid catchment and lower river sites has been about 1.0 MCI unit/km over the surveyed length. Therefore rates of decline over the 2012–2013 period were slightly lower in spring and far higher in summer than the historical average rate.

3.2.8 Maketawa Stream

Two sites, originally surveyed as components of the Maketawa catchment baseline investigation (Stark, 2003), were included in the 2002-03 SEM programme in recognition of the fisheries significance of this sub-catchment of the Manganui River catchment. The results from the surveys performed in the 2012-2013 monitoring year are presented in Table 140 and 141, Appendix I.

3.2.8.1 Derby Road site (MKW000200)

3.2.8.1.1 Taxa richness and MCI

Twenty-five surveys have been undertaken at this upper reach site in the Maketawa Stream between March 1998 and April 2012. These results are summarised in Table 34 together with the results from the current period, and illustrated in Figure 30.

Table 34 Results of previous surveys performed in the Maketawa Stream at Derby Road together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to April 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MKW000200	25	8-33	23	100-141	128	21	139	19	128

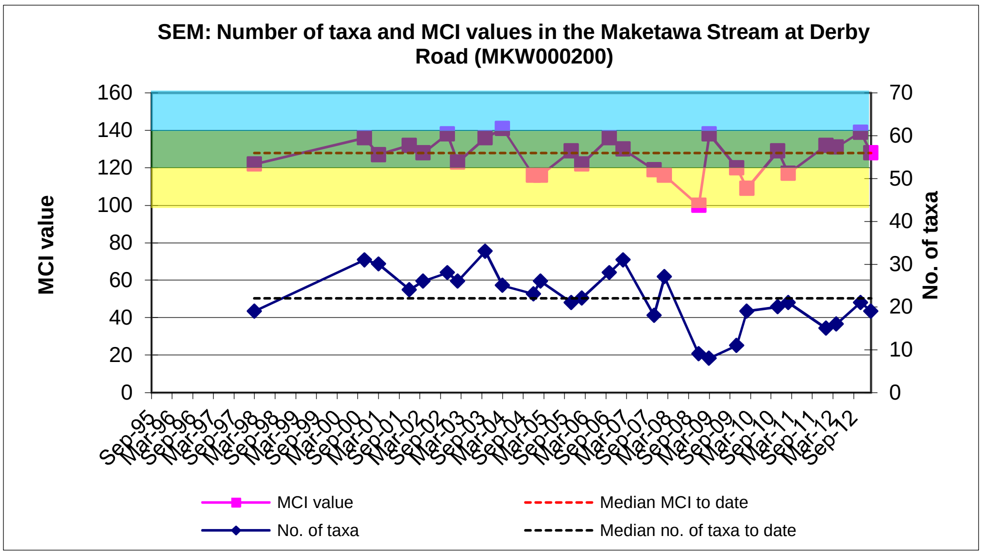


Figure 30 Number of taxa and MCI values in the Maketawa Stream at Derby Road

A very wide range of richnesses (8 to 33 taxa) has been found as a result of the impacts of headwater erosion events, with a median richness of 23 taxa (slightly lower than typical richnesses found in the upper reaches of ringplain streams and rivers). During the 2012-2013 period, spring (21 taxa) and summer (19 taxa) richnesses were similar but lower than this median richness indicative of incomplete recovery from previous erosion events (Figure 30).

MCI values have had a very wide range (41 units) at this site, atypical of a site in the upper reaches of a ringplain stream mainly due to headwater erosion effects. The median value (128 units) however, has been more typical of upper reach sites elsewhere on the ringplain, with the spring, 2012 (139 units) and summer 2013 (128 units) scores significantly different and higher than or equal with the historical median score (128 units). These scores categorised this site as having 'very good' generic health (Table 1) in spring and summer and, in terms of predictive relationships (Table 2), 'better than expected' and 'expected' health for the upper reaches of a ringplain stream in spring and summer respectively. The historical median score (128 units) placed this site in the 'very good' and 'expected' categories for the generic and predictive methods of assessment respectively.

3.2.8.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 35.

Table 35 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Maketawa Stream at Derby Road between 1995 and April 2012 [25 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
EPHEMEROPTERA	*Austroclima*	7	1	4		
	Coloburiscus	7	11	44		
	Deleatidium	8	25	100	XA	XA
	Nesameletus	9	15	60	VA	VA
PLECOPTERA	*Megaleptoperla*	9	10	40		
	Zelandoperla	8	20	80	VA	VA
COLEOPTERA	Elmidae	6	22	88	VA	VA
	Hydraenidae	8	3	12		
TRICHOPTERA	*Aoteapsyche*	4	10	40		
	Costachorema	7	5	20		
	Hydrobiosis	5	1	4		
	Beraeoptera	8	11	44		
	Helicopsyche	10	8	32		
	Olinga	9	1	4		
	Pycnocentrodes	5	8	32		
DIPTERA	*Aphrophila*	5	11	44		A
	Eriopterini	5	4	16		
	Maoridiamesa	3	5	20		
	Orthocladiinae	2	7	28		

Prior to the current 2012-2013 period, 19 taxa have characterised the community at this site on occasions. These have comprised eight 'highly sensitive', eight 'moderately sensitive', and three 'tolerant' taxa i.e. a predominance of 'sensitive' taxa as would be expected in the upper reaches of a ringplain stream. Predominant taxa have included three 'highly sensitive' taxa [mayflies (*Deleatidium* on every occasion, and *Nesameletus*) and stonefly (*Zelandoperla*)]; one 'moderately sensitive' taxon [elmid beetles]; but no 'tolerant' taxa. All four of these predominant taxa were dominant in the spring, 2012 community while the summer, 2013 community was characterised by the four taxa dominant in spring, together with one additional 'moderately sensitive' taxon, which also previously had been characteristic of this site's communities (Table 35). The similarity in the seasonally most dominant taxa composition was evident in the very similar $SQMCI_s$ scores which varied by only 0.1

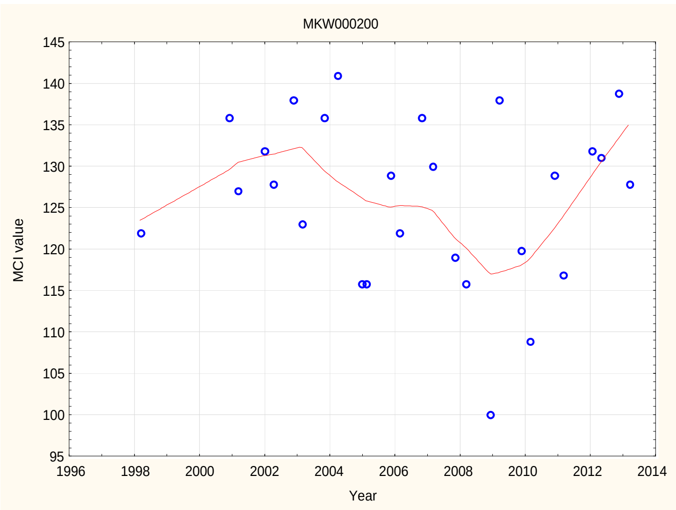
unit (Tables 140 and 141). The taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 60% to 100% of past survey occasions.

3.2.8.1.3 Predicted stream 'health'

The Maketawa Stream site at Derby Road is 2.3 km downstream of the National Park boundary at an altitude of 380 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 123 (altitude) and 121 (distance) for this site. The historical site median (128 units) is 5 units higher than the altitude prediction and 7 units above the distance predictive value. The spring, 2012 survey score (139 units) was up to a significant (Stark, 1998) 18 units higher than the predictive values while the summer, 2013 score (128 units) was higher (by 5 to 7 units) than both predictive values. Of the 27 surveys to date at this site, 26% of MCI scores have been less than 121 units while 59% have been greater than 123 units.

3.2.8.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the thirteen years of SEM results collected to date from the site in the Maketawa Stream at Derby Road. The MCI has been chosen as the preferable indicator 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 31.



N = 27
Kendall tau = -0.064
p level = 0.641 [>FDR, p = 0.709]
N/S at $p < 0.05$

Figure 31 LOWESS trend plot of MCI data at the Derby Rd. site

No significant temporal trend in the overall slight decline in MCI scores has been found over the thirteen year monitoring period at this relatively pristine site. Scores have tended to decrease gradually, particularly since the headwater erosion events, prior to recovery over the more recent four-year period. The variability in LOWESS-smoothed scores (range of 18 units) represented moderate ecological significance during the period accentuated by the impact of headwater erosion events during 2008.

Overall, smoothed scores remained indicative of 'very good' generic stream health (Table 1) for the majority of the period, dropping to 'good' health between 2008 and 2010. In terms of predictive relationships (Table 2) for a site in the upper reaches of a ringplain stream, stream health has been 'expected' almost throughout the thirteen year period, but falling toward 'worse than expected' for two years following the headwater erosion events of 2008 and very recently improving markedly toward the 'better than expected' category.

3.2.8.2 Tarata Road site (MKW000300)

3.2.8.2.1 Taxa richness and MCI

Twenty-four surveys have been undertaken at this mid-reach site at Tarata Road in the Maketawa Stream between March 1998 and April 2012. These results are summarised in Table 36, together with the results from the current period, and illustrated in Figure 32.

Table 36 Results of previous surveys performed in the Maketawa Stream at Tarata Road together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Apr 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MKW000300	24	12-31	23	90-115	103	19	116	24	117

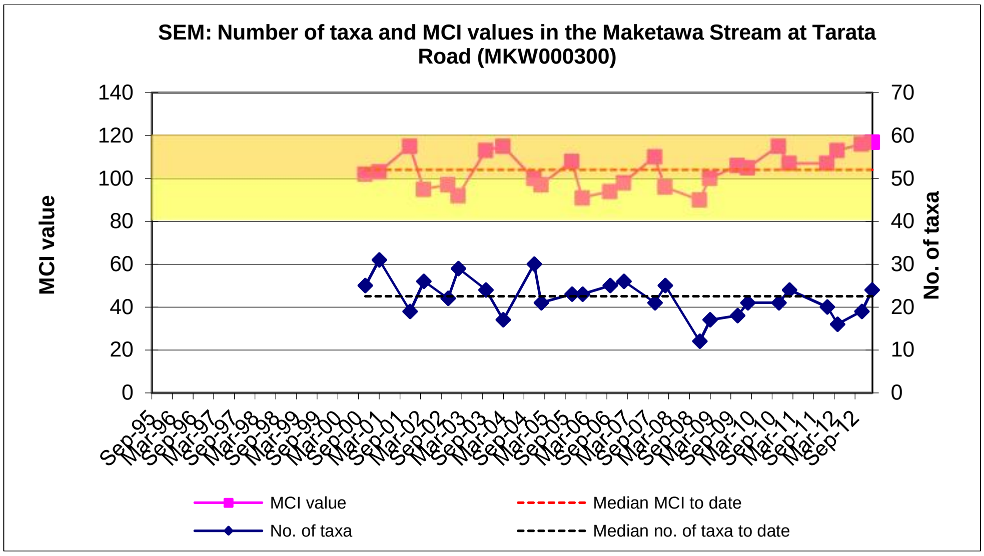


Figure 32 Number of taxa and MCI values in the Maketawa Stream at Tarata Road

A wide range of richnesses (12 to 31 taxa) has been found; wider than might be expected, with a median richness of 23 taxa which is more representative of typical richnesses in the mid-reaches of ringplain streams and rivers. During the 2012-2013 period, spring (19 taxa) and summer (24 taxa) richnesses were lower and higher respectively than the median taxa number coincident with an increase in substrate periphyton cover in summer.

MCI scores have had a relatively wide range (25 units) at this site, more typical of sites in the lower reaches of ringplain streams. The median value (103 units) has been relatively typical of mid-reach sites elsewhere on the ringplain. The spring, 2012 (116 units) and summer, 2013 (117 units) scores were very similar and within the range typical for such a site and higher than the historical median by a significant (Stark, 1998) 13 to 14 units. These scores categorized this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' health on both occasions for the mid-reaches of a ringplain stream. The historical median score (103 units) placed this site in the 'good' category for generic health and the 'expected' category for the predictive method of assessment.

3.2.8.2.2 Community composition

Characteristic macroinvertebrate taxa abundant in the communities at this site prior to the 2012-2013 period are listed in Table 37.

Table 37 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Maketawa Stream at Tarata Road between 1995 and April 2012 [24 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	7	29		
EPHEMEROPTERA	*Austroclima*	7	3	13		
	Coloburiscus	7	11	46	A	A
	Deleatidium	8	15	63	XA	XA
	Nesameletus	9	1	4		
PLECOPTERA	*Acroperla*	5	1	4		
COLEOPTERA	Elmidae	6	7	29		A
MEGALOPTERA	*Archichauliodes*	7	2	8		A
TRICHOPTERA	*Aoteapsyche*	4	14	58		VA
	Costachorema	7	11	46		
	Hydrobiosis	5	8	33		
	Neurochorema	6	3	13		
	Beraeoptera	8	2	8		
	Confluens	5	2	8		
	Oxyethira	2	4	17		
	Pycnocentrodes	5	1	4		
DIPTERA	*Aphrophila*	5	20	83		VA
	Maoridiamesa	3	16	67	A	A
	Orthocladiinae	2	23	96		A
	Tanytarsini	3	7	29		
	Empididae	3	1	4		
	Muscidae	3	4	17		
	Austrosimulium	3	2	8		

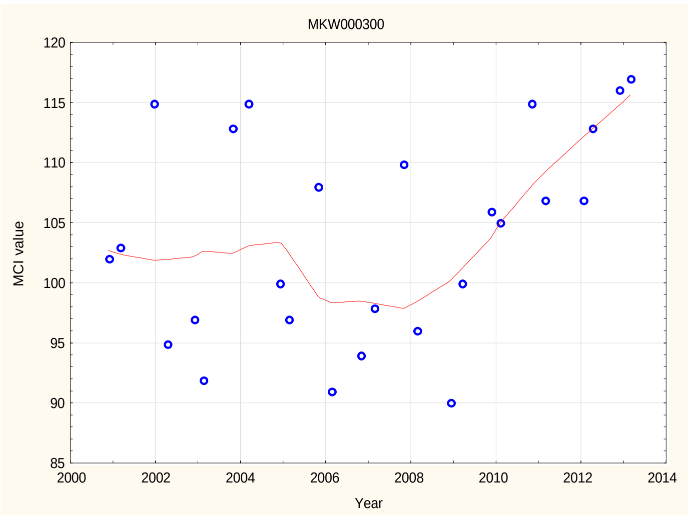
Prior to the current 2012-2013 period 23 taxa have characterised the community at this site on occasions. These have comprised three 'highly sensitive', eleven 'moderately sensitive', and nine 'tolerant' taxa i.e. a predominance of 'sensitive' taxa as might be expected in the mid-reaches of a ringplain stream. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; one 'moderately sensitive' taxon [cranefly (*Aphrophila*)]; and three 'tolerant' taxa [net-building caddisfly (*Aoteapsyche*) and midges (orthoclads and *Maoridiamesa*)]. Only three of the historically characteristic taxa were dominant in the spring 2012 community. These comprised one 'highly sensitive', one 'moderately sensitive', and one 'tolerant' taxa, whereas these three taxa together with another three 'moderately sensitive' and three 'tolerant' taxa, comprised the dominant taxa in the summer community. Therefore, only three of these eight taxa were dominant in both spring and summer communities (Table 37). These seasonal dominance differences resulted in a decrease in summer $SQMCI_s$ score of 0.9 unit (Tables 140 and 141). The three taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 58% to 83% of past survey occasions.

3.2.8.2.3 Predicted stream 'health'

The Maketawa Stream site at Tarata Road is 15.5 km downstream of the National Park boundary at an altitude of 150 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 100 (altitude) and 101 (distance) for this site. The historical site median (103 units) is only three units above the altitude prediction and two units above the predictive distance value, while the spring, 2012 survey score (116 units) and the summer, 2013 score (117 units) were significantly higher (by at least 15 units) than these predictive values. Of the 26 surveys to date at this site, 35% of MCI scores have been less than 100 units while 58% have been greater than 101 units.

3.2.8.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006), has been performed on the thirteen years of SEM results collected to date from the site, in the Maketawa Stream at Tarata Road. The MCI has been chosen as the preferable indicator 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend pot of MCI data is presented in Figure 33.



N = 26
Kendall tau = +0.261
p level = 0.061 [>FDR, p = 0.106]
N/S at p < 0.05

Figure 33 LOWESS trend plot at the Tarata Road site

The increasing trend in MCI scores found over the thirteen year monitoring period has not been statistically significant. Ecological variability in LOWESS-smoothed scores (which ranged over 18 units) has been significant ecologically with scores indicative of 'good' generic stream health (Table 1) trending downward to 'fair' stream health, between 2005 and 2009 before returning to 'good' health more recently. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a ringplain stream, stream health has been in the 'expected' category for the majority of the period, before very recently improving to 'better than expected'.

3.2.8.3 Discussion

Seasonal MCI values were significantly different between spring and summer at the upper reach (Derby Road) site and greater than the historical median decrease (7 units) recorded for this site (Appendix II). Values atypically remained very similar in summer (within 1 unit) at the mid-reach site, in comparison with the historical median 7 unit summer decrease (Appendix II). Seasonal communities at the upper reach site shared 15 common taxa (60% of the 25 taxa found at this site in 2012-2013) compared with 13 shared common taxa (43% of the 30 taxa found in 2012-2013) at the mid-reaches site (Tarata Road); dissimilar seasonal changes in community structures at the two sites and more pronounced at the site in the mid reaches. The two sites shared 11 common taxa (38% of the 29 taxa) in spring and 15 common taxa (54% of 28 taxa) in summer, indicative of the dissimilarity in spatial community structures in both spring and summer, although atypically more so in the spring.

MCI scores typically fell in a downstream direction in both spring (by 23 units) and atypically, less markedly in summer (by 11 units), over a stream distance of 15.1 km downstream from the Denby Road site. These falls in MCI scores equated to rates of decline of 1.5 units/km in spring decreasing to 0.7 units/km in summer, compared with a predicted rate of 1.5 units/km over the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009).

Using the long-term median SEM MCI scores for both sites (Appendix II), the rate of overall decline has been 1.6 MCI units/km over the river's length (i.e. similar to the predicted rate). Therefore rates of decline over the 2012–2013 period were similar in spring, but much lower in summer than the long term average rate to date.

3.2.9 Waitara River

The results found by the 2012-2013 surveys are presented in Table 142 and Table 143, Appendix I.

3.2.9.1 Mamaku Road site (WTR000850)

3.2.9.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this lower reach site in the Waitara River between November 1995 and April 2012. These results are summarised in Table 38, together with the results from the current period, and illustrated in Figure 34.

Table 38 Results of previous surveys performed in the Waitara River at Mamaku Road together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to April 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WTR000850	34	9-32	20	64-101	86	18	107	14	83

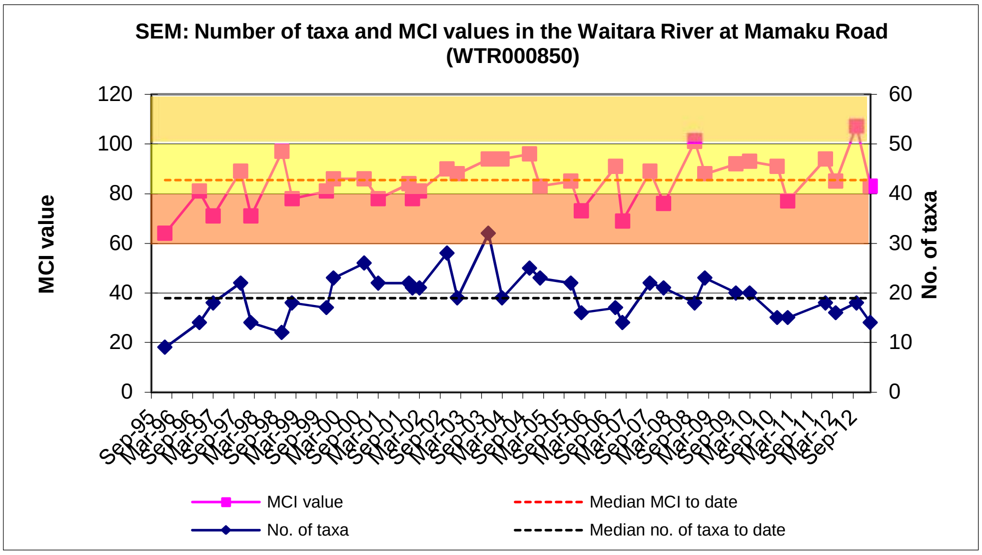


Figure 34 Numbers of taxa and MCI values in the Waitara River upstream of Methanex at Mamaku Road

A very wide range of richnesses (9 to 32 taxa) has been found with a moderate median richness of 20 taxa (more representative of typical richnesses in the lower reaches of streams and rivers). During the 2012-2013 period, spring and summer richnesses (18 and 14 taxa respectively) were lower than this median richness.

MCI values have had a wide range (37 units) at this site which were typical for a site in the lower reaches of large rivers. The historical median value (86 units) has also been typical of lower reach sites elsewhere. The spring, 2012 (107 units) score was a significant (Stark, 1998) 21 units higher than the historical median and also 6 units above the maximum to date despite widespread periphyton mats and patchy filamentous algal substrate cover. The summer, 2013 (83 units) score was 3 units lower than the historical median when periphyton mats were thicker. These scores categorised this site as having 'good' health generically (Table 1) in spring and 'fair' health in summer and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower reaches of a river with some ringplain catchment component. The historical median score (86 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.9.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 39.

Table 39 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waitara River at Mamaku Road between 1995 and April 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	2	6		
ANNELIDA	Oligochaeta	1	23	68		VA
	Branchiura	1	1	3		
	Polychaeta	3	2	6		
MOLLUSCA	*Latia*	5	10	29		
	Potamopyrgus	4	17	50		
CRUSTACEA	Tanaidacea	3	1	3		
	Paratya	3	13	38		
EPHEMEROPTERA	*Deleatidium*	8	11	32	VA	
COLEOPTERA	Elmidae	6	1	3		
TRICHOPTERA	*Aoteapsyche*	4	20	59		VA
	Oxyethira	2	10	29		
	Pycnocentrodes	5	4	12		
DIPTERA	*Aphrophila*	5	12	35	A	A
	Maoridiamesa	3	3	9		
	Orthocladiinae	2	24	71	A	VA
	Tanytarsini	3	9	26		A
	Austrosimulium	3	1	3		

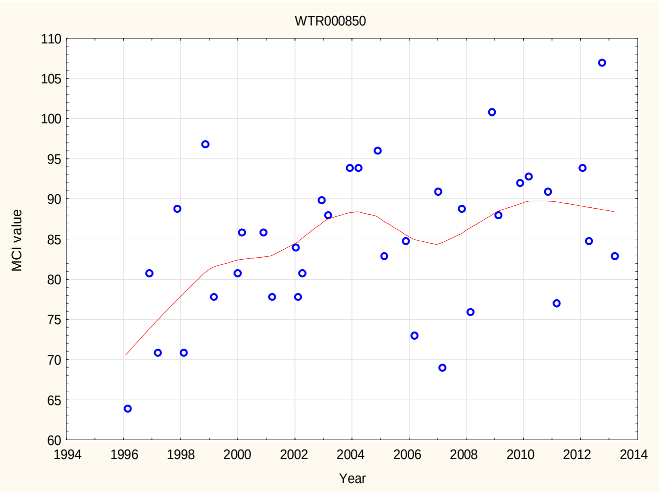
Prior to the current 2012-2013 period, 18 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', four 'moderately sensitive', and thirteen 'tolerant' taxa i.e. a high proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain river. Predominant taxa have included only four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and orthoclad midges]. Only one of these predominant taxa were dominant in the spring, 2012 community together with two of the other historically characteristic 'sensitive' taxa. The summer, 2013 community was characterised by three additional taxa (all 'tolerant') and one fewer ('highly sensitive') taxon than those dominant in spring (Table 39). As a result of these seasonal differences in characteristic taxa, particularly the decrease in numbers of the 'highly sensitive' mayfly taxon and increase in abundances within three 'tolerant' taxa in the summer survey, there was a significant decrease in $SQMCI_s$ scores of 3.7 units(Tables 142 and 143) in summer. The taxa recorded as very abundant during spring/summer had characterised this site's communities on 32% to 71% of past survey occasions.

3.2.9.2.1 Predicted stream 'health'

The Waitara River site at Mamaku Road, at an altitude of 15 m asl, is in the lower reaches of a river draining a catchment comprised of both eastern hill country and ringplain sub-catchments. A relationship for ringplain streams and rivers developed between MCI and altitude (Stark and Fowles, 2009) predicts a MCI value of 86 units for this site. The historical site median (86 units) was equivalent with this altitude prediction while the spring, 2012 (107 units) score was higher than this predictive value by a significant 21 units and the summer, 2013 score (83 units) was lower by three units than the predictive value. These two surveys' scores were also 31 and 7 units higher than the median MCI (76 units) found from 216 previous surveys of 'control' sites below 25 m asl in hill country catchment streams and rivers (TRC, 1999 (updated, 2012)). Of the 36 surveys to date at this river site, 14% of MCI scores have been less than 75 units while 47% have been greater than 86 units.

3.2.9.3 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years (1995-2011) of SEM results collected to date from the site in the Waitara River at Mamaku Road. The MCI has been chosen as the preferable indicator 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 35.



N = 36
Kendall tau = +0.273
p level = 0.019 [>FDR, p = 0.041]
Significant at p< 0.05; and
after FDR application

Figure 35 LOWESS trend plot of MCI data for the Mamaku Road site

Despite a statistically significant improvement in MCI scores over the first ten years of monitoring ($p < 0.05$ after FDR), and more recent higher scores, have resulted in an overall positive trend for the eighteen year period which has been as significant overall statistically ($p < 0.05$ after FDR). This may be consistent with the earlier assessment that linked improvement with climatic factors in this large, predominantly eastern hill country catchment. The range of LOWESS-smoothed scores (19 units) has been ecologically significant over the period. These MCI scores have been indicative of a general improvement from 'poor' to 'fair' generic river health (Table 1). In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain river (recognising the partial ringplain component of this catchment), river health has been within the 'expected' category almost throughout the eighteen year period.

3.2.9.4 Discussion

Seasonal MCI values typically decreased between spring and summer (but by a large 24 units) at this lower reach site with the percentage community compositions of 'tolerant' taxa increasing by 29% in summer. This decrease was much higher than the historical median (13 units) for this site (Appendix II). Seasonal communities at this site shared only 8 common taxa (33% of the 24 taxa found at this site in 2012-2013), a low percentage of common taxa as reflected in the difference in MCI values between the seasonal surveys. An increase in abundance of several 'tolerant' taxa and decrease in one 'highly sensitive' taxon abundances in particular, accounted for the decrease in $SQMCI_s$ value (3.7 units) in summer, despite patchy to widespread periphyton substrate cover on both occasions, although mats were thicker under summer low flow conditions.

3.2.10 Mangati Stream

The results found by the 2012-2013 surveys are presented in Table 144 and Table 145, Appendix I.

3.2.10.1 Site downstream of railbrige (MGT000488)

3.2.10.1.1 Taxa richness and MCI

Thirty-five surveys have been undertaken at this site in the mid reaches of this small coastal stream draining an industrial catchment between October 1995 and March 2012. These are summarised in Table 40, together with the results from the current period, and illustrated in Figure 36.

Table 40 Results of previous surveys performed in the Mangati Stream at the site downstream of the railbridge, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MGT000488	35	9-29	16	56-85	78	17	86	18	91

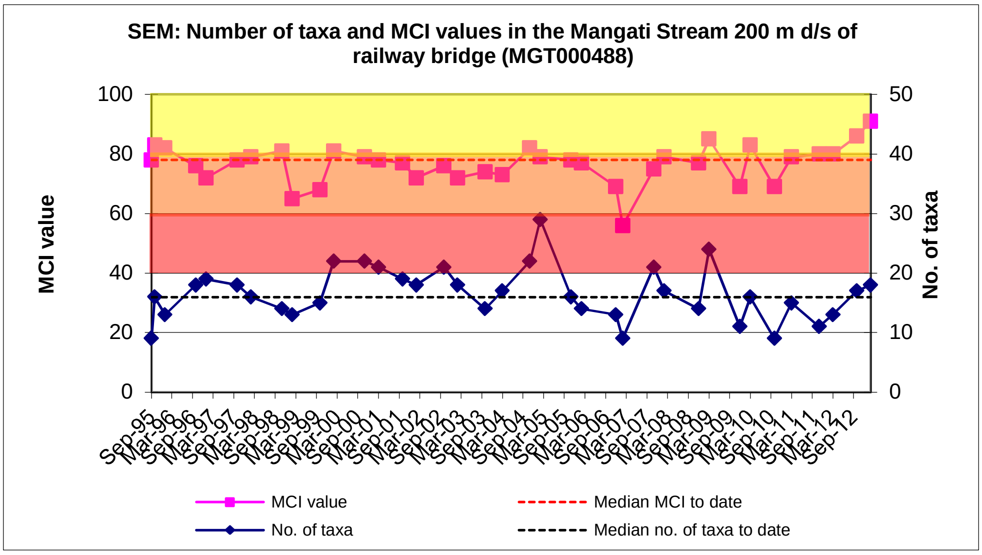


Figure 36 Numbers of taxa and MCI values in the Mangati Stream downstream of the railbridge

A very wide range of richnesses (9 to 29 taxa) has been found; with a median richness of 16 taxa (more representative of typical richnesses in upper, swampy reaches of small coastal streams (TRC, 1999 (updated 2012)). During the 2012-2013 period, spring (17 taxa) and summer (18 taxa) richnesses were very similar and slightly higher than this median richness.

MCI values have had a relatively wide range (29 units) at this site, relatively typical of a site in a small coastal stream. The median value (78 units) also has been typical of such streams elsewhere on the ringplain, and the spring, 2012 (86 units) and summer, 2013 (91 units) scores, were both well above the historical median with the summer score 6 units higher than the historical maximum. These scores were also significantly 19 and 24 units higher than the median score previously recorded by 39 surveys at 'control' sites in lowland coastal streams at altitudes between 25 m and 50 m asl (TRC, 1999 (updated, 2012)). These scores categorised this site as having 'fair' health generically (Table 1) in spring and summer. The historical median score (78 units) placed this site in the 'poor' category for the generic method of assessment.

3.2.10.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 41.

Table 41 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Mangati Stream at the site downstream of the railbridge between 1995 and March 2012 [35 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
PLATYHELMINTHES	*Cura*	3	5	14		
NEMERTEA	Nemertea	3	3	9		
ANNELIDA	Oligochaeta	1	31	89		A
	Lumbricidae	5	1	3		
MOLLUSCA	*Physa*	3	4	11		
	Potamopyrgus	4	33	94	XA	XA
	Sphaeriidae	3	1	3		
CRUSTACEA	Ostracoda	1	9	26		
	Paracalliope	5	29	83	VA	XA
	Phreatogammarus	5	1	3		
EPHEMEROPTERA	*Austroclima*	7	12	34		
	Zephlebia group	7	1	3		
HEMIPTERA	*Microvelia*	3	1	3		
TRICHOPTERA	*Hydrobiosis*	5	1	3		
	Polyplectropus	6	1	3		
	Oxyethira	2	3	9		
DIPTERA	Orthocladiinae	2	17	49	A	
	Polypedilum	3	1	3		A
	Austrosimulium	3	21	60	VA	

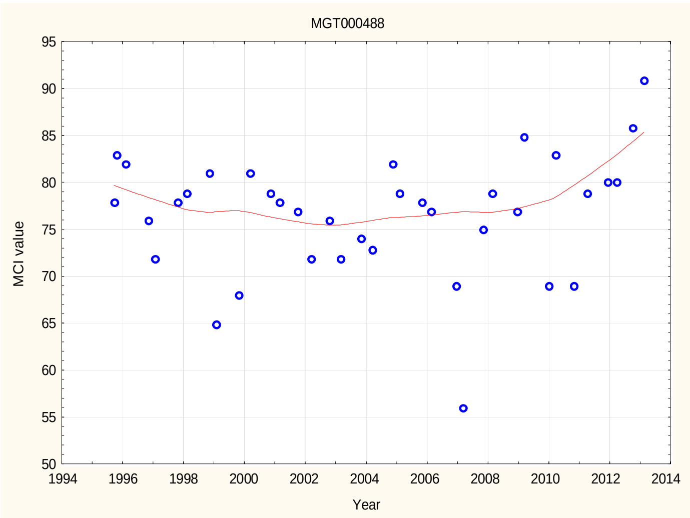
Prior to the current 2012-2013 period, 19 taxa have characterised the community at this site on occasions. These have comprised seven 'moderately sensitive' and twelve 'tolerant' taxa i.e. a higher proportion of 'tolerant' taxa as would be expected in the swampier upper reach of a softer bottom, macrophyte dominated, small coastal stream. Predominant taxa have included only one 'moderately sensitive' taxon (amphipod (*Paracalliope*) and three 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), and sandfly (*Austrosimulium*)]. Only four of the characteristic taxa were dominant in the spring, 2012 community. The summer, 2013 community was characterised by two of the taxa dominant in spring, together with two additional 'tolerant' taxa, both of which had been characteristic of this site's communities in the past (Table 41). The decrease in abundance of 'tolerant' sandflies and increase in abundance of 'moderately sensitive' amphipods in summer was reflected in the increase (of 0.4 unit) between spring and summer SQMCI$_s$ scores (Tables 144 and 145). The three taxa recorded as extremely or very abundant during spring and/or summer had characterised this site's communities on 60% to 94% of past surveys.

3.2.10.1.3 Predicted stream 'health'

The Mangati Stream site downstream of the railbridge is in the mid, swampier reaches of a small coastal stream at an altitude of 30 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) however, are not applicable in this type of small coastal stream.

3.2.10.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Mangati Stream at the site downstream of the railbridge. The MCI has been chosen as the preferable indicator 'stream/river health/ for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 37.



N = 37
Kendall tau = + 0.095
p level = 0.407 [>FDR, p = 0.524]
N/S at $p < 0.05$

Figure 37 LOWESS trend plot of MCI data at the site downstream of the railbridge

There was a slight negative overall trend identified in the MCI scores, which had no statistical significance, with a more pronounced recent improvement since 2010. This site's scores have had a LOWESS-smoothed range of 9 units indicative of no ecological significance over the period.

Overall, smoothed scores remained indicative of 'poor' generic stream health (Table 1) throughout most of the period improving to 'fair' generic health in the most recent two years. It also must be recognised that trends in the health of this 'soft-bottomed' lowland stream might be assessed more appropriately in future by application of the more recently established soft-bottomed MCI (SBMCI).

3.2.10.2 Te Rima Place, Bell Block site (MGT000520)

3.2.10.2.1 Taxa richness and MCI

Thirty-five surveys have been undertaken at this lower reach site at SH45 in the Mangati Stream between October 1995 and March 2012. These results are summarised in Table 42, together with the results from the current period, and illustrated in Figure 38.

Table 42 Results of previous surveys performed in the Mangati Stream at Te Rima Place, Bell Block together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MGT000520	35	3-22	10	44-78	64	12	75	17	75

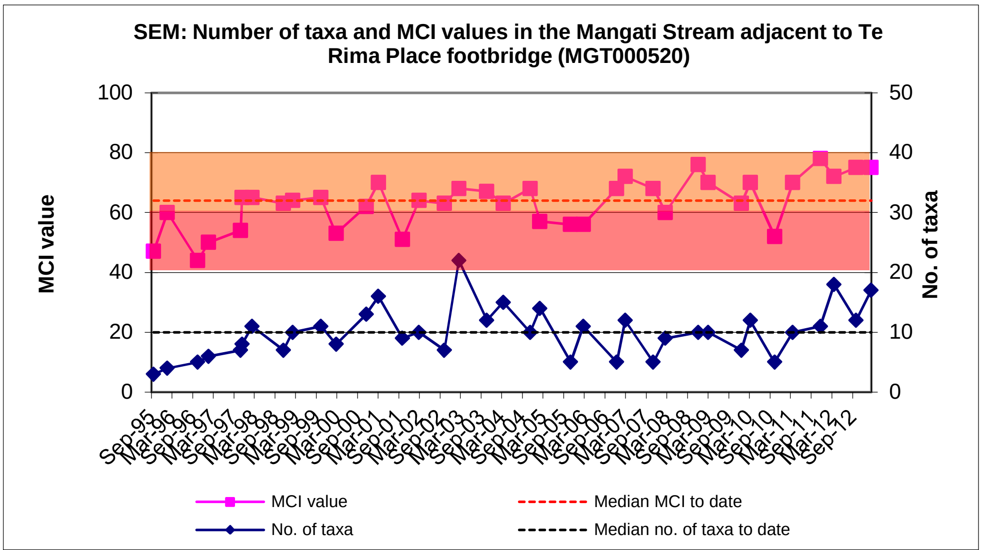


Figure 38 Numbers of taxa and MCI values in the Mangati Stream at Te Rima Place footbridge

A wide range of richnesses (3 to 22 taxa) has been found; wider than might be expected with a median richness of 10 taxa but not atypical of richnesses in the lower reaches of small coastal streams. During the 2012-2013 period, spring (12 taxa) and summer (17 taxa) richnesses were dissimilar and from two to seven taxa above the median taxa number.

MCI scores have had a relatively wide range (34 units) at this site, more typical of sites in the lower reaches of small coastal streams. The median value (64 units) also has been relatively typical of lower reach sites in coastal streams with the spring, 2012 (75 units) score only three units below the historical maximum for this site and the summer, 2013 (75 units) score within the range typical for such a site. These scores were well above the historical median. The scores were also equal with (spring and summer), the median score found by 225 surveys at 'control sites' in similar lowland coastal streams at altitudes of less than 25 m asl (TRC, 1999 (updated, 2012)). These scores categorised this site as having 'poor' (spring and summer) health generically (Table 1). The historical median score (64 units) placed this site in the 'poor' category for the generic method of assessment.

3.2.10.2.2 Community composition

Characteristic macroinvertebrate taxa abundant in the communities at this site prior to the 2012-2013 period are listed in Table 43.

Table 43 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Mangati Stream at Te Rima Place, Bell Block between 1995 and March 2012 [35 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	2	6		
ANNELIDA	Oligochaeta	1	35	100	VA	VA
MOLLUSCA	*Potamopyrgus*	4	17	49	XA	XA
CRUSTACEA	Ostracoda	1	1	3		
TRICHOPTERA	*Oxyethira*	2	1	3		
	Triplectides	5	3	9		A
DIPTERA	Orthocladiinae	2	26	74	A	
	Empididae	3	2	6		
	Austrosimulium	3	5	14		

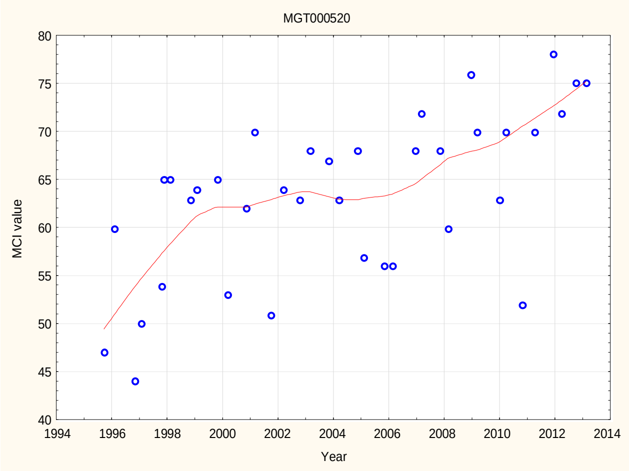
Prior to the current 2012-2013 period a small number of taxa (9) has characterised the community at this site on occasions. These have comprised one 'moderately sensitive' and eight 'tolerant' taxa i.e. a majority of 'tolerant' taxa as would be expected in the lower reaches of a small coastal ringplain stream. Predominant taxa have included no 'moderately sensitive' taxa but two 'tolerant' taxa [oligochaete worms and orthoclad midges]. Only three of the historically characteristic taxa were dominant in the spring, 2012 community. These were all 'tolerant' taxa whereas two of these and one 'moderately sensitive' taxa comprised the dominant taxa in the summer community and the significant abundances of the shared taxa on both occasions were reflected in the identical seasonal $SQMCI_s$ scores (Tables 144 and 145). Two of these four taxa were dominant in both spring and summer communities (Table 43). The two taxa recorded as very or extremely abundant during spring and summer had characterised this site's communities on 49% to 100% of past surveys.

3.2.10.2.3 Predicted stream 'health'

The Mangati Stream at Te Rima Place, Bell Block is in the lower, more gravel-bottomed reaches of a small coastal stream at an altitude of 20 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) however, are not applicable in this type of small coastal stream.

3.2.10.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site, in the Mangati Stream at the Te Rima Place, Bell Block site. The MCI has been chosen as the preferable indicator 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 39.



N = 37
Kendall tau = + 0.466
p level = < 0.0001 [>FDR, p < 0.001]
Significant at p < 0.05 and p < 0.01
after FDR application

Figure 39 LOWESS trend plot at the Te Rima Place, Bell Block

A positive temporal trend in MCI scores, statistically significant ($p < 0.01$) prior to and after FDR analysis, indicated continued improvement coincident with better control and treatment of industrial point source discharges in the upper and mid-catchment and wetland installation (stormwater interception) in mid catchment with this improvement continuing in recent years. The LOWESS-smoothed range of scores (24 units) was ecologically significant with MCI scores indicative of a shift from 'very poor' to 'poor' generic stream health (Table 1) during the period.

This trend of improvement in stream 'health' at this site has been much more pronounced than the trend at the site 1.5 km upstream, indicating that activities in the catchment between these two sites have had a significant influence.

3.2.10.3 Discussion

Seasonal MCI values increased (by 5 units) between spring and summer at the upstream site, where seasonal median values have been within 2 units (Appendix II), but remained identical at the lower site with minimal change in the percentage composition of 'tolerant' taxa in the summer community. Seasonal communities at the upper reach site shared 12 common taxa (52% of the 23 taxa found at this site in 2012-2013) compared with eight shared common taxa (38% of the 21 taxa) at the lower reaches site, a more pronounced seasonal change in community structure at the lower reach site. MCI values did not change in summer at this site where long-term median scores have shown a very small 2 unit summer increase to date (Appendix II). The two sites shared ten taxa (53% of the 19 taxa) in spring and eleven taxa (46% of 24 taxa) in summer, indicative of the dissimilarity in spatial community structures in both spring and more so in summer.

MCI score typically fell in a downstream direction in both spring (by 16 units) and less markedly in summer (by 11 units), over a stream distance of 1.4 km equating to relatively wide ranging rates of decline typical of a small coastal developed catchment stream.

Using the long-term median SEM MCI scores for both sites (Appendix II), the overall rate of decline has been 10 MCI units/km over the surveyed length of the stream. Therefore rates of decline over the 2012-2013 period were slightly higher in spring and lower in summer than this historical average rate.

3.2.11 Waimoku Stream

The results found by the 2012-2013 surveys are presented in Table 146 and Table 147, Appendix I.

3.2.11.1 Lucy's Gully site (WMK000100)

3.2.11.1.1 Taxa richness and MCI

Twenty-four surveys have been undertaken at this upper reach site in the Waimoku Stream (in the Kaitake Ranges) between December 1999 and March 2012. These results are summarised in Table 44, together with the results from the current period, and illustrated in Figure 40.

Table 44 Results of previous surveys performed in the Waimoku Stream at Lucy's Gully, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WMK000100	26	22-37	30	121-141	132	38	129	38	124

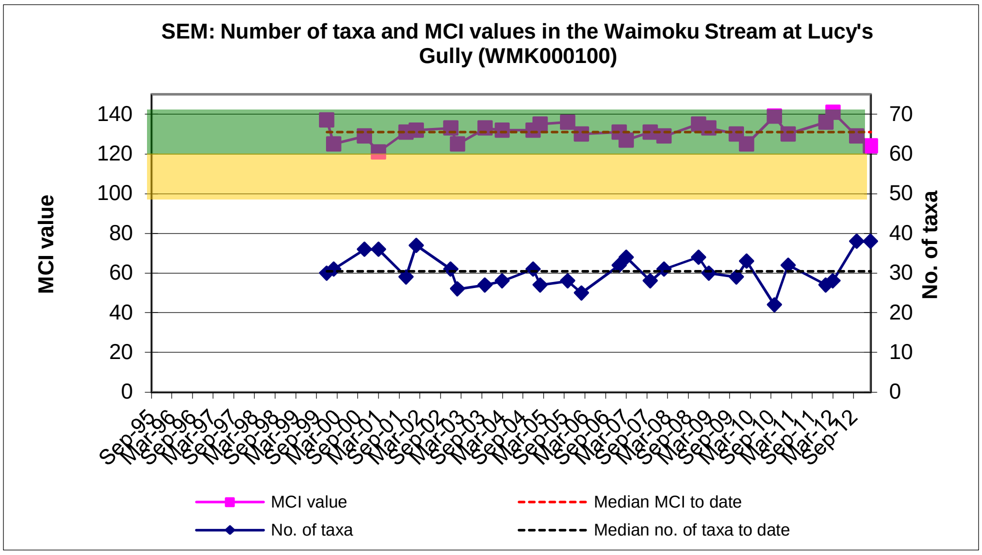


Figure 40 Numbers of taxa and MCI values in the Waimoku Stream at Lucy's Gully

A moderate range of richnesses (22 to 37 taxa) has been found, with a median richness of 30 taxa which is more representative of typical richnesses in the upper reaches of ringplain streams and rivers. During the 2012-2013 period, spring (38 taxa) and summer (38 taxa) richnesses were identical, well above this median richness, and one taxon more than found on any previous occasion.

MCI values also have had a moderate range (20 units) at this site, slightly wider than typical of a site in the upper reaches of a ringplain stream. The median value (132 units) however, has been typical of upper reach sites elsewhere on the ringplain, and the spring, 2012 (129 units) and summer, 2013 (124 units) scores were three units and eight units below the historical median respectively. These scores categorised this site as having 'very good' health generically (Table 1) in spring and in summer and, in terms of predictive relationships (Table 2), 'expected' health for the upper reaches of a ringplain stream on spring and summer occasions. The historical median score

(132 units) placed this site in the 'very good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.11.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 45.

Table 45 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waimoku Stream at Lucy's Gully between 1999 and March 2012 [26 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	3	12		
MOLLUSCA	*Potamopyrgus*	4	5	19		
EPHEMEROPTERA	*Austroclima*	7	20	77		
	Coloburiscus	7	26	100	VA	VA
	Deleatidium	8	24	92	A	
	Ichthybotus	8	1	4		
	Zephlebia group	7	22	85	VA	VA
PLECOPTERA	*Austroperla*	9	18	69	A	A
	Stenoperla	10	2	8		
	Zelandobius	5	1	4		
COLEOPTERA	Elmidae	6	1	4		A
	Ptilodactylidae	8	5	19		
MEGALOPTERA	*Archichauliodes*	7	3	12		
TRICHOPTERA	*Hydrobiosella*	9	7	27		
	Orthopsyche	9	26	100	VA	VA
DIPTERA	Orthocladiinae	2	19	73		
	Polypedilum	3	6	23		

Prior to the current 2012-2013 period, 17 taxa have characterised the community at this site on occasions. These have comprised seven 'highly sensitive', six 'moderately sensitive', and four 'tolerant' taxa i.e. a very high proportion of 'sensitive' taxa as would be expected in the upper reaches of a ringplain stream within the National Park's Kaitaki Ranges. Predominant taxa have included three 'highly sensitive' taxa [mayfly (*Deleatidium*), stonefly (*Austroperla*), and free-living caddisfly (*Orthopsyche*)]; three 'moderately sensitive' taxa [mayflies (*Austroclima, Coloburiscus,* and *Zephlebia* group)]; and one 'tolerant' taxon [orthoclad midges]. Five of these predominant taxa were characteristic of the spring, 2012 community. The summer, 2013 community was characterised by all but one of the taxa dominant in spring and one other ('moderately sensitive') taxon which previously had only once been characteristic of this site's communities (Table 45). Taxa recorded as very abundant during spring and/or summer had characterised this site's communities on 85% to 100% of past surveys.

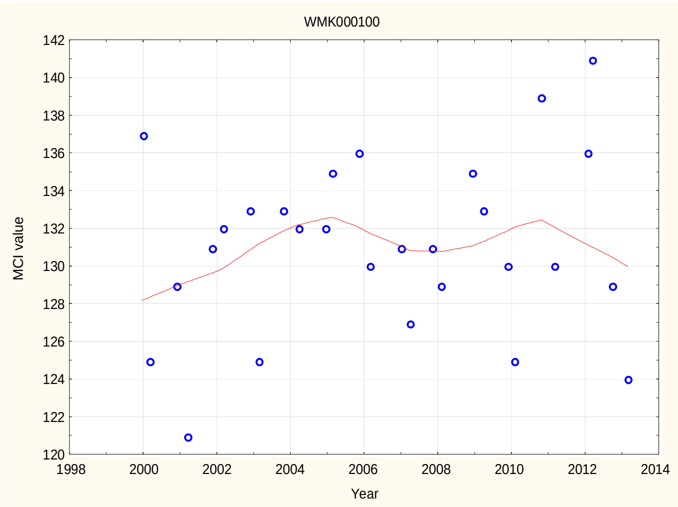
3.2.11.1.3 Predicted stream 'health'

The Waimoku Stream site at Lucy's Gully is within the Kaitaka Ranges of the National Park boundary at an altitude of 160 m asl.

Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 101 (altitude) and 132 (distance) for this site. The historical site median (132 units) is significantly 31 units higher than the altitude prediction and equal with the distance predictive value. The spring, 2012 survey score (129 units) was 3 units below the distance predictive value while the summer score (124 units) was 8 units below the distance predictive value. Of the 28 surveys to date at this site, no MCI scores have been less than 101 units while 36% have been greater than 132 units.

3.2.11.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the fourteen years of SEM results collected to date from the site in the Waimoku Stream at Lucy's Gully. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 41.



N = 28
Kendall tau = +0.054
p level = 0.685 [>FDR, p = 0.712]
N/S at p < 0.05 level

Figure 41 LOWESS trend plot of MCI data at the Lucy's Gully site

No significant temporal trend in MCI scores has been found over the fourteen year period at this pristine site within the National Park although minor improvement has been apparent at times. The LOWESS-smoothed range of scores (5 units) has not been ecologically significant and these MCI scores have continuously indicated 'very good' generic stream health (Table 1). In terms of predictive relationships (Table 2) for a site in the upper reaches of a ringplain stream, smoothed LOWESS MCI scores have indicated stream health as 'expected' through the fourteen year period.

3.2.11.2 Oakura Beach site (WMK000298)

3.2.11.2.1 Taxa richness and MCI

Twenty-six surveys have been undertaken at this lower reach site at Oakura Beach in the Waimoku Stream between December 1999 and March 2012. These results are summarised in Table 46, together with the results from the current period, and illustrated in Figure 42.

Table 46 Results of previous surveys performed in the Waimoku Stream at Oakura Beach together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WMK000298	26	10-26	20	75-101	89	13	95	27	87

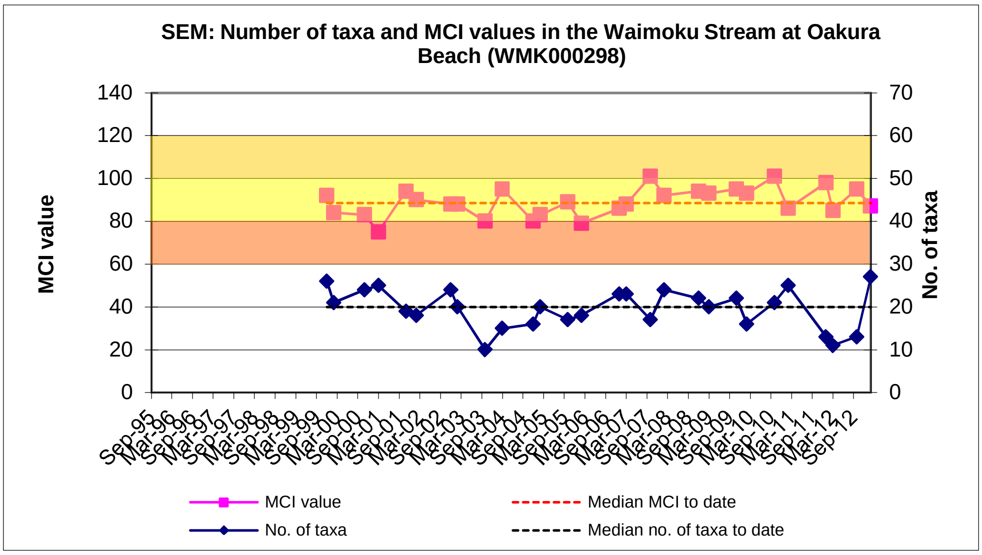


Figure 42 Numbers of taxa and MCI values in the Waimoku Stream at Oakura Beach

A wide range of richness (10 to 26 taxa) has been found; wider than might be expected, with a median richness of 20 taxa which was more representative of typical richnesses in ringplain streams and rivers in the lower reaches. During the 2012-2013 period, spring (13 taxa) and summer (27 taxa) richnesses were very different and much lower (spring) and much higher (summer) than the median taxa number coincident with minimal periphyton substrate cover in spring and widespread cover in summer. The summer richness was one taxon higher than the previous maximum for this site.

MCI scores have had a relatively wide range (26 units) at this site, typical of sites in the lower reaches of ringplain streams. The historical median value (89 units) has been relatively typical of lower reach sites elsewhere on the ringplain. The spring, 2012 (95 units) and summer, 2013 (87 units) scores were within the range typical for such a site and higher than the historical median by 6 units in spring and 2 units lower than the historical median in summer. These scores categorised this site as having ‘fair’ (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), ‘expected’ (spring and summer) health for the lower reaches of a ringplain stream. The historical median score (89 units) placed this site in the 'fair' category for generic and 'expected' category for predictive methods of assessment.

3.2.11.2.2 Commnity composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 47.

Table 47 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waimoku Stream at Oakura Beach between 1999 and March 2012 [26 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	12		
ANNELIDA	Oligochaeta	1	20	77		A
MOLLUSCA	*Potamopyrgus*	4	17	65		XA
	Sphaeriidae	3	1	4		
CRUSTACEA	Ostracoda	1	1	4		
	Paratya	3	1	4		
EPHEMEROPTERA	*Austroclima*	7	5	19		
	Coloburiscus	7	3	12	A	
	Deleatidium	8	1	4		
	Zephlebia group	7	1	4	A	
TRICHOPTERA	*Hydrobiosis*	5	4	15		A
	Oxyethira	2	3	12		
	Triplectides	5	4	15		
DIPTERA	*Aphrophila*	5	6	23		A
	Maoridiamesa	3	1	4		
	Orthocladiinae	2	25	96		A
	Polypedilum	3	4	15		
	Empididae	3	1	4		A
	Austrosimulium	3	12	46		

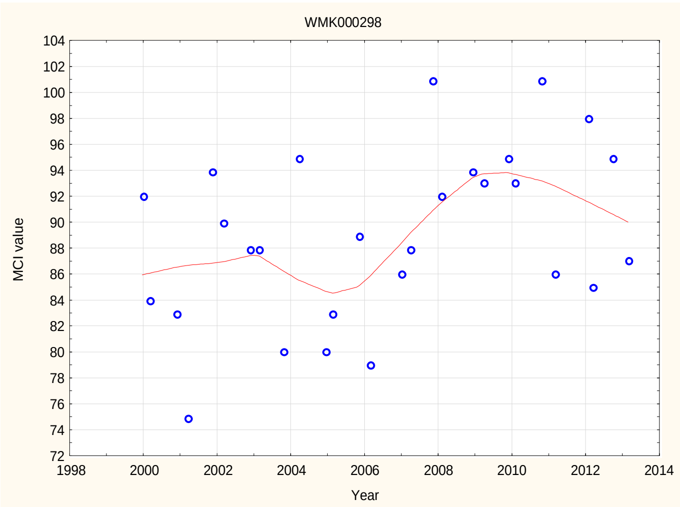
Prior to the current 2012-2013 period 19 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', six 'moderately sensitive', and twelve 'tolerant' taxa i.e. a majority of 'tolerant' taxa as would be expected in the lower reaches of a ringplain stream. Predominant taxa have included no 'highly' or 'moderately sensitive' taxa, but three 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), and orthoclad midges]. Only two of the historically characteristic taxa were dominant in the spring 2012, community. These were both 'moderately sensitive' taxa, while two other 'moderately sensitive' and four 'tolerant' taxa comprised the dominant taxa in the summer community. Therefore no taxon was dominant in both spring and summer communities (Table 47) but the increased summer numerical abundance of 'tolerant' snails in particular and three other 'tolerant' taxa were responsible for the decrease of 1.9 $SQMCI_s$ units between seasons. The one taxon recorded as extremely abundant during spring and/or summer had characterised this site's communities on 65% of past surveys.

3.2.11.2.3 Predicted stream 'health'

The Waimoku Stream at Oakura Beach site at an altitude of 1 m asl is only 4 km downstream of the National Park boundary. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 85 (altitude) and 116 (distance) for this site. The historical site median (89 units) is slightly higher (by 4 units) than the altitude prediction but 27 units lower than the predictive distance value, due to the atypically short distance between the National Park boundary and the coast for a ringplain stream. The spring, 2012 survey score (95 units) was higher than the predictive altitude value by 10 units while the summer 2013 score (87 units) was two units above the predictive altitude value. Of the 28 surveys to date at this site, 25% of MCI scores have been less than 85 units while no scores have been greater than 116 units nor within a significant 15 units of this score.

3.2.11.2.4 Temporal trends in 1995 to 2013 data

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the fourteen years of SEM results collected to date from the site in the Waimoku Stream at Oakura Beach. The MCI has been chosen as the preferable indicator 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 43.



N = 28
Kendall tau = +0.261
p level = 0.051 [>FDR, p = 0.092]
N/S at $p < 0.05$ level

Figure 43 LOWESS trend plot of MCI data at the Oakura Beach site

An overall positive temporal trend in MCI scores has been recorded during the fourteen year monitoring period but this trend has not been statistically significant. The range of LOWESS-smoothed scores (9 units) has not been of ecological significance over this period.

These smoothed MCI scores have consistently indicated 'fair' generic stream health (Table 1) at this site in the lower reaches of the stream. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, the stream health improved from 'expected' toward 'better than expected' in the 2008 to 2010 period.

3.2.11.3 Discussion

Seasonal MCI values were insignificantly lower (by 5 units) in summer compared with the historical median summer decrease of 3 units at the upper reach Lucy's Gully site. A more typical, seasonal decrease (of 8 units) was found at the lower reach site which was above the historical seasonal median decrease of 5 units (Appendix II), and the percentage composition of 'tolerant' taxa increased by 10% in the summer community. Seasonal communities at the upper reach site shared 31 common taxa (69% of the 45 taxa) compared with 11 shared common taxa (38% of the 29 taxa) at the lower reach site (Oakura Beach); a much more pronounced seasonal change in community structure at the lower reach site. The two sites shared 11 common taxa (27% of the 41 taxa) in spring and 15 common taxa (30% of 50 taxa) in summer, indicative of the dissimilarity in spatial community structures in spring and to a similar extent in summer.

MCI score typically fell in a downstream direction in both spring (by 34 units) and slightly more in summer (by 37 units), over a stream distance of 4.0 km downstream from the National Park boundary. These falls in MCI scores equated to rates of decline of 8.5 units/km in spring increasing to 9.3 units/km in summer. These were far higher than the predicted rate of 2.0 units/km over the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009), due to the short distance between the Kaitake Ranges area of the National Park and the western ringplain coastline.

Using the long-term median SEM MCI scores for both sites (Appendix II), the rate of overall decline has been 10.5 MCI units/km over the surveyed length. Therefore rates of decline over the 2012–2013 period were slightly lower in spring and in summer than the average rate to date.

3.2.12 Waiau Stream

The results found by the 2012-2013 surveys are presented in Table 148 and Table 149, Appendix I for this small lowland stream.

3.2.12.1 Inland North site (WAI000110)

3.2.12.1.1 Taxa richness and MCI

Twenty-five surveys have been undertaken in this mid-reach site in the Waiau Stream between February 1998 and March 2012. These results are summarised in Table 48, together with the results from the current period, and illustrated in Figure 44.

Table 48 Results of previous surveys performed in Waiau Stream at Inland North Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1998 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WAI000110	27	17-30	21	80-99	90	20	100	30	90

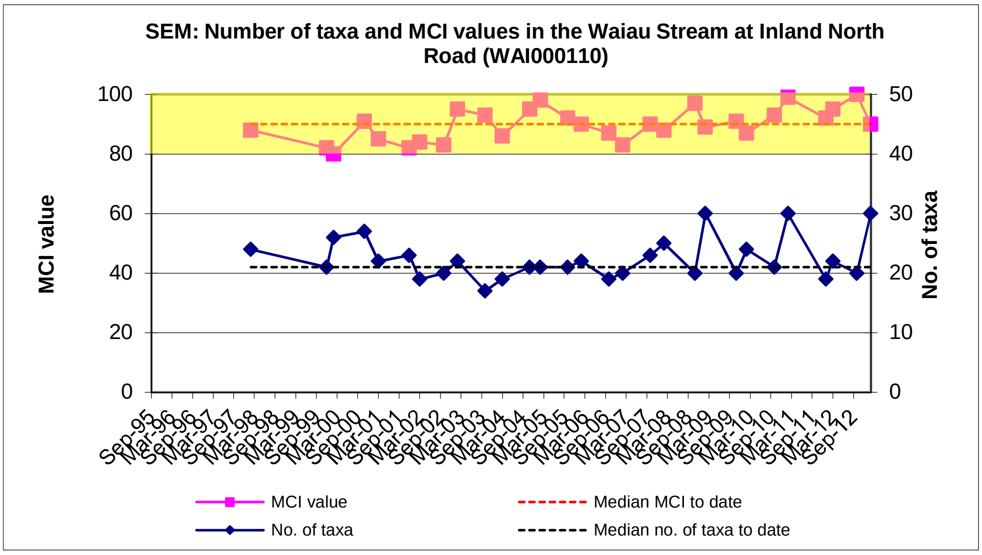


Figure 44 Numbers of taxa and MCI values in the Waiau Stream at the Inland North Road site

A moderate range of richnesses (17 to 30 taxa) has been found, with a median richness of 21 taxa (more representative of typical richnesses in small lowland streams not on the ringplain where a median richness of 20 taxa has been recorded from 92 previous surveys of 'control' sites at similar altitudes (TRC, 1999 (updated, 2012)). During the 2011-2012 period, spring (20 taxa) and summer (30 taxa) richnesses were relatively different and were close to this median richness in spring and well above in summer coincident with more widespread periphyton substrate cover.

MCI values have had a moderate range (19 units) at this site. The median value (90 units) is typical of lower reach sites in ringplain streams and rivers however, and the spring, 2012 (100 units) score was higher than typical for this site when the score was one unit higher than the previous maximum. These scores varied from 10 units higher than the historical median in spring and equivalent with this median in summer and categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1). They were significantly higher (Stark, 1998) than the median MCI score (78 units) recorded by 92 previous surveys of 'control' sites between 50 and 79 m asl in small, lowland streams in Taranaki (TRC, 1999 (updated, 2012)). The historical median score (90 units) placed this site in the 'fair' category for the generic method of assessment and was also significantly higher than the median score recorded at similar sites elsewhere.

3.2.12.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 49.

Table 49 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiau Stream at Inland North Road between 1998 and March 2012 [27 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	11		A
ANNELIDA	Oligochaeta	1	17	63		VA
MOLLUSCA	*Latia*	5	10	37		A
	Potamopyrgus	4	27	100		VA
CRUSTACEA	*Paracalliope*	5	15	56		
EPHEMEROPTERA	*Austroclima*	7	24	89	XA	A
PLECOPTERA	*Zelandobius*	5	1	4		
COLEOPTERA	Elmidae	6	27	100	VA	XA
TRICHOPTERA	*Aoteapsyche*	4	25	93	VA	VA
	Hydrobiosis	5	9	33	A	
	Hudsonema	6	2	7		A
	Oxyethira	2	7	26		
	Pycnocentria	7	11	41	VA	A
	Pycnocentrodes	5	21	78	VA	A
DIPTERA	*Aphrophila*	5	14	52	A	
	Maoridiamesa	3	1	4		
	Orthocladiinae	2	21	78		VA
	Polypedilum	3	1	4		
	Tanytarsini	3	1	4		
	Austrosimulium	3	5	19		
ACARINA	Acarina	5	1	4		

Prior to the current 2012-2013 period, 21 taxa had characterised the community at this site on occasions. These have comprised eleven 'moderately sensitive' and ten 'tolerant' taxa i.e. an absence of 'highly sensitive' taxa and a relatively high proportion of 'tolerant' taxa as would be expected in the mid reaches of a lowland stream beyond the ringplain.

Predominant taxa have included five 'moderately sensitive' taxa [amphipod (*Paracalliope*), mayfly (*Austroclima*), elmid beetles, stony-cased caddisfly (*Pycnocentrodes*), and cranefly (*Aphrophila*)] and four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and orthoclad midges].

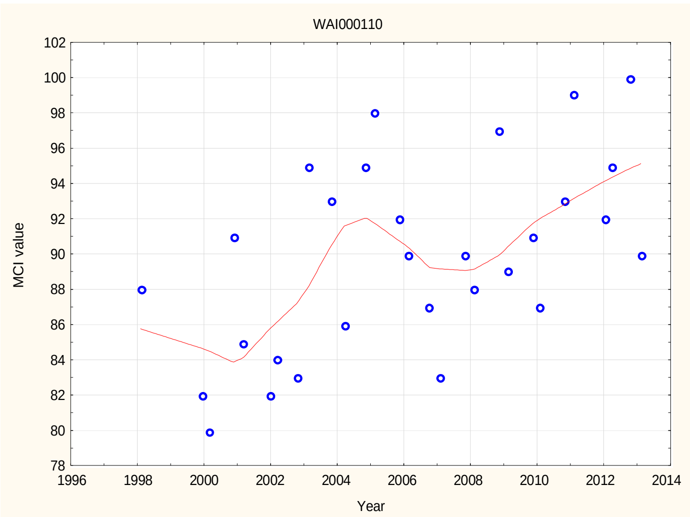
Seven of the historically characteristic taxa were dominant in the spring, 2012 community and comprised five of the predominant taxa (above). The summer, 2013 community was characterised by five of the taxa dominant in spring, together with an additional two 'moderately sensitive' taxa, both of which previously had not been predominantly characteristic of this site's communities (Table 49) and four 'tolerant' taxa. The increased abundances within several of the 'tolerant' taxa and reduced abundances within three of the 'moderately sensitive' taxa at the time of the summer survey was reflected in the decline in $SQMCI_s$ scores (1.5 units) between seasons (Tables 148 and 149). All taxa which were recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 41% to 100% of past surveys.

3.2.12.1.3 Predicted stream 'health'

The Waiau Stream rises as seepage beyond the ringplain and the site at Inland North Road is in the mid reaches at an altitude of 50 m asl. Relationships for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), therefore are not applicable in this type of small lowland stream.

3.2.12.1.4 Temporal trends in 1995 to 2013 data

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the fifteen years of SEM results collected to date from the site, in the Waiau Stream at Inland North Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 45.



N = 29
Kendall tau = +0.383
p level = 0.004 [>FDR, p = 0.009]
Significant at $p < 0.05$ level and $p < 0.01$ after FDR application

Figure 45 LOWESS trend plot of MCI data at the Inland North Road site

A strong overall improving temporal trend in MCI scores has been found which remains statistically significant ($p < 0.01$) after FDR application over the fifteen year monitoring term at this site, with the initial trend of increasing scores having been followed by some decline in scores improving again to scores slightly above those recorded in mid-programme. The range of LOWESS-smoothed scores (11 units) has been of marginal ecological significance. LOWESS-smoothed MCI scores have been indicative of 'fair' generic stream health (Table 1) throughout the period.

3.2.12.2 Discussion

Seasonal MCI values typically decreased between spring and summer (by 10 units) at this mid reach site of a lowland stream where the percentage composition of 'tolerant' taxa increased by 18% in the summer community coincident with an apparent seasonal increase in periphyton mats and filamentous algal substrate cover. Historical seasonal median scores (Appendix II) have indicated a 4 unit summer decrease at this site. Seasonal communities at this site shared 18 common taxa (56% of the 32 taxa found at this site in 2012-2013), a relatively high percentage of common taxa for this mid reach site in a lowland stream, thereby contributing to the moderate seasonal difference of 10 units in MCI values.

3.2.13 Punehu Stream

The results of the spring (2012) and summer (2012-2013) surveys are summarised in Table 150 and Table 151, Appendix I.

3.2.13.1 Wiremu Road site (PNH000200)

3.2.13.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Punehu Stream between 1995 and February 2012 at this open, upper mid-reach site in farmland, 4 km downstream of the National Park These results are summarised in Table 50 together with the results from the current period, and illustrated in Figure 46.

Table 50 Results of previous surveys performed in the Punehu Stream at Wiremu Road together with spring 2012 and summer 2012 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
PNH000200	34	19-31	27	104-133	122	25	134	27	119

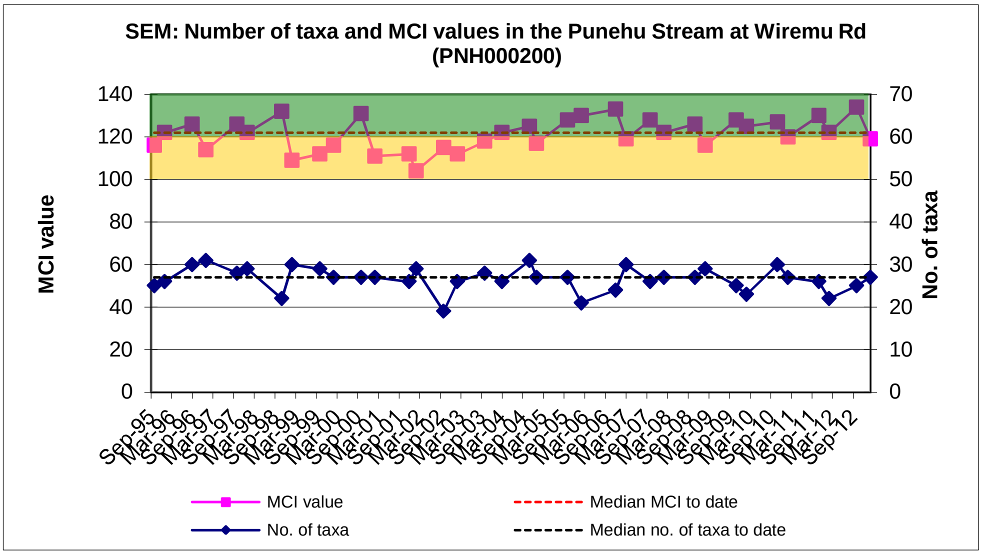


Figure 46 Numbers of taxa and MCI values in the Punehu Stream at Wiremu Road

A moderate range of richnesses (19 to 31 taxa) has been found with a median richness of 27 taxa (more representative of typical richnesses in the mid reaches of ringplain streams and rivers). During the 2012-2013 period, spring (25 taxa) and summer (27 taxa) richnesses were slightly lower and equal with the median richness respectively.

MCI values have had a moderate range (29 units) at this site, typical of a site in the (upper) mid reaches of a ringplain stream in more open farmland. The median value (122 units) has been typical of mid reach sites elsewhere on the ringplain. The spring, 2012 (134 units) and summer, 2013 (119 units) scores were a significant 12 units above and insignificant 3 units below the historical median respectively. The spring score was also one unit higher than the previous maximum score. These scores categorised this site as having 'very good' generic health (Table 1) in spring and 'good' health in summer and, in terms of predictive relationships (Table 2), 'above expected' health for the (upper) mid reaches of a ringplain stream in spring and 'expected' health in summer. The historical median score (122 units) placed this site in the 'very good' and 'expected' categories for the generic and predictive methods of assessment respectively.

3.2.13.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 51.

Table 51 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Punehu Stream at Wiremu Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	4	12		
MOLLUSCA	*Potamopyrgus*	4	1	3		
EPHEMEROPTERA	*Austroclima*	7	5	15		A
	Coloburiscus	7	31	91	A	A
	Deleatidium	8	34	100	VA	XA
	Nesameletus	9	29	85	VA	VA
PLECOPTERA	*Acroperla*	5	2	6		
	Megaleptoperla	9	5	15		
	Zelandoperla	8	24	71	VA	A
COLEOPTERA	Elmidae	6	34	100	A	VA
	Hydraenidae	8	5	15		
MEGALOPTERA	*Archichauliodes*	7	2	6		
TRICHOPTERA	*Aoteapsyche*	4	24	71		VA
	Costachorema	7	21	62		
	Hydrobiosis	5	10	29		
	Beraeoptera	8	14	41	VA	A
	Helicopsyche	10	4	12		
	Olinga	9	2	6		
	Oxyethira	2	1	3		
	Pycnocentrodes	5	21	62	A	
DIPTERA	*Aphrophila*	5	5	15		
	Eriopterini	5	8	24		
	Maoridiamesa	3	14	41		A
	Orthocladiinae	2	19	56		
	Empididae	3	1	3		

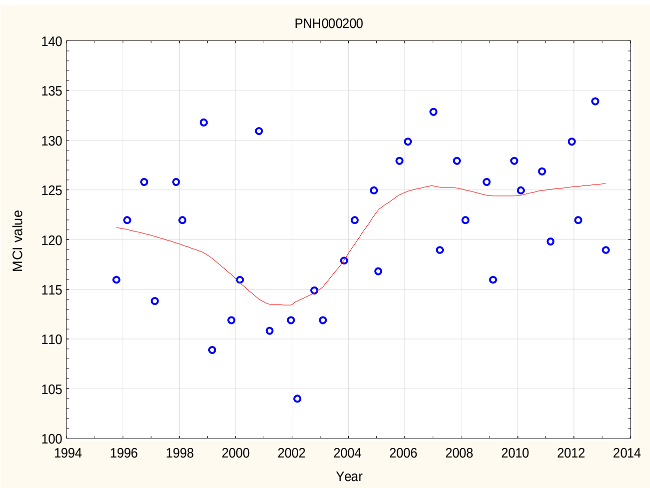
Prior to the current 2012-2013 period, 25 taxa have characterised the community at this site on occasions. These have comprised eight 'highly sensitive', ten 'moderately sensitive', and seven 'tolerant' taxa i.e. a predominance of 'sensitive' taxa as would be expected in the (upper) mid reaches of a ringplain stream. Predominant taxa have included three 'highly sensitive' taxa [mayflies (*Deleatidium* on every occasion, and *Nesameletus*) and stonefly (*Zelandoperla*)]; four 'moderately sensitive' taxa [mayfly (*Coloburiscus*), elmid beetles (on very occasion), stony-cased caddisfly (*Pycnocentrodes*), and free-living caddisfly (*Costachorema*)]; and two 'tolerant' taxa [net-building caddisfly (*Aoteapsyche*) and orthoclad midges]. Six of these predominant taxa were dominant in the spring, 2012 community together with one other ('highly sensitive') taxon while the summer, 2013 community was characterized by six of the taxa dominant in spring together with an additional two 'tolerant' and one 'moderately sensitive' taxa. All but three of these taxa previously had been predominantly characteristic of this site's communities (Table 51). Some increase in numerical dominance of one tolerant taxon and decreased abundances of two 'highly sensitive' taxa in summer were reflected by a small seasonal decrease in $SQMCI_s$ scores of 0.6 unit (Tables 150 and 151). All taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 41% to 100% of the past surveys.

3.2.13.1.3 Predicted stream 'health'

The Punehu Stream site at Wiremu Road is 4.4 km downstream of the National Park boundary at an altitude of 270 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 112 (altitude) and 115 (distance) for this site. The historical site median (122 units) is 10 units higher than the altitude prediction and 7 units above the distance predictive value. The spring, 2012 survey score (134 units) was significantly (Stark, 1998) higher than both predictive values while the summer, 2013 score (119 units) was higher (by 7 units) than the altitude predictive value and 4 units above the distance predictive value. Of the 36 surveys to date at this site, only 8% of MCI scores have been less than 112 units while 78% have been greater than 115 units.

3.2.13.1.4 Temporal trends 1995 to 2013

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site, in the Punehu Stream at Wiremu Road. The MCI has been chosen as the preferable indicator 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 47.



N = 36
Kendall tau = +0.251
p level = 0.031 [>FDR, p = 0.060]
Significant at $p < 0.05$ level; not significant after FDR

Figure 47 LOWESS trend plot of MCI data at the Wiremu Road site

Although a steady increase in MCI scores had been apparent between 2002 and 2007, the positive trend in scores over the entire period has not been statistically significant after FDR (at p<0.05 level). The range of LOWESS-smoothed scores (13 units) has some ecological significance, particularly since 2002 (coincident with localised riparian fencing and planting of the true left-bank of the stream). Overall, smoothed MCI scores were indicative of 'good' generic stream health (Table 1) until 2001 improving to 'very good' health in more recent years.

In terms of predictive relationships (Table 2) for a site in the upper mid reaches of a ringplain stream, stream health has been in the' expected' category for the period to 2004 improving to 'better than expected' since that date.

3.2.13.2 SH 45 site (PNH000900)

3.2.13.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this lower reach site at SH 45 in the Punehu Stream between 1995 and February 2012. These results are summarised in Table 52, together with the results from the current period, and illustrated in Figure 48.

Table 52 Results of previous surveys performed in the Punehu Stream at SH 45 together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
PNH000900	34	10-26	21	70-105	87	19	106	22	85

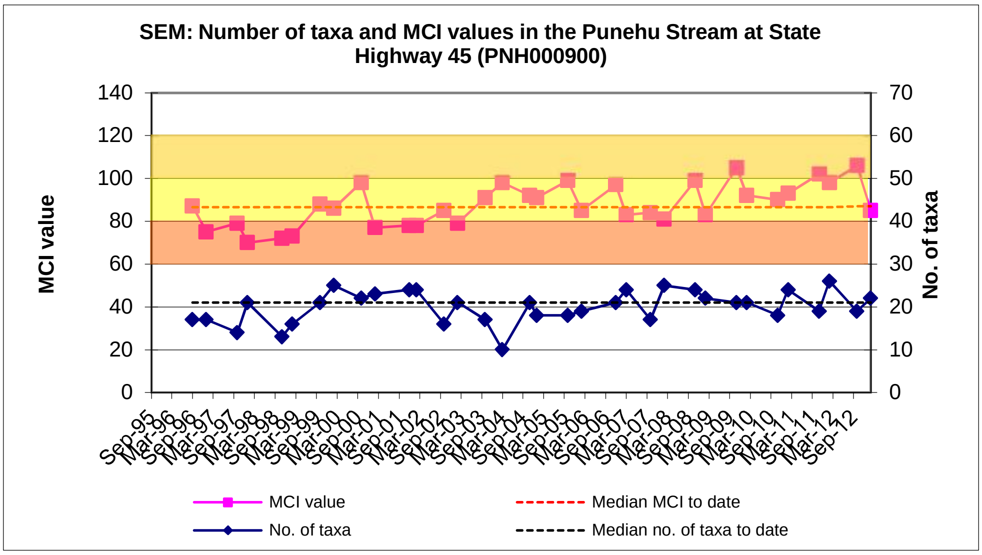


Figure 48 Numbers of taxa and MCI values in the Punehu Stream at SH 45

A moderate range of richnesses (10 to 26 taxa) has been found with a median richness of 21 taxa, relatively typical of richnesses in the lower reaches of ringplain streams and rivers. During the 2012-2013 period, spring (19 taxa) and summer (22 taxa) richnesses varied over a small range but were within two taxa of the median taxa number on both occasions coincident with similar substrate patchy periphyton cover but higher water temperature at the time of the summer survey.

MCI scores have had a relatively wide range (35 units) at this site, typical of sites in the lower reaches of ringplain streams. The median value (87 units) has been relatively typical of lower reach sites elsewhere on the ringplain. The spring, 2012 (106 units) score was higher than typical for such a site and above the historical median by a significant 19 units whereas the summer 2013 score (85 units) was typical for such a site and within two units of the historical median value. These scores categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower reaches of a ringplain stream. The historical median score (87 units) placed this site in the 'fair' category for generic and 'expected' category for predictive methods of assessment.

3.2.13.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 53.

Table 53 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Punehu Stream at SH 45 between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	29	85		
MOLLUSCA	*Potamopyrgus*	4	19	56		A
EPHEMEROPTERA	*Austroclima*	7	2	6		
	Coloburiscus	7	2	6	A	
	Deleatidium	8	12	35	XA	VA
PLECOPTERA	*Acroperla*	5	1	3		
COLEOPTERA	Elmidae	6	22	65	A	A
MEGALOPTERA	*Archichauliodes*	7	6	18		A
TRICHOPTERA	*Aoteapsyche*	4	16	47	A	VA
	Hydrobiosis	5	16	47		
	Oxyethira	2	4	12		
	Pycnocentrodes	5	11	32	VA	A
DIPTERA	*Aphrophila*	5	17	50		
	Maoridiamesa	3	17	50		
	Orthocladiinae	2	30	88		A
	Tanytarsini	3	9	26		
	Ceratopogonidae	3	1	3		
	Empididae	3	6	18		
	Muscidae	3	2	6		
	Austrosimulium	3	4	12		

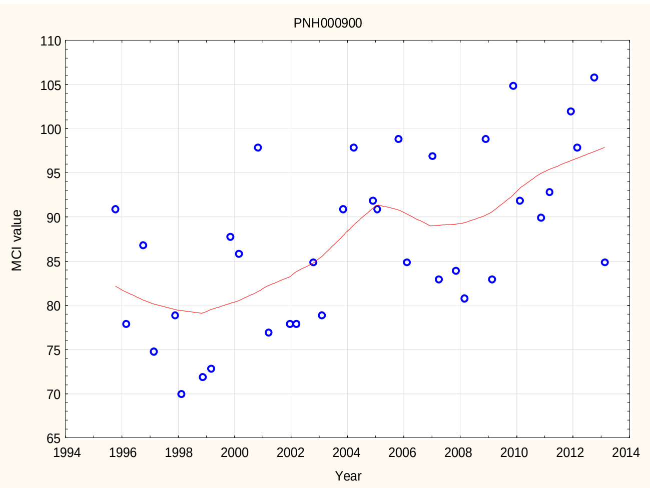
Prior to the current 2012-2013 period 20 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', eight 'moderately sensitive', and eleven 'tolerant' taxa i.e. a higher proportion of 'tolerant' taxa as might be expected in the lower reaches of a ringplain stream. Predominant taxa have included no 'highly sensitive' taxa; two 'moderately sensitive' taxa [elmid beetles, and cranefly (*Aphrophila*)]; and four 'tolerant' taxa (oligochaete worms, snail (*Potamopyrgus*), and midges (orthoclads and *Maoridiamesa*)]. Five of the historically characteristic taxa, one of which had been predominant, were dominant in the spring 2012 community. These comprised one 'highly sensitive', three 'moderately sensitive', and one 'tolerant' taxa, whereas one 'highly sensitive', three 'moderately sensitive', and three 'tolerant' taxa comprised the dominant taxa in the summer community. Four of these eight taxa were dominant in both spring and summer communities (Table 53). An increase in the proportional dominance by 'tolerant' taxa, an increase in numerical abundance within one 'tolerant' taxon, and a decrease in numerical abundance within the one 'highly sensitive' taxon combined to cause a decrease in summer $SQMCI_s$ score of 1.8 units (Tables 150 and 151). The three taxa recorded as very or extremely abundant during spring and summer had characterised this site's communities on 32% to 47% of past surveys.

3.2.13.2.3 Predicted stream 'health'

The Punehu Stream site at SH 45 is 20.9 km downstream of the National Park boundary at an altitude of 20 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 86 (altitude) and 98 (distance) for this site. The historical site median (87 units) was only one unit above the altitude prediction but a significant (Stark, 1998) 11 units lower than the distance predictive value. The spring, 2012 survey score (106 units) was from 8 to 20 units above these predictive values and the summer, 2013 score (85 units) was from 1 to a significant 13 units below predictive values. Of the 36 surveys to date at this site, 36% of MCI scores have been less than 86 units while only 17% have been greater than 98 units.

3.2.13.2.4 Temporal trends in 1995 to 2013

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006), has been performed on the eighteen years of SEM results collected to date from the site, in the Punehu Stream at SH 45. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 49.



N = 36
Kendall tau = +0.415
p level = 0.0004 [>FDR, p = 0.001]
Significant at p <0.05 and p < 0.01;
and after FDR

Figure 49 LOWESS trend plot of MCI data at the SH 45 site

This site's MCI scores show a strong positive temporal trend over the eighteen year period which was statistically significant ($p<0.01$) after FDR application.

The LOWESS-smoothed MCI scores' range (19 units) has been ecologically significant within this period with scores mainly indicative of 'poor' generic stream health (Table 1) prior to early 1999 improving to 'fair' health throughout the subsequent period and approaching 'good' health more recently. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, stream health improved from 'worse than expected' very early in the period to the 'expected' category throughout the remainder of the period although issues have occurred with consented dairy shed discharge compliance and cumulative impacts of such discharges in the Mangatawa Stream sub-catchment in the local vicinity of the site (TRC, 2011).

3.2.13.3 Discussion

Seasonal MCI values typically deteriorated between spring and summer at the upper mid-reach (Wiremu Road) site but by a wider (significant) margin of 15 units which was seven units more than the historical median seasonal difference for this site (Appendix II). A larger and typical decrease (21 units) was found at the lower reach site (SH 45) in comparison with the historical seasonal median decrease of 8 units (Appendix II). Seasonal communities at the upper mid reach site shared 23 common taxa (79% of the 29 taxa found at this site in 2012-2013) compared with 14 shared common taxa (52% of the 27 taxa found in 2012-2013) at the lower reaches site (SH 45), a typically more pronounced seasonal change in community structure at the lower of the two sites. The two sites shared 14 common taxa (47% of the 30 taxa) in spring and 15 common taxa (44% of 34 taxa) in summer, indicative of the dissimilarity in spatial community structures to almost the same degree in both spring and summer.

MCI score typically fell in a downstream direction in both spring (by 28 units) and typically more markedly in summer (by 34 units), over a stream distance of 16.5 km through the mid to lower reaches of this stream. These falls in MCI scores equated to rates of decline of 1.7 units/km in spring increasing to 2.1 units/km in summer, compared with a predicted rate of 1.0 unit/km over the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009) indicative of additional impacts from diffuse and point-source discharges within this extensive dairying catchment.

Using the long-term median SEM MCI scores for both sites (Appendix II), the overall rate of decline has been 2.1 MCI units/km over the surveyed length. Therefore rates of decline over the 2012–2013 period were lower in spring and equivalent in summer to the historical average rate.

3.2.14 Patea River

The results of spring (2012) and summer (2012-2013) surveys are presented in Table 152 and Table 153, Appendix I.

3.2.14.1 Barclay Road site (PAT000200)

3.2.14.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this upper reach, shaded site adjacent to the National Park boundary in the Patea River between 1995 and February 2012. These results are summarised in Table 54, together with the results from the current period, and illustrated in Figure 50.

Table 54 Results of previous surveys performed in the Patea River at Barclay Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
PAT000200	34	25-35	31	127-145	137	24	138	32	141

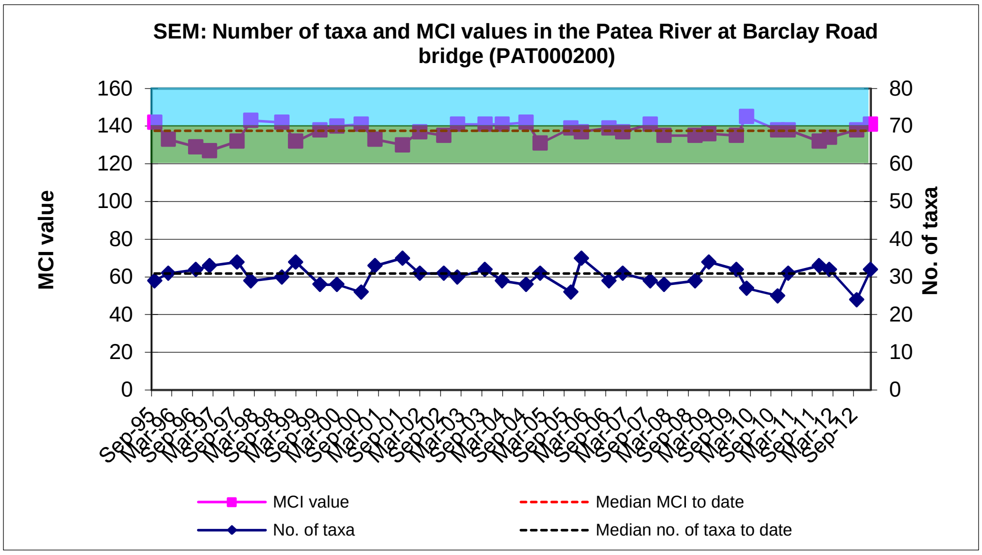


Figure 50 Numbers of taxa and MCI values in the Patea River at Barclay Road

A moderate range of richnesses (25 to 35 taxa) has been found with a relatively high median richness of 31 taxa, typical of richnesses in ringplain streams and rivers near the National Park boundary. During the 2012-2013 period spring richness (24 taxa) was well below the median and one fewer than the historical minimum at this site while summer (32 taxa) richness was one taxon above this median richness. This was coincident with very thin periphyton mat layers on the predominantly stony-bouldery substrate of this shaded site on both survey occasions.

MCI values have had a moderate range (18 units) at this site, more typical of a National Park boundary site. The high median value (137 units) has been typical of upper reach sites elsewhere on the ringplain, and the spring, 2012 (138 units) and summer, 2013 (141 units) scores continued this trend for such a site. These scores were higher (by 1 to 4 units) than the historical median and within four units of the previous maximum value at this site in summer.

They categorised this site as having 'very good' (spring) and 'excellent' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' and 'better than expected' health for the upper reaches of a ringplain stream in spring and summer respectively. The historical median score (137 units) placed this site in the 'very good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.14.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 55.

Table 55 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Patea River at Barclay Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
EPHEMEROPTERA	*Austroclima*	7	6	18		
	Coloburiscus	7	34	100	A	VA
	Deleatidium	8	34	100	XA	VA
	Nesameletus	9	4	12		A
PLECOPTERA	*Acroperla*	5	1	3		
	Austroperla	9	1	3		
	Megaleptoperla	9	14	41		A
	Zelandobius	5	13	38		
	Zelandoperla	8	26	76	A	A
COLEOPTERA	Elmidae	6	30	88		A
	Hydraenidae	8	11	32		A
MEGALOPTERA	*Archichauliodes*	7	7	21		
TRICHOPTERA	*Costachorema*	7	1	3	VA	
	Hydrobiosis	5	1	3		
	Hydrobiosella	9	2	6		
	Orthopsyche	9	25	74		A
	Beraeoptera	8	15	44	A	
	Helicopsyche	10	13	38		
	Olinga	9	1	3		
	Zelolessica	7	1	3		
DIPTERA	*Aphrophila*	5	32	94		A
	Orthocladiinae	2	16	47		
	Polypedilum	3	2	6		

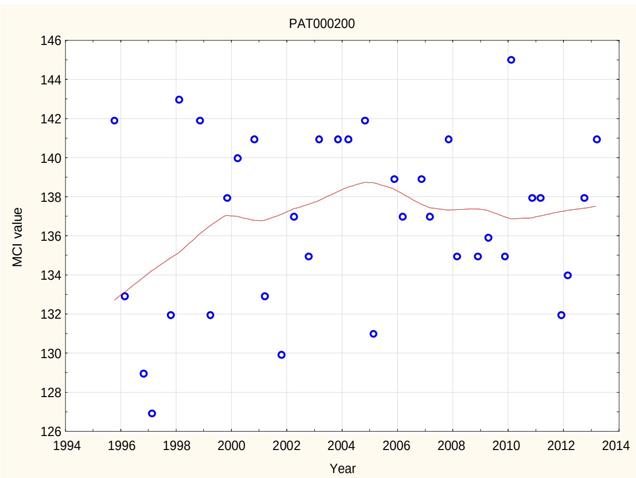
Prior to the current 2012-2013 period, 23 taxa had characterised the community at this site on occasions. These have comprised eleven 'highly sensitive', ten 'moderately sensitive', and only two 'tolerant' taxa i.e. a majority of 'highly sensitive' taxa as would be expected near the National Park boundary of a ringplain river. Predominant taxa have included three 'highly sensitive' taxa [mayfly (*Deleatidium* on every sampling occasion), stonefly (*Zelandoperla*), and caddisfly (*Orthopsyche*)]; three 'moderately sensitive' taxa [(mayfly (*Coloburiscus* on every occasion), elmid beetles, and cranefly (*Aphrophila*)); but no 'tolerant' taxa. Five of the characteristic taxa were dominant in the spring, 2012 community, four of which were predominant taxa. Three of these taxa again were dominant in the summer, 2013 community together with four additional 'highly sensitive' and two 'moderately sensitive' taxa, all historically characteristic of this site. No 'tolerant' taxa were dominant in either spring or summer. Despite some variability amongst the 'highly' and 'moderately' sensitive taxa numerical dominances, seasonal $SQMCI_s$ values were within 0.2 unit (Tables 152 and 153). The three taxa recorded as very or extremely abundant during spring and/or summer had characterized this site's communities on 3% to 100% of past surveys.

3.2.14.1.3 Predicted stream 'health'

The Patea River site at Barclay Road is 1.9 km downstream of the National Park boundary at an altitude of 500 m asl. Some bush cover extends from the National Park adjacent to most of the reach upstream of this site which is situated in farmland. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 135 (altitude) and 125 (distance) for this site. The historical site median (137 units) is only 2 units higher than the altitude prediction but a significant 12 units above the distance predictive value. The spring, 2012 score (138 units) and summer, 2013 score (141 units) were very similar and 3 to a significant 16 units higher than these predictive values. Of the 36 surveys to date at this site, no MCI scores have been less than 125 units while 64% have been greater than 135 units.

3.2.14.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Patea River at Barclay Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 51.



N = 36
Kendall tau = +0.049
p value = 0.673 [>FDR, p = 0.712]
N/S at p <0.05

Figure 51 LOWESS trend plot of MCI data at the Barclay Road site

No statistically significant temporal trend in MCI scores has been found at this upper catchment site over the 18 year monitoring period during which there has been a minimal overall trend. Neither has the range of LOWESS-smoothed scores (6 units) shown ecological significance. Smoothed MCI scores have consistently indicated 'very good', bordering on 'excellent', generic river health (Table 1) at this relatively pristine site just outside the National Park boundary and in terms of predictive relationships (Table 2), river health has been in the 'expected' category throughout the period.

3.2.14.2 Swansea Road site (PAT000315)

3.2.14.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Patea River at this mid-reach site at Swansea Road, Stratford between 1995 and February 2012. These results are summarised in Table 56, together with the results from the current period, and illustrated in Figure 52.

Table 56 Results of previous surveys performed in the Patea River at Swansea Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
PAT000315	34	20-32	26	99-130	110	23	118	30	106

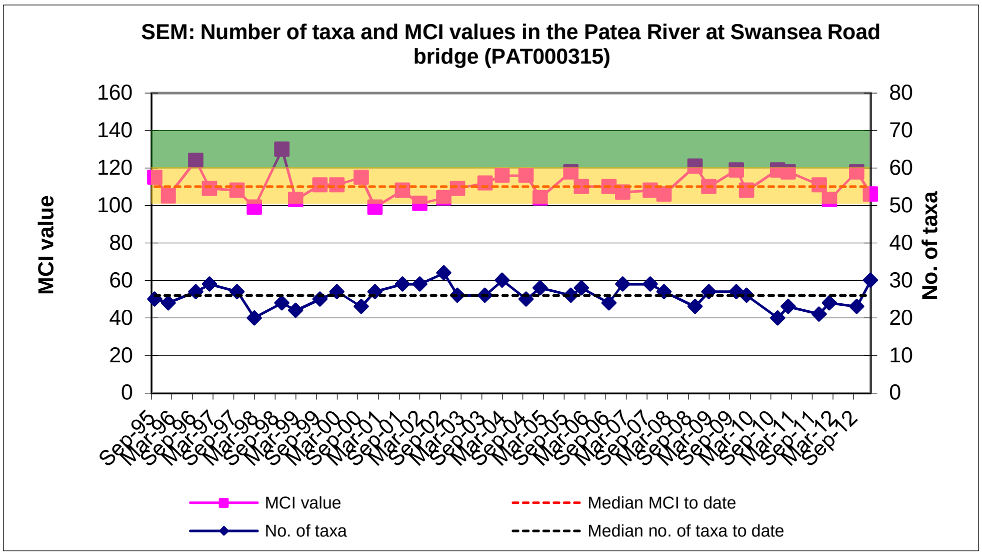


Figure 52 Numbers of taxa and MCI values in the Patea River at Swansea Road

A moderate range of richnesses (20 to 32 taxa) has been found, with a median richness of 26 taxa typical of richnesses in the mid reaches of ringplain streams and rivers. During the 2012-2013 period, spring (23 taxa) and summer (30 taxa) richnesses were relatively different; slightly lower than the median taxa number in spring coincident with minimal substrate periphyton cover, and four taxa above the median in summer when periphyton substrate cover was more widespread.

MCI values have had a relatively wide range (31 units) at this site, more so than typical of many sites in the mid reaches of ringplain rivers. The median value (110 units) has been relatively typical of scores in mid-reach sites elsewhere on the ringplain however, with the spring, 2012 (118 units) and summer, 2013 (106 units) scores eight units above and four units below the historical median respectively. These scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' health on both occasions for the mid reaches of a ringplain river. The historical median score (110 units) placed this site in the 'good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.14.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 57.

Table 57 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Patea River at Swansea Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	8	24		
EPHEMEROPTERA	*Austroclima*	7	13	38		
	Coloburiscus	7	34	100	VA	VA
	Deleatidium	8	27	79	XA	XA
	Nesameletus	9	12	35	A	A
PLECOPTERA	*Acroperla*	5	4	12		
	Zelandoperla	8	11	32		
COLEOPTERA	Elmidae	6	21	62		VA
	Hydraenidae	8	5	15		A
MEGALOPTERA	*Archichauliodes*	7	13	38		A
TRICHOPTERA	*Aoteapsyche*	4	25	74		VA
	Costachorema	7	19	56		
	Hydrobiosis	5	5	15		
	Neurochorema	6	4	12		
	Beraeoptera	8	8	24		
	Pycnocentrodes	5	4	12		
DIPTERA	*Aphrophila*	5	30	88	A	VA
	Eriopterini	5	1	3		
	Maoridiamesa	3	24	71		
	Orthocladiinae	2	31	91		A
	Tanytarsini	3	10	29		
	Muscidae	3	2	6		
	Austrosimulium	3	10	29		

Prior to the current 2012-2013 period, 23 taxa had characterised the community at this site on occasions. These have comprised five 'highly sensitive', eleven 'moderately sensitive', and seven 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and a downstream increase in 'tolerant' taxa as would be expected in the mid reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; four 'moderately sensitive' taxa [mayfly (*Coloburiscus*), elmid beetles, free-living caddisfly (*Costachorema*), and cranefly (*Aphrophila*)]; and three 'tolerant' taxa [net-building caddisfly (*Aoteapsyche*) and midges (*Maoridiamesa* and orthoclads)]. Four of these historically characteristic taxa (three predominant taxa) were dominant in the spring 2012 community. These comprised two 'highly sensitive' taxa, two 'moderately sensitive' taxa, but no 'tolerant' taxa, whereas three 'highly sensitive', four 'moderately sensitive', and two 'tolerant' taxa comprised the dominant taxa of the summer community. Four of these nine taxa were dominant in both spring and summer communities (Table 57). Increases in numerical dominances amongst some 'moderately sensitive' and 'tolerant' taxa were reflected in the decrease of 0.9 unit in $SQMCI_s$ score in summer (Tables 152 and 153).

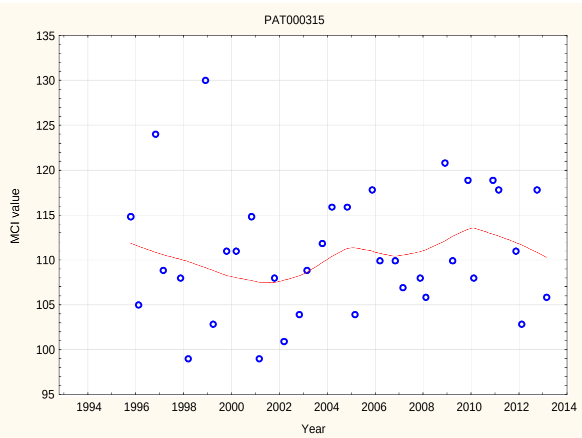
The five taxa found as very and/or extremely abundant by the seasonal surveys have characterised this site's communities on 62% to 100% of past survey occasions.

3.2.14.2.3 Predicted stream 'health'

The Patea River site at Swansea Road, Stratford is 12.9 km downstream of the National Park boundary at an altitude of 300 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 115 (altitude) and 103 (distance) for this site. The historical site median (110) is 5 units lower than the altitude prediction and 7 units higher than the distance predictive value while the spring, 2012 survey score (118 units) was 3 units higher than the predictive altitude value but a significant 15 units higher than the predictive distance value. The summer, 2013 score (106 units) was 9 units below the predictive altitude value but 3 units above the predictive distance value. Of the 36 surveys to date at this site, only 8% of MCI scores have been less than 103 units while 28% have been greater than 115 units.

3.2.14.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Patea River at Swansea Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 53.



N = 36
Kendall tau = +0.095
p value = 0.413 [>FDR, p = 0.524]
N/S at p <0.05

Figure 53 LOWESS trend plot at the Swansea Road site

The slight positive temporal trend in MCI scores was not statistically significant over the eighteen year period. The range of LOWESS-smoothed scores (6 units) was of no ecological significance. Smoothed MCI scores consistently indicated 'good' generic river health (Table 1) throughout the monitoring period. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a ringplain river, health has been in the 'expected' category for the entire period.

3.2.14.3 Skinner Road site (PAT000360)

3.2.14.3.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Patea River at this mid-reach site at Skinner Road (some 6 km downstream of the Swansea Road, Stratford site), between 1995 and February 2012. These results are summarised in Table 58, together with the results from the current period, and illustrated in Figure 54.

Table 58 Results of previous surveys performed in the Patea River at Skinner Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
PAT000360	34	15-33	24	86-105	98	18	100	25	95

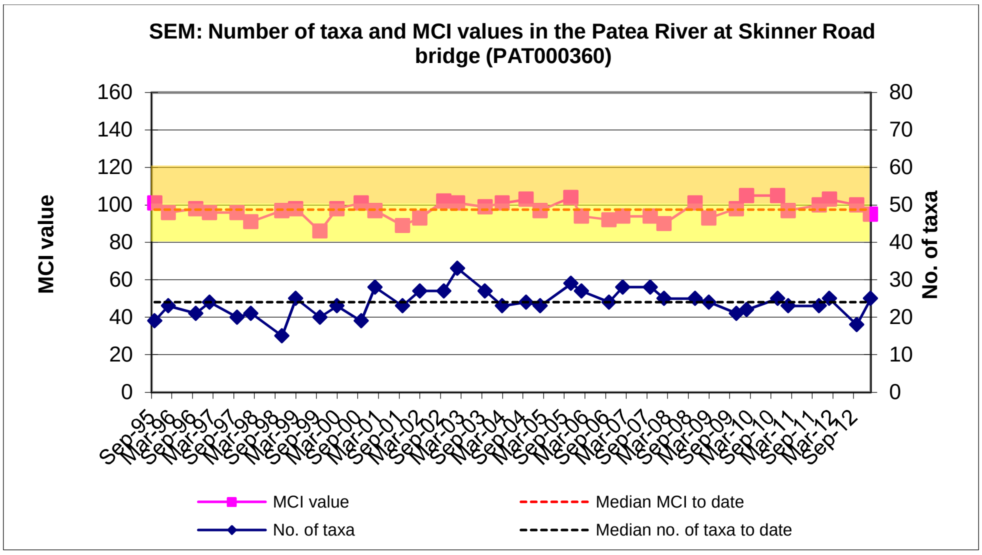


Figure 54 Numbers of taxa and MCI values in the Patea River at Skinner Road

A wide range of richnesses (15 to 33 taxa) has been found with a median richness of 24 taxa (more representative of typical richnesses in the mid-reaches of ringplain streams and rivers). During the 2012-2013 period spring (18 taxa) and summer (25 taxa) richnesses were dissimilar and six taxa lower than the median taxa number in spring and one taxon above median richness in summer when substrate periphyton cover was more widespread.

MCI values have had a moderate range (19 units) at this site, typical of sites in the mid-reaches of ringplain streams and rivers. The median value (98 units) has been relatively typical of the range of scores at mid-reach sites elsewhere on the ringplain however. The spring, 2012 (100 units) and summer, 2013 (95 units) scores were relatively similar, typical of scores for such a site, and 2 units higher (spring) and 3 units lower (summer) than the historical median. These scores categorised this site as having 'good' (spring) and 'fair'(summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' health in spring and summer respectively for the mid-reaches of a ringplain river. The historical median score (98 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.14.3.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 59.

Table 59 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Patea River at Skinner Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	5	15		
ANNELIDA	Oligochaeta	1	24	71		
MOLLUSCA	*Potamopyrgus*	4	10	29		
CRUSTACEA	*Paracalliope*	5	1	3		
EPHEMEROPTERA	*Austroclima*	7	1	3		
	Coloburiscus	7	9	26	A	A
	Deleatidium	8	13	38	XA	VA
PLECOPTERA	*Acroperla*	5	2	6		
COLEOPTERA	Elmidae	6	26	76	A	VA
MEGALOPTERA	*Archichauliodes*	7	13	38	A	A
TRICHOPTERA	*Aoteapsyche*	4	27	79	A	XA
	Costachorema	7	11	32		
	Hydrobiosis	5	17	50		A
	Oxyethira	2	4	12		
	Pycnocentrodes	5	8	24		
DIPTERA	*Aphrophila*	5	25	74	A	VA
	Maoridiamesa	3	28	82		VA
	Orthocladiinae	2	34	100	A	A
	Tanytarsini	3	16	47		VA
	Empididae	3	2	6		
	Muscidae	3	8	24		
	Austrosimulium	3	8	24		

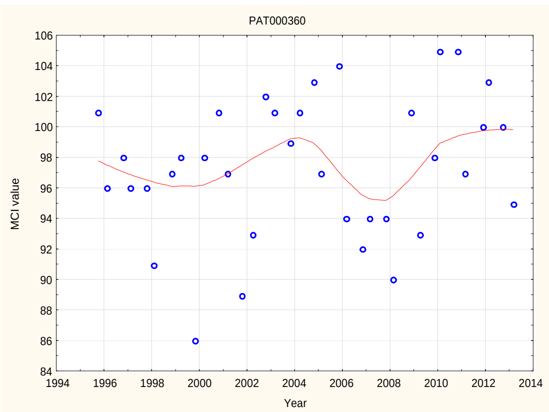
Prior to the current 2012-2013 period, 22 taxa had characterised the community at this site on occasions. These have comprised one 'highly sensitive', ten 'moderately sensitive', and eleven 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and relatively high proportions of 'moderately sensitive' and 'tolerant' taxa as would be expected in the mid-reaches of a ringplain river. Predominant taxa have included no 'highly sensitive' taxa, three 'moderately sensitive' taxa [elmid beetles, free-living caddisfly (*Hydrobiosis*), and cranefly (*Aphrophila*)] and four 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]. Seven of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised one 'highly sensitive', four 'moderately sensitive', and two 'tolerant' taxa, whereas one 'highly sensitive', five 'moderately sensitive', and four 'tolerant' taxa comprised the dominant taxa of the summer, 2013 community. Seven of these ten taxa were dominant in both spring and summer communities (Table 61). A typical increase in summer numerical dominance by three 'tolerant' taxa (e.g. oligochaete worms) in particular and some decrease in the abundance of the single characteristic 'highly sensitive' mayfly, were reflected in the significant decrease of 2.7 units in $SQMCI_s$ scores between spring and summer (Tables 152 and 153). The six taxa found as very and/or extremely abundant by the seasonal surveys have characterised this site's communities on 38% to 82% of past survey occasions although only one of these taxa was predominant in both spring and summer surveys.

3.2.14.3.3 Predicted stream 'health'

The Patea River site at Skinner Road is 19.2 km downstream of the National Park boundary at an altitude of 240 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 109 (altitude) and 99 (distance) for this site. The historical site median (98) is a significant (Stark, 1998) 11 units lower than the altitude prediction but only one unit lower than the distance predictive value. The spring, 2012 survey score (100 units) was 9 units lower than the altitude predictive value while the summer, 2013 score (95 units) was a significant 14 units lower than the predictive altitude value but 4 units below the predicted distance value. Of the 36 surveys to date at this site, 61% of MCI scores have been less than 99 units while no scores have been greater than 109 units, indicative of some deterioration in river 'health' in comparison with the historical record at the nearest upstream site (at Swansea Road) in Stratford township.

3.2.14.3.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Patea River at Skinner Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 55.



N = 34
Kendall tau = +0.143
p value = 0.219 [>FDR, p = 0.306]
N/S at $p < 0.05$

Figure 55 LOWESS trend plot at the Skinner Road site

The small positive temporal trend in MCI scores over the eighteen year period was not statistically significant. An apparent decline in scores between 2004 and 2008 has been followed by some recent improvement. The range of LOWESS-smoothed scores (5 units) had no ecological significance over the period. Smoothed MCI scores consistently indicated 'fair' generic river health (Table 1) briefly bordering on 'good' health six years ago and most recently. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a ringplain river, health has remained in the 'expected' category throughout the eighteen year period, briefly bordering on 'worse than expected' in 2007-2008.

3.2.14.4 Discussion

Seasonal MCI values remained very similar between spring and summer at one site (typically at Barclay Road) while at the Swansea Road and Skinner Road sites, more typical summer decreases in MCI score (12 units and 5 units) were recorded which were within 3 units of the historical median seasonal differences for these sites (Appendix II). Seasonal communities shared 56% of the 36 taxa at the upper site, 57% of 36 taxa at Swansea Road, and 54% of 28 taxa at the furthest downstream site in the middle reaches indicative of a very narrow range in seasonal community composition dissimilarities at all sites.

The MCI scores fell in a downstream direction between the upper site and the furthest downstream middle reaches site by 38 units in spring and 46 units in summer, over a river distance of 17.3 km. These seasonal falls in MCI scores equated to rates of decline of about 2.2 units/km in spring and 2.7 units/km in summer, higher than a predicted rate of 1.5 units/km for the equivalent length and reach of a National Park-sourced river (Stark and Fowles, 2009). This was typical of the general trend of past summer seasonal increases in rates of decline.

Between the upper reach site and Swansea Road mid-reach site, the spring (1.8 units/km) and summer (3.2 units/km) rates of decline were similar to (spring) and much higher than (summer) the predicted rate (2.0 units/km) for the equivalent river reach. For the Swansea Road mid-reach to Skinner Road mid-reach sites, the spring (2.9 units/km) rate of decline was well above the predicted rate of 0.6 unit/km, but atypically there was a lower rate of decline (1.7 units/km) in summer.

Using the long-term median SEM MCI scores for each site (Appendix II), the rates of decline between upper reach and Swansea Road mid-reach, and the Swansea Road mid-reach and Skinner Road mid-reach sites have been about 2.6 and 1.9 units per km respectively with an overall average rate of decline of 2.3 MCI units/km over the surveyed length. Therefore rates of MCI decline in the 2012-2013 period were mainly higher, but more variable in summer, than average rates for the 1995 to 2012 period for the various surveyed reaches of the river.

Community composition varied markedly through the upper to mid-reach length of the river surveyed. A total of 34 taxa was recorded in spring of which only 10 taxa were present at all three sites. These included two 'highly sensitive', seven 'moderately sensitive', and one 'tolerant' taxa with only the 'highly sensitive' ubiquitous mayfly *Deleatidium* and one 'moderately sensitive' taxon [mayfly (*Coloburiscus*)] abundant at all three sites. A higher total of 50 taxa was found along the river's length by the summer survey of which only 8 taxa were present at all three sites. These were very similar to the widespread taxa in spring with the loss of one 'moderately sensitive' and one 'tolerant' taxa. Only the one 'highly sensitive' mayfly taxon and three 'moderately sensitive' taxa were abundant at all three sites. These dissimilarities in spatial community structure along the surveyed length (upper to mid-reaches) of the Patea River typically were more pronounced in summer.

3.2.15 Mangaehu River

The results found by the 2012-2013 surveys are presented in Table 154 and Table 155 Appendix I for this single site in the lower reaches of a large eastern hill country river.

3.2.15.1 Raupuha Road site (MGH000950)

3.2.15.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this lower reach site in the Mangaehu River between 1995 and February 2012. These results are summarised in Table 60, together with the results from the current period, and illustrated in Figure 56.

Table 60 Results of previous surveys performed in the Mangaehu River at Raupuha Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MGH000950	34	13-26	19	77-104	88	20	101	24	91

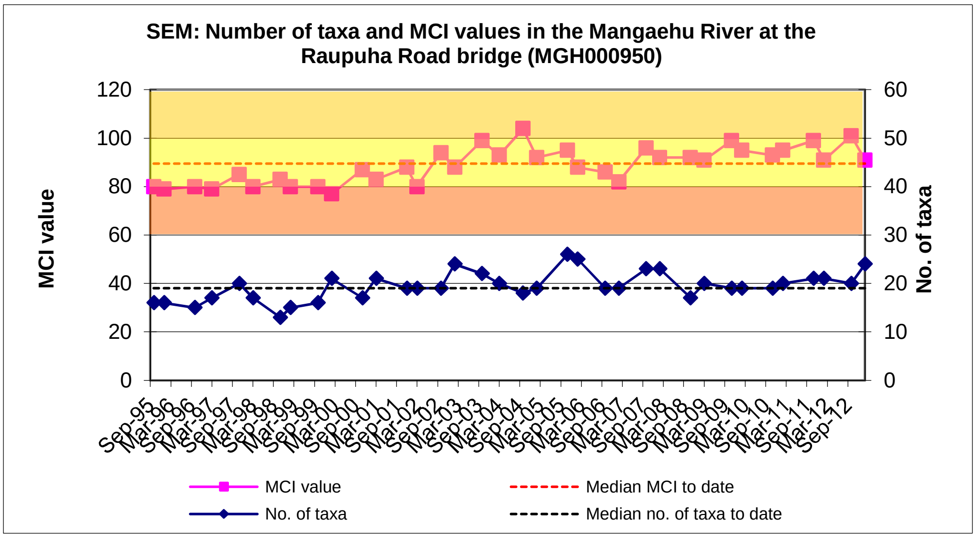


Figure 56 Numbers of taxa and MCI values in the Mangaehu River at Raupuha Road

A relatively wide range of richnesses (13 to 26 taxa) has been found with a moderate median richness of 19 taxa (lower than typical richnesses in the lower reaches of eastern hill country rivers). During the 2012-2013 period, spring (20 taxa) and summer (24 taxa) richnesses were higher in summer under more widespread substrate periphyton cover and slightly higher than this median richness.

MCI values have had a relatively wide range (27 units) at this site and typical of a site in the lower reaches of streams and rivers. The median value (88 units) has been more typical of lower reach sites elsewhere and four units less than the median score (92 units) recorded by 47 previous surveys at 'control' sites located at similar altitudes (to the Raupuha Road site) in eastern hill country rivers and streams (TRC, 1999 (updated, 2012)). The spring, 2012 (101 units) and summer, 2013 (91 units) scores were 3 to a significant 13 units higher than the historical median. These scores categorised this site as having 'good' and 'fair' health generically (Table 1) in spring and summer respectively.

The historical median score (88 units) placed this site in the 'fair' category for the generic method of assessment.

3.2.15.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 61.

Table 61 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Mangaehu River at Raupuha Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	3		
ANNELIDA	Oligochaeta	1	4	12		
MOLLUSCA	*Potamopyrgus*	4	9	26		
CRUSTACEA	*Paracalliope*	5	5	15		
EPHEMEROPTERA	*Austroclima*	7	8	24		VA
	Coloburiscus	7	0			A
	Deleatidium	8	2	6	VA	
	Mauiulus	5	1	3		
	Zephlebia group	7	3	9	A	
PLECOPTERA	*Acroperla*	5	8	24		
COLEOPTERA	Elmidae	6	4	12		
TRICHOPTERA	*Aoteapsyche*	4	19	56		VA
	Costachorema	7	6	18	A	
	Hydrobiosis	5	15	44		VA
	Oxyethira	2	2	6		
	Pycnocentrodes	5	13	38		A
DIPTERA	*Aphrophila*	5	27	79	A	VA
	Maoridiamesa	3	22	65	VA	A
	Orthocladiinae	2	32	94	VA	VA
	Tanytarsini	3	14	41		VA
	Empididae	3	4	12		
	Muscidae	3	7	21		
	Austrosimulium	3	6	18		

Prior to the current 2012-2013 period, 22 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', ten 'moderately sensitive', and 11 'tolerant' taxa i.e. a high proportion of 'tolerant' taxa as would be expected in the lower reaches of an eastern hill-country river. Predominant taxa have included only one 'moderately sensitive' taxon [cranefly (*Aphrophila*)] and three 'tolerant' taxa [net-building caddisfly (*Aoteapsyche*) and midges (*Maoridiamesa* and orthoclads)]. Three of these predominant taxa were dominant in the spring, 2012 community together with three other historically characteristic taxa. The summer, 2013 community was characterised by three more than the taxa dominant in spring, together with one additional taxon ['moderately sensitive' mayfly (*Coloburiscus*)], which previously had not been characteristic of this site's communities (Table 61). Despite several seasonal differences in characteristic taxa, particularly a decrease in the summer numerical abundance of the 'highly sensitive' mayfly, there was only a small decrease of 0.3 unit in the summer $SQMCI_s$ score (Tables 154 and 155).

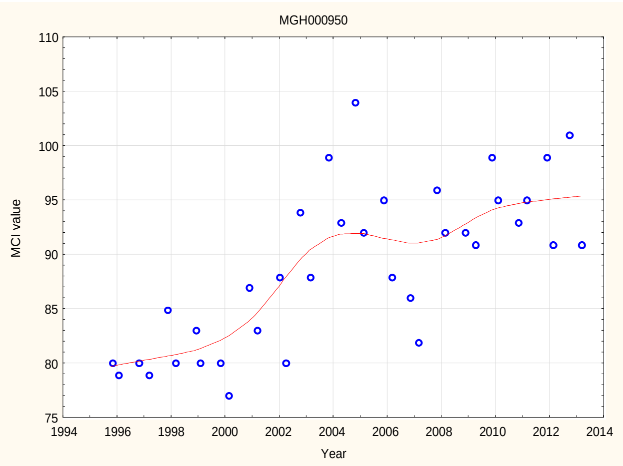
Those taxa recorded as very or extremely abundant during spring and/or summer surveys had been characteristic of this site's communities on 6% to 94% of past survey occasions.

3.2.15.1.3 Predicted stream 'health'

The Mangaehu River site at Raupuha Road, at an altitude of 100 m asl, is in the lower reaches of a river draining an eastern hill country catchment. Relationships for ringplain streams and river developed between MCI and altitude and distance from the National Park (Stark and Fowles, 2009) are therefore not appropriate for this river.

3.2.15.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years (1995-2012) of SEM results collected to date from the site in the Mangaehu River at Raupuha Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 57.



N = 36
Kendall tau = +0.536
p value < 0.0001 [>FDR, p = 0.0001]
Significant at p<0.05 and p<0.01

Figure 57 LOWESS trend plot of MCI data for the Raupuha Road site

A very strong, statistically significant, positive temporal trend in MCI scores (p<0.01 after FDR) was found at this lower river reach, eastern hill country site. This trend is partially explained by an apparent reduction in river bed sedimentation possibly related to fewer severe flood events particularly since 2000 with scores tending to plateau since peaking in 2004 before slightly improving again more recently. The range of LOWESS-smoothed MCI scores (15 units) has also been ecologically significant, particularly over the period since 2000.

Smoothed MCI scores originally bordering on 'poor/fair' generic river health (Table 1) have trended upward into 'fair' where they currently remain (Figure 57).

3.2.15.2 Discussion

Seasonal MCI values typically decreased (by 10 units) between spring and summer at this lower reach site, although by a greater degree than the median five unit difference found to date (Appendix II), with the percentage community composition of 'tolerant' taxa increasing by 15% at the time of the summer survey. However, seasonal communities at this site shared only 16 common taxa (57% of the 28 taxa found at this site in 2012-2013), a moderate percentage of common taxa, partly accounting for the dissimilarity in seasonal MCI values.

3.2.16 Waingongoro River

The results of spring (2012) and summer (2012-2013) surveys are summarised in Table 156 and Table 157, Appendix I.

3.2.16.1 Site near National Park boundary (WGG000115)

3.2.16.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this upper reach site, 700m downstream of the National Park boundary in the Waingongoro River, between 1995 and February 2012. These results are summarised in Table 62, together with the results from the current period, and illustrated in Figure 58.

Table 62 Results of previous surveys performed in the Waingongoro River 700m downstream of the National Park, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGG000115	34	24-40	31	122-139	133	33	130	34	132

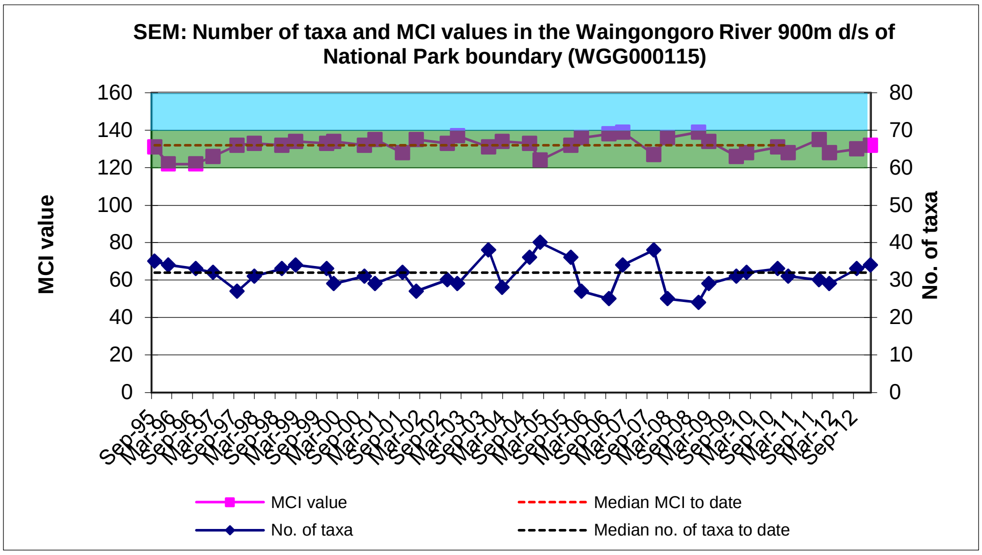


Figure 58 Numbers of taxa and MCI values in the Waingongoro River 700 m d/s National Park

A relatively wide range of richnesses (24 to 40 taxa) has been found with a high median richness of 31 taxa, typical of richnesses in ringplain streams and rivers near the National Park boundary. During the 2012-2013 period spring (33 taxa) and summer (34 taxa) richnesses were two to three taxa above this median richness.

MCI values have had a moderate range (17 units) at this site, more typical of a National Park boundary site. The median value (133 units) also has been typical of upper reach sites elsewhere on the ringplain and the spring, 2012 (130 units) and summer, 2013 (132 units) scores were within 3 units of the historical median. They categorised this site as having 'very good' (spring) and 'very good' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' (spring and summer) health for the upper reaches of a ringplain stream. The historical median score (133 units) placed this site in the 'very good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.16.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 63.

Table 63 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waingongoro River 700 m downstream of the National Park between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
EPHEMEROPTERA	*Austroclima*	7	21	60		
	Coloburiscus	7	35	100	XA	XA
	Deleatidium	8	35	100	XA	XA
	Nesameletus	9	18	51		VA
PLECOPTERA	*Acroperla*	5	3	9		
	Austroperla	9	2	6	A	
	Megaleptoperla	9	32	91	A	A
	Stenoperla	10	4	11		
	Zelandobius	5	3	9		
	Zelandoperla	8	35	100	VA	XA
COLEOPTERA	Elmidae	6	35	100	A	VA
	Hydraenidae	8	24	69	A	VA
MEGALOPTERA	*Archichauliodes*	7	7	20		
TRICHOPTERA	*Aoteapsyche*	4	32	91	A	A
	Beraeoptera	8	27	77	A	A
	Helicopsyche	10	16	46	A	A
	Olinga	9	24	69		A
	Pycnocentrodes	5	1	3		
	Zelolessica	7	11	31		
DIPTERA	*Aphrophila*	5	35	100	VA	VA
	Maoridiamesa	3	2	6		
	Orthocladiinae	2	17	49		

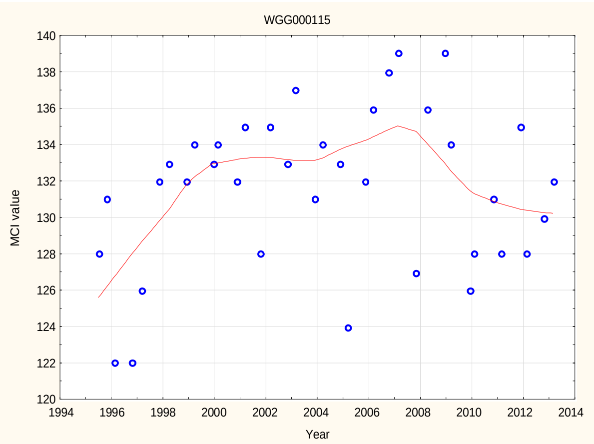
Prior to the current 2012-2013 period, 22 taxa had characterised the community at this site on occasions. These have comprised ten 'highly sensitive', nine 'moderately sensitive', and three 'tolerant' taxa i.e. a high proportion of 'highly sensitive' taxa as might be expected in the upper reaches of a ringplain river near the National Park. Predominant taxa have included seven 'highly sensitive' taxa [mayflies (*Nesameletus* and *Deleatidium*), stoneflies (*Megaleptoperla* and *Zelandoperla*), hydraenid beetles, and cased caddisflies (*Beraeoptera* and *Olinga*)]; four 'moderately sensitive' taxa [mayflies (*Coloburiscus* and *Austroclima*), elmid beetles, and cranefly (*Aphrophila*)]; and only one 'tolerant' taxon [free-living caddisfly (*Aoteapsyche*)]. Five of these taxa have been characteristic of communities on every occasion to date. Eleven of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised seven 'highly sensitive' taxa, three 'moderately sensitive' taxa, and one 'tolerant' taxon, whereas eight 'highly sensitive' taxa, three 'moderately sensitive' taxa, and one 'tolerant' taxon comprised the dominant taxa of the summer, 2013 community. Eight of these thirteen taxa were dominant in both spring and summer communities. All five taxa dominant on every previous survey occasion were included amongst these eight taxa (Table 63). The relatively similar seasonal dominances by high proportions of 'sensitive' taxa were reflected in the very similar seasonal $SQMCI_s$ scores (Tables 156 and 157).

3.2.16.1.3 Predicted stream 'health'

The Waingongoro River site near the National Park is 0.7 km downstream of the National Park boundary at an altitude of 540 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 140 (altitude) and 130 (distance) for this site. The historical site median (133 units) is 7 units lower than the altitude prediction and 3 units higher than the distance predictive value, while the spring, 2012 survey score (130 units) was 10 units lower than the altitude predictive value and the summer, 2013 score (132 units) was also lower, by 8 units, than this predictive value but both were no more than 2 units from the distance predictive value. Of the 36 surveys to date at this site, 28% of MCI scores have been less than 130 units while none have been greater than 140 units.

3.2.16.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waingongoro River near the National Park. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 59.



N = 37
Kendall tau = +0.140
p value = 0.224 [>FDR, p = 0.306]
N/S at $p<0.05$

Figure 59 LOWESS trend plot of MCI data at the site near the National Park

A temporal trend of some improvement in MCI scores has been found over the eighteen year period. This has not been statistically significant at the 5% level however, although previously (prior to 2008) there had been a statistically significant improvement over a shorter period. Most recently there has been some decline but the overall range of LOWESS-smoothed MCI scores remains close to ecologically significant (9 units). Throughout the period, smoothed MCI scores have indicated 'very good' generic river health (Table 1), while in terms of predictive relationships (Table 2) for a site in the upper reaches of a ringplain stream near the boundary of the National Park, river health has remained in the 'expected' category throughout the eighteen year period.

3.2.16.2 Opunake Road site (WGG000150)

3.2.16.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Waingongoro River at this upper mid-reach site at Opunake Road (approximately 7km downstream of the National Park) between 1995 and February 2012. These results are summarised in Table 64, together with the results from the current period, and illustrated in Figure 60.

Table 64 Results of previous surveys performed in the Waingongoro River at Opunake Road together with spring 2012 and summer 2013 results.

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGG000150	34	24-39	28	119-139	129	24	134	25	130

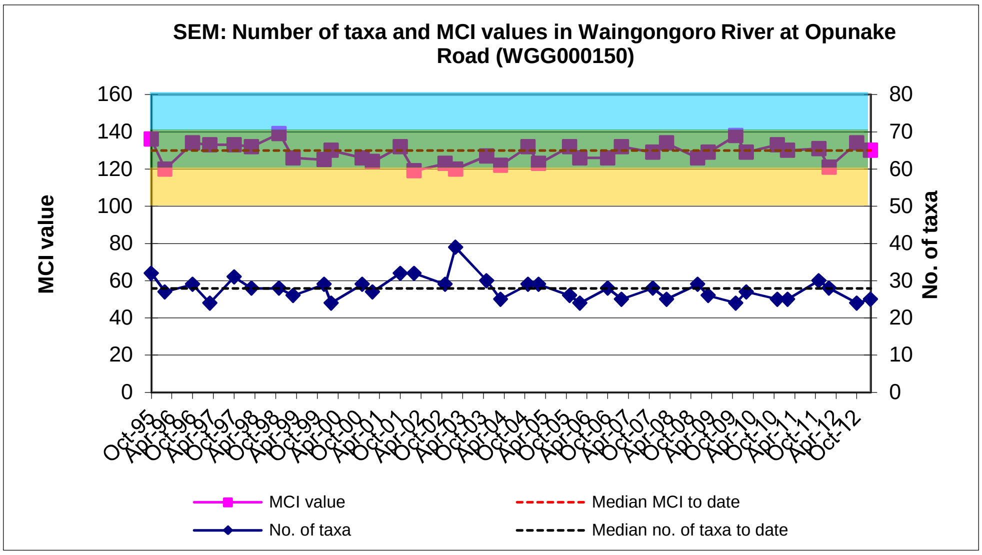


Figure 60 Numbers of taxa and MCI values in the Waingongoro River at Opunake Road

A relatively wide range of richnesses (24 to 39 taxa) has been found; wider than might be expected, with a median richness of 28 taxa (more representative of typical richnesses in the upper mid reaches of ringplain streams and rivers). During the 2012-2013 period spring (24 taxa) and summer (25 taxa) richnesses were very similar and three to four taxa below the median taxa number coincidental with minimal substrate periphyton cover on both occasions.

MCI values have had a moderate range (20 units) at this site, typical of sites in the upper mid reaches of ringplain rivers. The median value (129 units) has been higher than typical of mid reach sites elsewhere on the ringplain however, with the spring, 2012 (134 units) and summer, 2013 (130 units) scores one to five units above the historical median. These scores categorised this site as having 'very good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' health on both occasions for the mid reaches of a ringplain river. The historical median score (129 units) placed this site in the 'very good' and 'better than expected' categories for generic and predictive methods of assessment respectively.

3.2.16.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 65.

Table 65 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waingongoro River at Opunake Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	2	6		
EPHEMEROPTERA	*Austroclima*	7	26	76	A	VA
	Coloburiscus	7	34	100	XA	XA
	Deleatidium	8	34	100	XA	XA
	Nesameletus	9	28	82	A	XA
PLECOPTERA	*Acroperla*	5	1	3		
	Megaleptoperla	9	2	6		
	Zelandoperla	8	24	71	XA	VA
COLEOPTERA	Elmidae	6	34	100	A	VA
	Hydraenidae	8	20	59		A
MEGALOPTERA	*Archichauliodes*	7	22	65	A	VA
TRICHOPTERA	*Aoteapsyche*	4	28	82	A	VA
	Costachorema	7	1	3		
	Hydrobiosis	5	4	12		A
	Beraeoptera	8	27	79	VA	VA
	Confluens	5	2	6		
	Helicopsyche	10	2	6		
	Olinga	9	8	24		A
	Pycnocentrodes	5	14	41		
DIPTERA	*Aphrophila*	5	34	100	A	VA
	Eriopterini	5	1	3		
	Orthocladiinae	2	6	18		

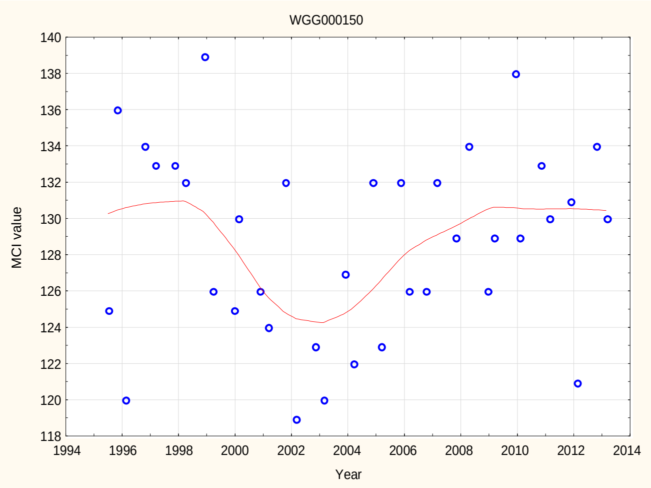
Prior to the current 2012-2013 period, 22 taxa had characterised the community at this site on occasions. These have comprised eight 'highly sensitive', eleven 'moderately sensitive', and three 'tolerant' taxa i.e. a majority of 'sensitive' taxa as would be expected toward the upper mid-reaches of a ringplain stream. Predominant taxa have included five 'highly sensitive' taxa [mayflies (*Deleatidium* on every sampling occasion, and *Nesameletus*), stonefly (*Zelandoperla*), hydraenid beetles, and cased caddisfly (*Beraeoptera*)]; five 'moderately sensitive' taxa [mayflies (*Coloburiscus* and *Austroclima*), elmid beetles, dobsonfly (*Archichauliodes*), and cranefly (*Aphrophila*)]; and one 'tolerant' taxon [net-building caddisfly (*Aoteapsyche*)]. Ten of the characteristics taxa were dominant in the spring, 2012 community. These were comprised of four 'highly sensitive', five 'moderately sensitive', and one 'tolerant' taxa. All of these taxa were again dominant in the summer, 2013 community together with three additional (all 'sensitive') taxa. Two taxa ('highly sensitive' mayfly, *Deleatidium* and 'moderately sensitive' mayfly, *Coloburiscus*) were recorded as extremely abundant in both spring and summer communities. The numerical dominance by similar proportions of 'highly sensitive' and 'moderately sensitive' taxa in both seasons was reflected in the similarity in seasonal $SQMCI_s$ values which differed by only 0.1 unit (Tables 156 and 157). All taxa recorded as very or extremely abundant during spring and/or summer had characterized this site's communities on 65% to 100% of past survey occasions.

3.2.16.2.3 Predicted stream 'health'

The Waingongoro River site at Opunake Road is 7.2 km downstream of the National Park boundary at an altitude of 380 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 123 (altitude) and 110 (distance) for this site. The historical site median (129) is 6 units higher than the altitude prediction and a significant (Stark, 1998) 19 units higher than the distance predictive value while the spring, 2012 survey score (134 units) was significantly (Stark, 1998) higher than both predictive values. The summer, 2013 score (130 units) was also higher than both the altitude and the distance predictive values by 7 to a significant 11 units. Of the 36 surveys to date at this site, no MCI scores have been less than 110 units while 81% have been greater than 123 units, further indicative of the better than predicted health of the river at this site.

3.2.16.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waingongoro River at Opunake Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 61.



N = 37
Kendall tau =+ 0.042
p value = 0.717 [>FDR, p = 0.731]
N/S at p <0.05

Figure 61 LOWESS trend plot of MCI data at the Opunake Road site

An overall temporal trend of minimal change (very slight increase) in MCI scores has not been statistically significant at this site in the upper mid-reaches of the river (some 7 km below the National Park). The LOWESS-smoothed range of scores (7 units) has also been ecologically insignificant over the eighteen year period. Localised erosion had caused sediment deposition on the riverbed during 1999 with a subsequent five year decline in MCI scores which was of no ecological significance (LOWESS-smoothed range of 7 units). This decline ceased with a gradual improvement in MCI scores towards earlier levels over the latter ten years. The erosion event was very localised and site specific, as corresponding biological and physiochemical monitoring data showed no significant trends at the nearest downstream site (Eltham Road). Smoothed MCI scores have been consistently indicative of 'very good' generic river health (Table 1) although trending downward toward 'good' immediately following the erosion event. In terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain river, health has been in the 'better than expected' category for almost the entire period, the exception being over the 2001 to 2005 period when health fell into the 'expected' category.

3.2.16.3 Eltham Road site (WGG000500)

3.2.16.3.1 Taxa richness and MCI

Thirty-eight surveys have been undertaken in the Waingongoro River at this mid-reach site at Eltham Road between October 1995 and February 2012. These results are summarised in Table 66, together with the results from the current period, and illustrated in Figure 62.

Table 66 Results of previous surveys performed in the Waingongoro River at Eltham Road, together with spring 2012 and summer 2013 results.

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGG000500	38	16 - 32	23	91-115	101	19	124	26	112

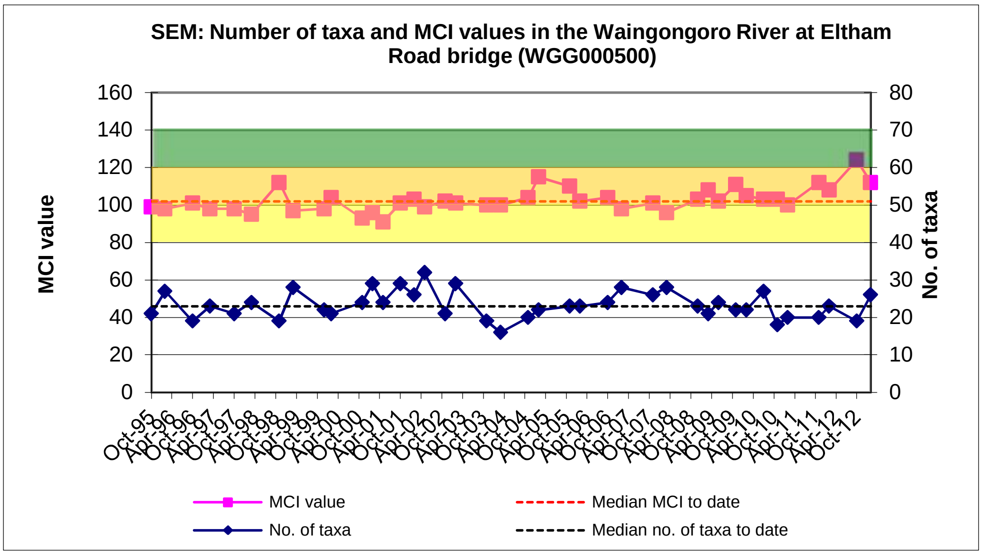


Figure 62 Numbers of taxa and MCI values in the Waingongoro River at Eltham Road

A wide range of richnesses (16 to 32 taxa) has been found with a median richness of 23 taxa, typical of richnesses in the mid reaches of ringplain streams and rivers. During the 2012-2013 period spring (19 taxa) and summer (26 taxa) richnesses were dissimilar but were within 4 taxa of the median taxa number.

MCI values have had a moderate range (24 units) at this site, typical of sites in the mid reaches of ringplain rivers. The median value (101 units) has been relatively typical of mid reach sites elsewhere on the ringplain with the spring, 2012 (124 units) and summer, 2013 (112 units) scores above those typical for such a site and significantly from 11 to 23 units higher than the historical median. The spring score was 9 units higher than the historical maximum at this site (coincident with minimal periphyton substrate cover). These scores categorised this site as having 'very good' (spring) and 'good' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring and summer) health for the mid reaches of a ringplain river. The historical median score (101 units) placed this site in the 'good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.16.3.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 67.

Table 67 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waingongoro River at Eltham Road between 1995 and February 2012 [38 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	4	11		
ANNELIDA	Oligochaeta	1	12	32		
MOLLUSCA	*Potamopyrgus*	4	7	18		
EPHEMEROPTERA	*Austroclima*	7	9	24		A
	Coloburiscus	7	21	55	VA	VA
	Deleatidium	8	27	71	XA	XA
PLECOPTERA	*Zelandobius*	5	5	13	A	
COLEOPTERA	Elmidae	6	36	95	A	VA
	Hydraenidae	8	0	0		A
MEGALOPTERA	*Archichauliodes*	7	22	58		A
TRICHOPTERA	*Aoteapsyche*	4	32	84		XA
	Costachorema	7	16	42		
	Hydrobiosis	5	24	63		A
	Beraeoptera	8	1	3		
	Oxyethira	2	2	5		
	Pycnocentrodes	5	10	26	A	
DIPTERA	*Aphrophila*	5	8	21		A
	Eriopterini	5	6	16		
	Maoridiamesa	3	17	45		
	Orthocladiinae	2	23	61		
	Tanytarsini	3	9	24		
	Ceratopogonidae	3	1	3		
	Empididae	3	3	8		
	Austrosimulium	3	13	34		

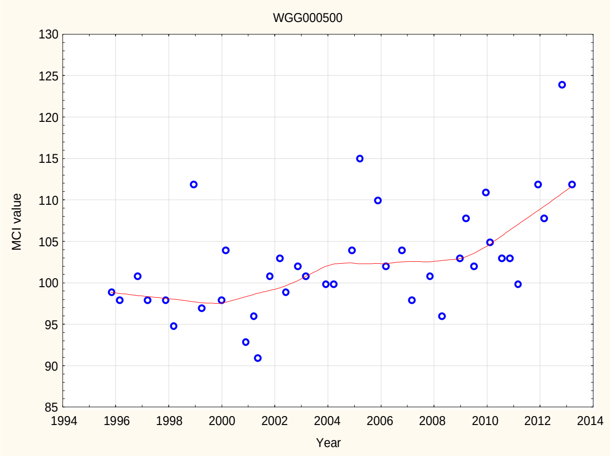
Prior to the current 2012-2013 period, 23 taxa had characterised the community at this site on occasions. These have comprised two 'highly sensitive', ten 'moderately sensitive', and eleven 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and a downstream increase in 'tolerant' taxa as would be expected in the mid reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; four 'moderately sensitive' taxa [mayfly (*Coloburiscus*), elmid beetles, free-living caddisfly (*Hydrobiosis*), and dobsonfly (*Archichauliodes*)]; and two 'tolerant' taxa [free-living caddisfly (*Aoteapsyche*) and orthoclad midges]. Five of these historically characteristic taxa were dominant in the spring, 2012 community. These comprised one 'highly sensitive' taxon, four 'moderately sensitive' taxa, but no 'tolerant' taxa, whereas three of these taxa and one additional 'highly sensitive', four 'moderately sensitive', and one 'tolerant' taxa comprised the dominant taxa of the summer community. The additional 'highly sensitive' taxon (hydraenid beetles) had not been recorded previously in abundance at this site. Three of these eleven taxa were dominant in both spring and summer communities (Table 67). The increased seasonal numerical dominance within the single 'tolerant' taxon resulted in a decrease (1.4 units) in $SQMCI_s$ scores between spring and summer (Tables 156 and 157). The four taxa recorded as very or extremely abundant during spring and/or summer have characterised this site's communities on 55% to 95% of past survey occasions.

3.2.16.3.3 Predicted stream 'health'

The Waingongoro River site at Eltham Road is 23.0 km downstream of the National Park boundary at an altitude of 200 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 105 (altitude) and 97 (distance) for this site. The historical site median (101) is 4 units lower than the altitude prediction and 4 units higher than the distance predictive value while the spring, 2012 survey score (124 units) was significantly 19 to 27 units above predictive values and the summer, 2013 score (111 units) was 6 units above the predictive altitude value and a significant 14 units above the predictive distance value. Of the 40 surveys to date at this site, 13% of MCI scores have been less than 97 units while 23% have been greater than 105 units.

3.2.16.3.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waingongoro River at Eltham Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 63.



N = 40
Kendall tau = +0.419
p value = 0.0001 [>FDR, p < 0.001]
Significant at p < 0.05 and p < 0.01; and after FDR application

Figure 63 LOWESS trend plot of MCI data at the Eltham Road site

A positive temporal trend in MCI scores has been found over the eighteen-year period which has been statistically significant at the 5% and 1% levels and after FDR application. This has been more pronounced since 2001 but scores plateaued for about three years before a more recent further improvement. The narrow range of LOWESS-smoothed range of scores (14 units) has been of ecological significance over the eighteen year period although particularly influenced by very recent scores. MCI scores consistently bordered on 'fair' to 'good' generic river health (Table 1) remaining 'good' since 2003. In terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain river, health has been consistently in the 'expected' category since 2003, prior to which it bordered on the 'worse than expected' category

3.2.16.4 Stuart Road site (WGG000665)

3.2.16.4.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Waingongoro River at this mid-reach site at Stuart Road between 1995 and February, 2012. These results are summarised in Table 68, together with the results from the current period, and illustrated in Figure 64.

Table 68 Results of previous surveys performed in the Waingongoro River at Stuart Road, together with spring 2012 and summer 2013 results.

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGG000665	34	14-30	20	77-111	94	20	108	25	105

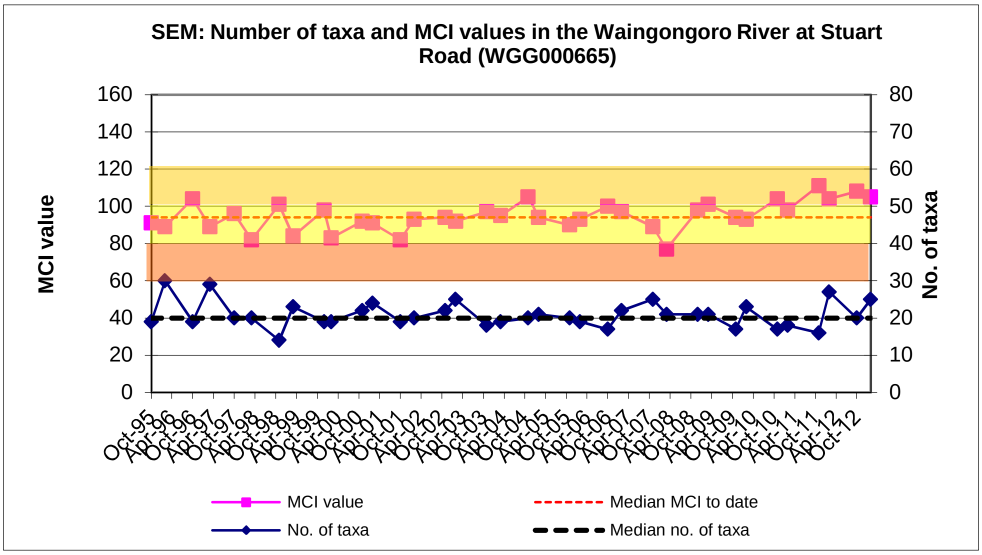


Figure 64 Numbers of taxa and MCI values in the Waingongoro River at Stuart Road

A wide range of richnesses (14 to 30 taxa) has been found with a median richness of 20 taxa (more representative of typical richnesses in the mid reaches of ringplain streams and rivers). During the 2012-2013 period spring (20 taxa) and summer (25 taxa) richnesses varied from equal with the median taxa number in spring to well above median taxa number in summer, coincidental with a typically more widespread substrate periphyton cover in summer.

MCI values have had a moderately wide range (34 units) at this site, typical of sites in the mid reaches of ringplain rivers. The median value (94 units) has been lower than typical of mid reach sites elsewhere on the ringplain however, with the spring, 2012 (108 units) and summer, 2013 (105 units) scores higher than typical of this site and significantly above the historical median by 14 units in spring and 11 units in summer. This spring score was only 3 units below the maximum recorded at this site (in the previous spring), coincident with minimal substrate periphyton cover. These scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' (spring and summer) health for the mid reaches of a ringplain river. Improvements in biological 'health', as indicated by the higher than median scores have been coincidental with the July 2010 diversion of the major point source discharge (Eltham municipal wastewater) out of the catchment, a short distance upstream of this site. The historical median score (94 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.16.4.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 69.

Table 69 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waingongoro River at Stuart Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	3		
ANNELIDA	Oligochaeta	1	18	53		
CRUSTACEA	Ostracoda	1	1	3		
EPHEMEROPTERA	*Austroclima*	7	5	15		A
	Coloburiscus	7	1	3	A	A
	Deleatidium	8	18	53	XA	XA
PLECOPTERA	*Zelandobius*	5	2	6	A	
COLEOPTERA	Elmidae	6	27	79	A	VA
MEGALOPTERA	*Archichauliodes*	7	0			A
TRICHOPTERA	*Aoteapsyche*	4	27	79		VA
	Costachorema	7	6	18		
	Hydrobiosis	5	12	35		A
	Beraeoptera	8	0		A	
	Oxyethira	2	1	3		
	Pycnocentrodes	5	6	18	VA	
DIPTERA	*Aphrophila*	5	13	38		A
	Maoridiamesa	3	24	71		A
	Orthocladiinae	2	33	97		A
	Tanytarsini	3	8	24		A
	Ceratopogonidae	3	1	3		
	Empididae	3	2	6		
	Austrosimulium	3	11	32		

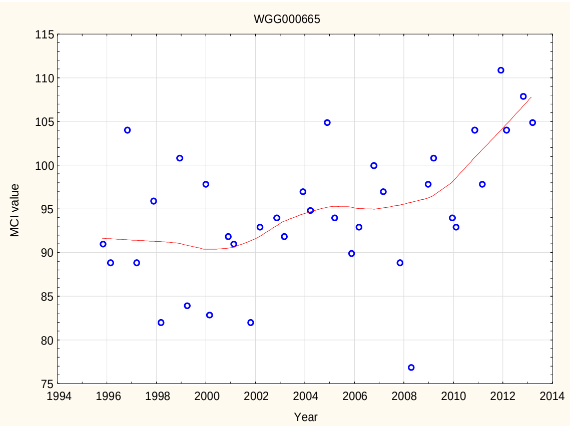
Prior to the current 2012-2013 period, 20 taxa had characterised the community at this site on occasions. These have comprised one 'highly sensitive', eight 'moderately sensitive', and eleven 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and a higher proportion of 'tolerant' taxa as might be expected in the mid reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; one 'moderately sensitive' taxon [elmid beetles]; and four 'tolerant' taxa [oligochaete worms, free-living caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]. Five of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised one 'highly sensitive' taxon and four 'moderately sensitive' taxa, together with one 'highly sensitive' taxon [cased caddisfly (*Beraeoptera*)], which previously had not been recorded in abundance at this site. Three of these taxa plus three additional 'moderately sensitive', four 'tolerant', and one 'moderately sensitive' (not previously characteristic at this site) taxa comprised the dominant taxa of the summer community. Only three of these 14 taxa were dominant in both spring and summer communities (Table 69). For the second occasion to date, the 'tolerant' orthoclad midges were not characteristic of this site's spring community, coincident with a marked reduction in spring periphyton substrate cover. A decreased numerical dominance within one 'moderately sensitive' taxon (stony-cased caddisfly) and increased abundance of one 'tolerant' taxon were reflected in the small decrease (0.6 unit) in summer $SQMCI_s$ score (Tables 156 and 157). The taxa (four) recorded as very or extremely abundant during spring and/or summer have characterised this site's communities on 18% to 79% of past survey occasions.

3.2.16.4.3 Predicted stream 'health'

The Waingongoro River site at Stuart Road is 29.6 km downstream of the National Park boundary at an altitude of 180 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 103 (altitude) and 94 (distance) for this site. The historical site median (94) is 9 units lower than the altitude prediction and equal with the distance predictive value. The spring, 2012 survey score (108 units) was 5 to a significant (Stark, 1998) 14 units higher than these predictive values and the summer, 2013 score (105 units) was 2 to a significant 11 units above both predictive values. Of the 36 surveys to date at this site, 47% of MCI scores have been less than 94 units while only 19% have been greater than 103 units.

3.2.16.4.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waingongoro River at Stuart Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 65.



N = 36
Kendall tau = +0.363
p value = 0.002 [>FDR, p = 0.005]
Significant at p <0.05 and p < 0.01 and after FDR application

Figure 65 LOWESS trend plot of MCI data at the Stuart Road site

A positive statistically significant trend in MCI scores has been found (at the 5% and 1% levels and after FDR application) over the period with a gradual improvement in MCI scores since 2002 (coincident with summer diversion of the treated meatworks wastes discharge (at Eltham) from the river to land irrigation) and particularly most recently (since 2009) following the diversion of treated municipal Eltham wastewater out of the catchment (to the Hawera WWTP and ocean outfall). The LOWESS-smoothed range of scores (17 units) has also been ecologically significant over the eighteen year period. Smoothed MCI scores consistently have been indicative of 'fair' generic river health until more recently when they have been more indicative of 'good' generic health (Table 1). In terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain river, health has been in the 'expected' category almost throughout the period until approaching the 'better than expected' category in the last two years.

3.2.16.5 SH45 site (WGG000895)

3.2.16.5.1 Taxa richness and MCI

Thirty-five surveys have been undertaken in the Waingongoro River at this lower reach site at SH45 between 1995 and February, 2012. These results are summarised in Table 70, together with the results from the current period, and illustrated in Figure 66.

Table 70 Results of previous surveys performed in the Waingongoro River at together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGG000895	35	16-24	20	73-105	95	16	106	23	96

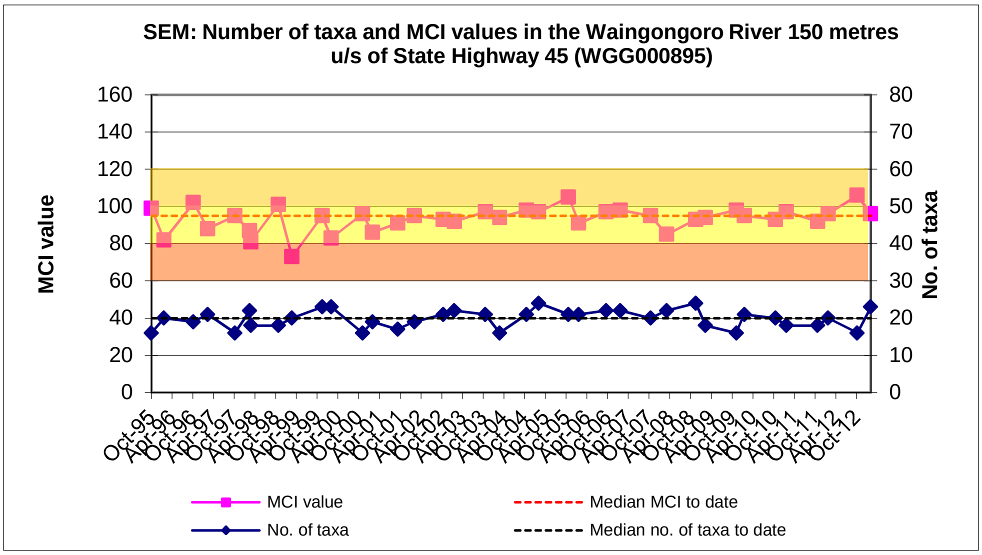


Figure 66 Numbers of taxa and MCI values in the Waingongoro River 150 m u/s of SH45

A moderate range of richnesses (16 to 24 taxa) has been found with a median richness of 20 taxa (more representative of typical richnesses in the lower reaches of ringplain streams and rivers). During the 2012-2013 period spring (16 taxa) and summer (23 taxa) richnesses showed a moderate summer increase coincident with greater substrate periphyton cover. The spring richness was lower than the median taxa number by 4 taxa whereas summer richness was higher by 3 taxa.

MCI values have had a wide range (32 units) at this site, more typical of sites in the lower reaches of ringplain streams and rivers. The median value (95 units) has been slightly higher than typical of scores at lower reach sites elsewhere on the ringplain (TRC, 1999 (updated, 2012)), however. The spring, 2012 (106 units) score was significantly above the median score and higher than the previous maximum score for this site, whereas the summer, 2013 (96 units) score was typical of scores at this site and one unit above the historical median. These scores categorised this site as having 'good' health (spring) and 'fair' health (summer) generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' health (spring) and 'expected' health (summer) for the lower reaches of a ringplain river. The historical median score (95 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.16.5.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 71.

Table 71 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waingongoro River at SH45 between 1995 and February 2012 [35 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	9		
ANNELIDA	Oligochaeta	1	29	83	VA	
	Lumbricidae	5	4	11		
MOLLUSCA	*Latia*	5	2	6		
	Potamopyrgus	4	33	94		VA
EPHEMEROPTERA	*Austroclima*	7	3	9		VA
	Deleatidium	8	18	51	XA	VA
PLECOPTERA	*Zelandobius*	5	3	9		
COLEOPTERA	Elmidae	6	33	94		
MEGALOPTERA	*Archichauliodes*	7	3	9		A
TRICHOPTERA	*Aoteapsyche*	4	35	100	A	VA
	Costachorema	7	2	6		
	Hydrobiosis	5	17	49		
	Pycnocentrodes	5	33	94	VA	VA
DIPTERA	*Aphrophila*	5	10	29		
	Maoridiamesa	3	17	49		
	Orthocladiinae	2	20	57		
	Tanytarsini	3	5	14		
	Austrosimulium	3	5	14		

Prior to the current 2012-2013 period, 19 taxa had characterised the community at this site on occasions. These have comprised one 'highly sensitive', ten 'moderately sensitive', and eight 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and relatively high proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; two 'moderately sensitive' taxa [elmid beetles and caddisfly (*Pycnocentrodes*)]; and four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and midge (orthoclads)]. Four of the historically characteristic and predominant taxa were dominant in the spring, 2012 community. These comprised one 'highly sensitive', one 'moderately sensitive', and two 'tolerant' taxa. Three of these taxa and two 'moderately sensitive' and one 'tolerant' taxa also comprised the dominant taxa of the summer, 2013 community but with a tendency for increased numerical dominance within some of the 'tolerant' taxa and a decreased abundance of the 'highly sensitive' taxon (Table 71). These subtle differences in seasonal dominances were reflected in the decrease of 0.5 unit in seasonal $SQMCI_s$ scores (Tables 156 and 157).

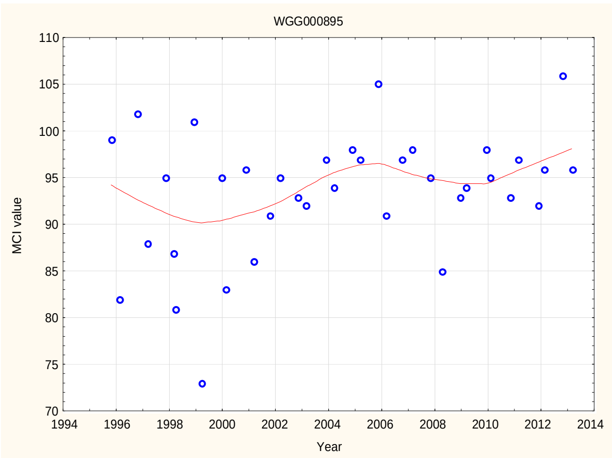
The six taxa recorded as very/extremely abundant during spring and/or summer have characterised this site's communities on 9% to 100% of past survey occasions.

3.2.16.5.3 Predicted stream 'health'

The Waingongoro River site at SH45 is 63.0 km downstream of the National Park boundary at an altitude of 40 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 89 (altitude) and 85 (distance) for this site. The historical site median (95) is 6 units higher than the altitude prediction and ten units higher than the predictive distance value. The spring, 2012 survey score (106 units) was significantly (Stark, 1998) 17 units higher than the altitude predictive value and 21 units higher than the predictive distance value while the summer, 2013 score (96 units) was 7 units higher than the predictive altitude value and a significant (Stark, 1998) 11 units above the predicted distance value. Of the 37 surveys to date at this site, 11% of MCI scores have been less than 85 units while 78% have been greater than 89 units.

3.2.16.5.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waingongoro River at SH45. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 67.



N = 37
Kendall tau = +0.190
p value = 0.098 [>FDR, p = 0.164]
N/S at $p < 0.05$

Figure 67 LOWESS trend plot of MCI data for the SH45 site

A positive trend in MCI scores has been found over the eighteen year period, particularly since 2000 followed by a plateauing in trend in 2005- 2006 followed by further improvement, but this has not been statistically significant. The LOWESS-smoothed range (8 units) of scores has not been ecologically significant. LOWESS-smoothed MCI scores have consistently indicated 'fair' generic river health (Table 1) throughout the period. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain river, river health has remained in the 'expected' category throughout the period, although health has more recently bordered on the 'better than expected' category.

3.2.16.6 Ohawe Beach site (WGG000995)

3.2.16.6.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Waingongoro River at this lower reach site at Ohawe Beach between 1995 and February 2012. These results are summarised in Table 72, together with the results from the current period, and illustrated in Figure 68.

Table 72 Results of previous surveys performed in the Waingongoro River at the Ohawe Beach site, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WGG000995	34	12-25	18	69-99	90	18	100	22	93

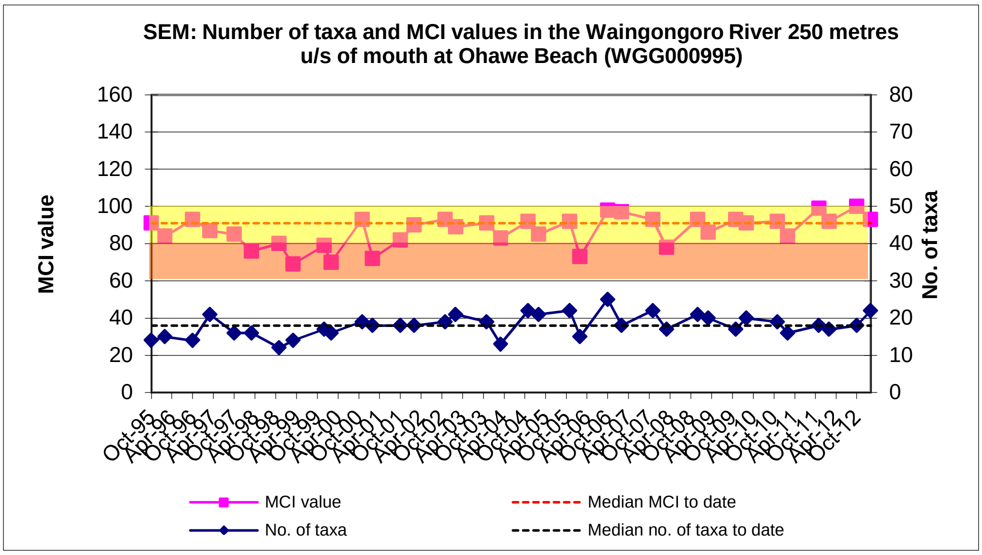


Figure 68 Numbers of taxa and MCI values in the Waingongoro River at the Ohawe Beach site

A wide range of richnesses (12 to 25 taxa) has been found, with a median richness of 18 taxa. During the 2012-2013 period spring (18 taxa) and summer (22 taxa) richnesses were relatively similar and equal with the median richness in spring and four taxa higher in summer when periphyton substrate cover was more extensive.

MCI values have had a moderate range (30 units) at this site, typical of sites in the lower reaches of ringplain streams and rivers. The median value (90 units) has been relatively typical of scores at lower reach sites elsewhere on the ringplain (TRC, 1999 (updated, 2012)). The spring, 2012 (100 units) and summer, 2013 (93 units) scores, were slightly above scores typical for such a site and 3 to 10 units above the historical median, but showed a typical summer seasonal decrease. These scores categorised this site as having 'good' and 'fair' health generically (Table 1) in spring and summer respectively and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower reaches of a ringplain river. The historical median score (90 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.16.6.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 73.

Table 73 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waingongoro River at the Ohawe Beach site between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	25	74		VA
	Lumbricidae	5	1	3		
MOLLUSCA	*Potamopyrgus*	4	26	76	A	VA
CRUSTACEA	*Paratya*	3	2	6		
EPHEMEROPTERA	*Austroclima*	7	2	6		
	Deleatidium	8	7	21	VA	
COLEOPTERA	Elmidae	6	22	65		
TRICHOPTERA	*Aoteapsyche*	4	33	97	VA	XA
	Costachorema	7	1	3		
	Hydrobiosis	5	3	9		
	Pycnocentrodes	5	27	79	XA	VA
DIPTERA	*Aphrophila*	5	5	15	A	A
	Maoridiamesa	3	26	76	VA	
	Orthocladiinae	2	32	94	XA	VA
	Tanytarsini	3	5	15		A
	Ephydridae	4	2	6		
	Austrosimulium	3	4	12		

Prior to the current 2012-2013 period, 17 taxa had characterised the community at this site on occasions. These have comprised one ‘highly sensitive’, seven ‘moderately sensitive’, and nine ‘tolerant’ taxa i.e. a lower proportion of ‘sensitive’ taxa and a higher proportion of ‘tolerant’ taxa as would be expected in the lower reaches of a ringplain river. Predominant taxa have included no ‘highly sensitive’ taxa; two ‘moderately sensitive’ taxa [elmid beetles and stony-cased caddisfly (*Pycnocentrodes*)]; and five ‘tolerant’ taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]. Seven of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised one ‘highly sensitive’ taxon, two ‘moderately sensitive’ taxa, and four ‘tolerant’ taxa, whereas five of these same taxa plus two additional ‘tolerant’ taxa comprised the dominant taxa of the summer, 2013 community. Although five of these nine taxa were dominant in both spring and summer communities (Table 73), some overall increase in numerical abundances within some ‘tolerant’ taxa and decreased abundances within two ‘sensitive’ taxa combined to reduce the summer $SQMCI_s$ score by 0.3 unit (Tables 156 and 157).

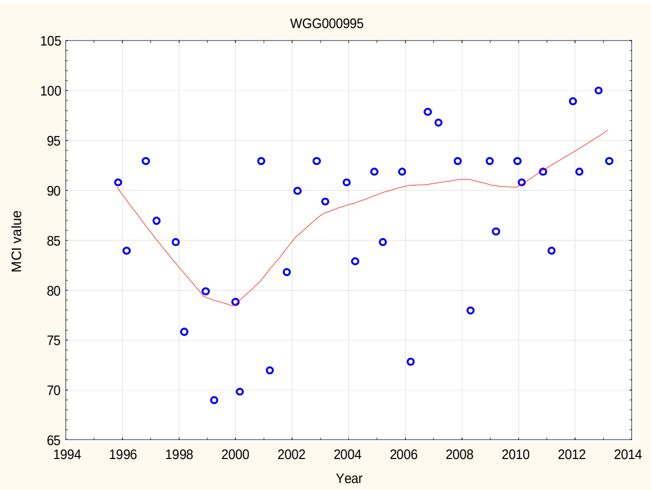
The seven taxa recorded as very or extremely abundant during spring and/or summer have characterised this site’s communities on 21% to 97% of past survey occasions.

3.2.16.6.3 Predicted stream 'health'

The Waingongoro River at the Ohawe Beach site is 66.6km downstream of the National Park boundary at an altitude of 5 m asl. Relationships for ringplain streams and rivers developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 85 (altitude) and 85 (distance) for this site. The historical site median (90) is 5 units higher than both the predictive values. The spring, 2012 survey score (106 units) was a significant 21 units higher than both predictive values while the summer score (93 units) was eight units higher than the predictive altitude and distance values. Of the 36 surveys to date at this site, 33% of MCI scores have been less than 85 units while 61% have been greater than 85 units.

3.2.16.6.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Waingongoro River at Ohawe Beach. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 69.



N = 36
Kendall tau = +0.329
p value = 0.005 [>FDR, p = 0.011]
Significant at p <0.05 and p < 0.01;
not significant after FDR (at p < 0.01)

Figure 69 LOWESS trend plot of MCI data at the Ohawe Beach site

There has been a marked trend of MCI scores improvement since 2001, which tended to plateau between 2006 and 2009, with a more recent improvement resulting in an overall eighteen-year temporal trend which has been statistically significant ($p > 0.05$ after FDR application) but not at $p < 0.01$ after FDR application. The range of LOWESS-smoothed scores (17 units) has been ecologically significant, mainly due to the influence of a series of low scores (<81 MCI units) between 1998 and 2001 and the elevation in scores subsequent to diversion of major mid-catchment point source discharges out of the river, particularly since 2009.

Smoothed MCI scores were consistently indicative of 'fair' generic river health (Table 1) with the exception of the 1998 to 2001 period when generic health fell to 'poor'. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, river health has remained in the 'expected' category until most recently when it has improved to border on the 'better than expected' category.

3.2.16.7 Discussion

Seasonal MCI values typically decreased between spring and summer at five of the six sites by 4, 12, 3, 10, and 7 units in a downstream direction with the exception of the uppermost site where there was a minimal increase (2 units). These decreases tended to be slightly higher than historical seasonal median differences (by 0 to 12 units) at the corresponding sites (Appendix II). Seasonal communities shared 63% of the 41 taxa found at the upper site near the National Park, 69% of 29 taxa at the Opunake Road upper mid-reach site, 55% of 29 taxa at the Eltham Road mid-reach site, 67% of 27 taxa at the Stuart Road mid-reach site, 56% of 25 taxa at the SH45 lower reach site, and 48% of 27 taxa at the furthest downstream site (Ohawe Beach) in the lower reaches. Seasonal community compositions in the 2012-2013 period therefore tended to follow typical trends of generally greater dissimilarity with increasing distance downstream from the National Park.

Community composition varied markedly through the length of the river surveyed. A total of 45 taxa was recorded in spring of which only eight taxa were present at all six sites. These included one 'highly sensitive' taxon, six 'moderately sensitive' taxa, and one 'tolerant' taxon with only the 'highly sensitive' mayfly (*Deleatidium*) abundant at all six sites. A higher total of 52 taxa was found along the river's length by the summer survey of which nine taxa were present at all six sites. These were similar to the eight widespread taxa found in spring with two additional 'tolerant' taxa and one fewer 'moderately sensitive' taxon. Only one 'tolerant' caddisfly was abundant at all six sites in summer. Dissimilarities in spatial community structure along the length of the Waingongoro River were slightly more pronounced in summer than in spring.

The MCI scores fell in a downstream direction between the upper site and the furthest downstream lower reaches site by 30 units in spring and 39 units in summer, over a river distance of 65.9 km. These seasonal falls in MCI scores equated to rates of decline of 0.5 unit/km (spring) and 0.6 unit/km (summer), compared with a predicted rate of 0.7 unit/km for the equivalent length and reach of a National Park-sourced river (Stark and Fowles, 2009). This was typical of most past trends, when there have been increased summers' seasonal rates of decline. These relatively low rates of decline (for a ringplain stream) may be explained in part by the extensive meander pattern unique to this catchment which has a total river length of 76 km from its source to the coast (67 km outside the National Park) compared with the average ringplain stream length of approximately 25 km.

Between the upper and mid-reach site at Eltham Road, the spring (0.3 unit/km) and summer (0.9 unit/km) rates of decline were lower than the predicted rate (1.5 units/km) for the equivalent river reach. For the mid-reach Eltham Road to Ohawe Beach lower reach site, spring (0.55 unit/km) and summer (0.4 unit/km) rates of decline were slightly higher than the predicted rate of 0.3 unit/km. Previously, more marked rates of decline had been recorded between the Eltham Road and Stuart Road mid-reach sites (6.6 km reach) in spring and summer compared with the predicted rate (0.5 units/km) for the equivalent reach of this river. This had been attributable to point source discharges of treated Eltham municipal wastes and treated industrial (meatworks) wastes within this reach but since the summer removal of the meatworks discharge and the complete diversion of the municipal wastes (post July 2010) these rates have reduced. The rate in spring 2012 (2.4 units/km) and in summer 2013 (1.1 units/km) reflected these seasonal differences.

Using the long-term median SEM MCI scores for each site (Appendix II), the rates of decline between upper and mid catchment (Eltham Road) and mid catchment (Eltham Road) and lower river sites have been about 1.3 and 0.25 units per km respectively with an overall average rate of decline of 0.6 MCI unit/km over the river's length. Therefore rates of decline over the 2012-2013 period were lower in spring and summer for the upper to mid reach and higher for the mid to lower reach of the river than have been typical of average rates prior to 2012.

3.2.17 Mangawhero Stream

The results found by the 2012-2013 surveys are presented in Table 158 and Table 159 Appendix I for this small stream draining the Ngaere swamp, with a lower sub-catchment (Mangawharawhara Stream) rising on the ringplain but outside of the National Park.

3.2.17.1 Site upstream of the Eltham Municipal WWTP discharge (MWH000380)

3.2.17.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in this mid-reach site in the Mangawhero Stream within about 3 km of the Ngaere swamp 1995 and February 2012. These results are summarised in Table 74, together with the results from the current period, and illustrated in Figure 70.

Table 74 Results of previous surveys performed in Mangawhero Stream upstream of Eltham WWTP, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MWH000380	34	10-23	15	58-85	74	12	85	24	85

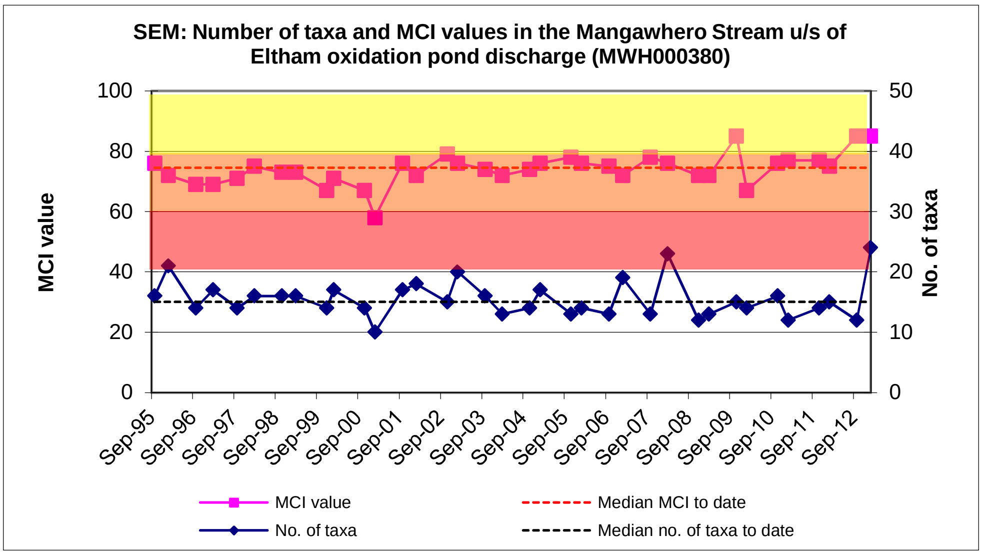


Figure 70 Numbers of taxa and MCI values in the Mangawhero Stream upstream of Eltham WWTP

A moderate range of richnesses (10 to 23 taxa) has been found, with a median richness of 15 taxa (more representative of typical richnesses in small swamp drainage streams where a median richness of 18 taxa has been recorded from 167 previous surveys of 'control' sites at similar altitudes (TRC, 1999 (updated, 2012)). During the 2012-2013 period, spring (12 taxa) richness was three taxa lower than this site's median richness while the markedly richer summer community (24 taxa) was 9 taxa higher than the historical median and one taxon more than the previous maximum. (However it is noted that rarities (fewer than 5 individuals per taxon) comprised 16% of the spring community compared with 68% of the summer community). The habitat was predominantly comprised of a hard clay substrate with patchy filamentous algae and thin algal mats substrate cover in both spring and summer.

MCI values have had a moderate range (27 units) at this site. The median value (74 units) has been typical of similar non-ringplain sites elsewhere in the region however, and the identical spring, 2012 (85 units) and summer, 2013 (85 units) scores were toward the maximum of the range for such a site. The scores were significantly (Stark, 1998) higher (by eleven units) in spring/summer than the historical median. These scores categorised this site as having 'fair' (spring and summer) health generically (Table 1) and were 7 units higher than the median MCI score (78 units) recorded by 167 previous surveys of similar 'control' sites in small, non ringplain streams in Taranaki (TRC, 1999 (updated, 2012)). The historical median score (74 units) placed this site in the 'poor' category for the generic method of assessment and was 4 units below the median score recorded at similar sites elsewhere in the region.

3.2.17.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 75.

Table 75 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Mangawhero Stream upstream of Eltham WWTP between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	3		
ANNELIDA	Oligochaeta	1	22	65	VA	
	Lumbricidae	5	1	3		
MOLLUSCA	*Potamopyrgus*	4	2	6		
CRUSTACEA	Ostracoda	1	8	24		VA
	Paracalliope	5	29	85		VA
EPHEMEROPTERA	*Austroclima*	7	31	91	VA	A
TRICHOPTERA	*Aoteapsyche*	4	14	41	A	A
	Hydrobiosis	5	6	18		
	Polyplectropus	6	1	3		
	Oxyethira	2	4	12		
DIPTERA	*Aphrophila*	5	16	47	A	A
	Chironomus	1	2	6		
	Maoridiamesa	3	8	24		
	Orthocladiinae	2	33	97	A	A
	Austrosimulium	3	15	44		A

Prior to the current 2012-2013 period, 16 taxa had characterised the community at this site on occasions. These have comprised six 'moderately sensitive' and ten 'tolerant' taxa i.e. an absence of 'highly sensitive' taxa and a relatively high proportion of 'tolerant' taxa as would be expected in the drain-like upper reaches of a non-ringplain, swampy, seepage stream.

Predominant taxa have included two 'moderately sensitive' taxa [amphipod (*Paracalliope*) and mayfly (*Austroclima*)]; and two 'tolerant' taxa [oligochaete worms and orthoclad midges].

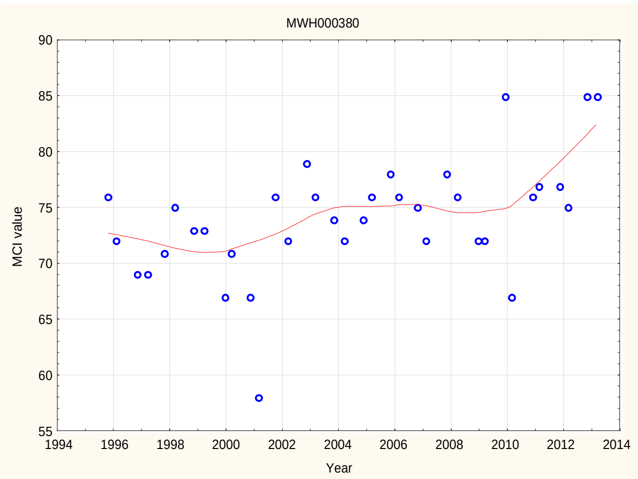
Six of the historically characteristic taxa were dominant in the spring, 2012 community and comprised three of the predominant taxa (above). The summer, 2013 community was characterised by four of the taxa dominant in spring plus one 'moderately sensitive' and two 'tolerant' taxa, three of which previously had been predominantly characteristic of this site's communities (Table 75). A small increase in abundance of one 'tolerant' summer taxon and some decrease in abundance of the 'moderately sensitive' mayfly resulted in a minor decrease in $SQMCI_s$ scores (0.5 unit) recorded between seasons (Tables 158 and 159). The four taxa which were recorded as very abundant during spring and/or summer had characterised this site's communities on 24% to 91% of past surveys.

3.2.17.1.3 Predicted stream 'health'

The Mangawhero Stream rises as seepage from the Ngaere swamp and is not a ringplain stream at the site upstream of the Eltham WWTP. This site is at an altitude of 200 m asl and toward its upper reaches. Relationships for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), therefore are not applicable in the upper reaches of this type of stream.

3.2.17.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Mangawhero Stream upstream of the Eltham WWTP discharge. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 71.



N = 36
Kendall tau = +0.369
p value = 0.002 [>FDR, p = 0.004]
Significant at p <0.05 and p < 0.01 levels; and after FDR application

Figure 71 LOWESS trend plot of MCI data at site upstream of the Eltham WWTP discharge

A positive and statistically significant temporal trend in MCI scores (p < 0.01 after FDR) has been found over the eighteen year monitoring period at this site with the early trend of slightly increasing scores having been followed by a plateauing of scores a few units above those recorded early in the programme and a more recent steady increase. However, the narrow range of LOWESS-smoothed scores (5 units) until 2012-2013 has not been of ecological significance but the range has widened to 10 units very recently. LOWESS-smoothed MCI scores consistently have been indicative of 'poor' generic stream health (Table 1) throughout the period until an improvement to 'fair' in the 2012-2013 period. However, due to the often weedy, more drain-like nature of this site, the more recently established $SQMCI_s$ may also be an appropriate index to consider in future.

3.2.17.2 Site downstream of the Mangawharawhara Stream confluence (MWH000490)

3.2.17.2.1 Taxa richness and MCI

Thirty-four surveys have been undertaken at this lower mid-reach site in the Mangawhero Stream between 1995 and February 2012. These results are summarised in Table 76, together with the results from the current period, and illustrated in Figure 72.

Table 76 Results of previous surveys performed in the Mangawhero Stream downstream of the Mangawharawhara Stream confluence, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
MWH000490	34	13-25	19	63-97	78	22	102	29	90

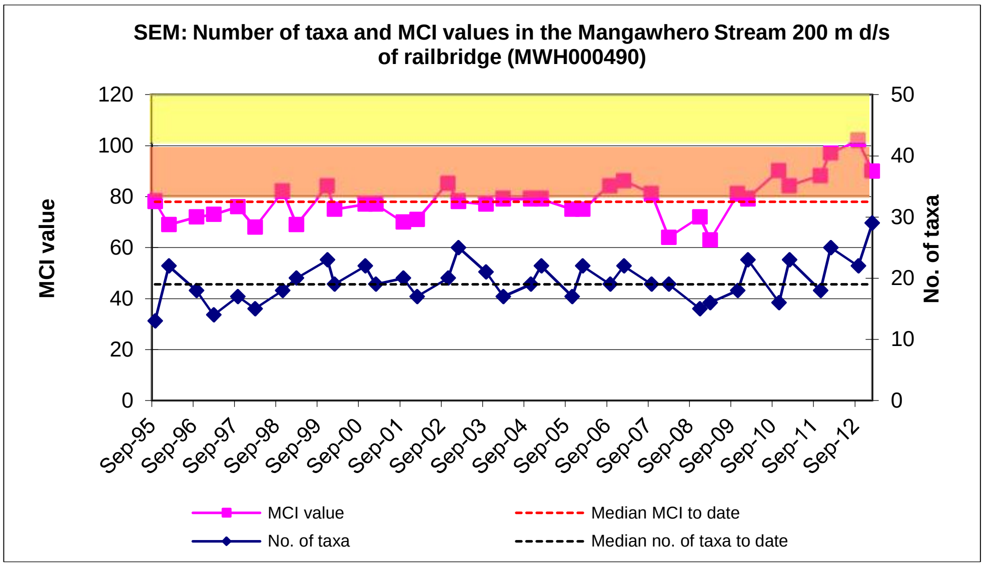


Figure 72 Numbers of taxa and MCI values in the Mangawhero Stream downstream of the railbridge and Mangawharawhara Stream confluence

A moderate range of richnesses (13 to 25 taxa) has been found with a moderate median richness of 19 taxa (more representative of typical richnesses in the lower-mid reaches of streams and rivers). During the 2012-2013 period, spring (22 taxa) and summer (29 taxa) richnesses were quite different and from three fewer to ten taxa more than this median richness, with the summer richness four taxa higher than the previous maximum taxa number.

MCI values have had a moderate range (34 units) at this site, more typical of a site in the middle to lower reaches of ringplain streams. However, the median value (78units) has been lower than typical of lower mid-reach sites elsewhere. However the spring, 2012 (102 units) and summer, 2013 (90 units) scores were a significant (Stark, 1998) 12 to 24 units higher than the historical median. These scores were coincident with the diversion of the major point source Eltham municipal wastewater discharge out of the Mangawhero Stream which was completed in June 2010. The spring, 2012 MCI score continued the more recent trend of exceeding (by 5 units) the previous maximum score. These scores categorised this site as having 'good' health generically (Table 1) in spring and 'fair' health in summer and, in terms of predictive relationships (Table 2), 'expected' (spring) and 'below expected' (summer) health for the equivalent mid-reaches of a stream with some ringplain catchment component (Mangawharawhara Stream which rises outside of the National Park). The historical median score (78 units) placed this site in the 'poor' and 'worse than expected' categories for generic and predictive methods of assessment respectively. The historical median score continues to reflect both the more lowland, swampy, nature of the headwaters of the Mangawhero Stream, but more particularly, the impact of the Eltham municipal wastewater treatment system's discharge on the water quality of the stream, prior to diversion in July, 2010.

3.2.17.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 77.

Table 77 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Mangawhero Stream downstream of the Mangawharawhara Stream confluence, between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	34	100	VA	A
MOLLUSCA	*Physa*	3	2	6		
	Potamopyrgus	4	9	26	A	
CRUSTACEA	Cladocera	5	3	9		
	Ostracoda	1	26	76		
	Paracalliope	5	32	94		
	Paraleptamphopidae	5	2	6		
	Talitridae	5	0	0		A
EPHEMEROPTERA	*Austroclima*	7	2	6		A
	Deleatidium	8	5	15	XA	XA
PLECOPTERA	*Zelandobius*	5	0	0	A	
COLEOPTERA	Elmidae	6	4	12	A	XA
TRICHOPTERA	*Aoteapsyche*	4	21	62	VA	VA
	Hydrobiosis	5	10	29		A
	Oxyethira	2	9	26		
	Pycnocentria	7	0	0		A
	Pycnocentrodes	5	3	9	XA	A
DIPTERA	*Aphrophila*	5	7	21	VA	
	Chironomus	1	2	6		
	Maoridiamesa	3	18	53		
	Orthocladiinae	2	31	91	A	A
	Tanypodinae	5	1	3		
	Tanytarsini	3	1	3		A
	Muscidae	3	1	3		
	Austrosimulium	3	13	38		

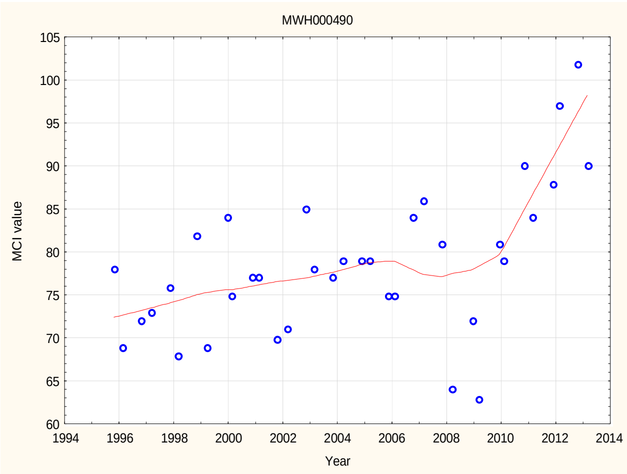
Prior to the current 2012-2013 period, 22 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', nine 'moderately sensitive', and twelve 'tolerant' taxa i.e. a higher proportion of 'tolerant' taxa than might be expected in the mid reaches of a small stream with a ringplain component. Predominant taxa have included one 'moderately sensitive' taxon [amphipod (*Paracalliope*)] and five 'tolerant' taxa [oligochaete worms, ostracod seed shrimps, net-building caddisfly (*Aoteapsyche*), and midges (orthoclads and *Maoridiamesa*)]. Three of these predominant taxa were dominant in the spring, 2012 community together with five of the other historically characteristic taxa (two of which, the extremely abundant 'moderately sensitive' stony-cased caddisfly and 'highly sensitive' mayfly had only been characteristic of the community on up to five previous occasions), and one taxon ['moderately sensitive' stonefly (*Zelandobius*)] which had not been recorded previously in abundance at this site. The summer, 2013 community was characterised by six of the same taxa dominant in spring, together with five additional taxa; two of which (both 'moderately sensitive') previously had not been characteristic of this site's communities (Table 77). The repeated extreme abundance of the 'highly sensitive' mayfly (*Deleatidium*) and additions of up to three characteristic 'moderately sensitive' taxa were further confirmation of improved water quality (and habitat) conditions following Eltham WWTP wastewater diversion. Despite seasonal similarities in many of the characteristic taxa, increased summer abundances within some 'moderately sensitive' taxa were reflected in the summer increase of 0.7 unit in $SQMCI_s$ scores (Tables 158 and 159). All taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 9% to 100 % of the past surveys with three of the 'sensitive' taxa on less than 16% of occasions coincident with recent habitat improvements.

3.2.17.2.3 Predicted stream 'health'

The Mangawhero Stream site below the Mangawharawhara Stream confluence, at an altitude of 190 m asl, is in the lower reaches of a stream draining a catchment comprised of the Ngaere Swamp drainage system and a mid-reach ringplain sub-catchment with its headwaters outside the National Park. A relationship for ringplain streams and rivers developed between MCI and altitude (Stark and Fowles, 2009) predicts a MCI value of 104 units for this site. The historical site median (78 units) was very significantly lower than this altitude prediction while the spring, 2012 (102 units) and summer, 2013 (90 units) scores were also below this predictive value by 2 and 14 units respectively although it must be noted that only part of the catchment is of ringplain derivation. Of the 36 surveys to date at this river site, all MCI scores have been less than 104 units.

3.2.17.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years 1995-2013) of SEM results collected to date from the site in the Mangawhero Stream downstream of the Mangawharawhara Stream confluence. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 73.



N = 36
Kendall tau = +0.410
p value = 0.0004 [>FDR, p = 0.001]
Significant at p <0.05 and p < 0.01; and significant after FDR

Figure 73 LOWESS trend plot of MCI data at the site downstream of the Mangawharawhara Stream confluence

A moderate and recently much more pronounced, and now statistically significant ($p < 0.01$, after FDR), temporal improvement in MCI scores has been illustrated at this more ringplain-like site in the lower reaches of the stream near its confluence with Waingongoro River. The wide range in LOWESS-smoothed scores (26 units) has more recently become ecologically significant over this eighteen year period. Scores trended downwards for 3 years after a steady improvement between 1995 and 2006 prior to the most recent marked improvement due to improved scores since the diversion of the Eltham WWTP wastes discharge out of the stream in July 2010.

The MCI scores generally have been indicative of 'poor' generic stream health (Table 1) with sporadic incursions into the 'fair' health category prior to 2010. The LOWESS-smoothed scores have remained in the 'poor' category through the period until 2010 and subsequently improved into the 'fair' category and most recently bordered on 'good' health. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a ringplain stream (recognising the partial ringplain component of this catchment and the position of the site in the lower reach of this small stream prior to joining the mid-reaches of a larger ringplain river), stream health has been 'worse than expected' almost throughout the entire eighteen year period, but entered the 'expected' category in the previous 2011-2012 survey period.

3.2.17.3 Discussion

Seasonal MCI values atypically remained the same between spring and summer at the upper reach (upstream of the Eltham WWTP) with scores identical, in comparison with the historical median summer decrease (3 units, Appendix II) for this site. A more typical decrease (12 units) was found at the lower site (downstream of the Mangawharawhara Stream confluence) in the absence of the WWTP discharge which had significantly impacted on water quality at this site prior to mid 2010. This was larger in comparison with a seasonal 4 unit median summer historical decrease at this site (Appendix II). Seasonal communities at the upper reach site shared only nine common taxa (33% of the 27 taxa found in 2012-2013, despite the identical MCI scores) compared with 21 shared common taxa (70% of the 30 taxa) at the lower site;

a more typical seasonal change in community structure historically found at the lower of the two sites. The two sites shared 12 common taxa (55% of the 22 taxa) in spring and 14 common taxa (36% of 39 taxa) in summer, indicative of the dissimilarity in spatial community structures in spring and more so in summer, as might be expected given the significantly different physical and physicochemical habitats at these two sites.

MCI scores typically (for this stream) improved in a downstream direction by 17 units in spring and in summer (only by 5 units), over a stream distance of 16.5 km between the upper and lower sites of this stream. This was principally as a result of the variability and improvement in physical habitat and physicochemical water quality conditions in a downstream direction between the two sites and enhanced in recent years by the diversion of the Eltham wastewater discharge out of the stream. The much lower flow conditions at the time of the summer survey also contributed to the seasonal difference in rate of downstream MCI improvement.

3.2.18 Huatoki Stream

The results of spring (2012) and summer (2012-2013) surveys are summarised in Table 160 and Table 161, Appendix I.

3.2.18.1 Hadley Drive site (HTK000350)

3.2.18.1.1 Taxa richness and MCI

Thirty-two surveys have been undertaken, between 1996 and March 2012, at this lower mid-reach, unshaded site, draining open developed farmland, on the outskirts of New Plymouth city. These results are summarised in Table 78, together with the results from the current period, and illustrated in Figure 74.

Table 78 Results of previous surveys performed in the Huatoki Stream at Hadley Drive together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to March 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		March 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
HTK000350	32	22-31	26	79-109	94	23	114	34	101

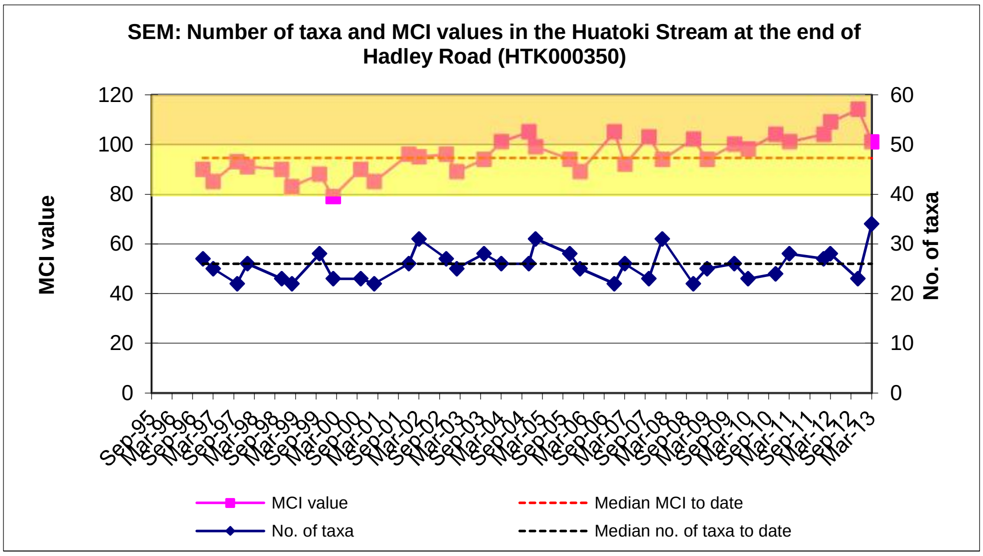


Figure 74 Numbers of taxa and MCI values in the Huatoki Stream at the end of Hadley Drive

A moderate range of richnesses (22 to 31 taxa) has been found with a relatively high median richness of 26 taxa, relatively typical of richnesses in the mid to lower reaches of ringplain streams rising outside of the National Park. During the 2012-2013 period spring (23 taxa) and summer (34 taxa) richnesses were very different, below (spring) and well above (summer) the historical median richness coincident with minimal periphyton cover (spring) and patchy periphyton mats and filamentous algae cover (summer) on the predominantly stony-bouldery substrate of this unshaded site. The summer richness was three taxa more than the maximum recorded to date for this site, although 53% of taxa were present as rarities (less than 5 individuals per taxon) in this community.

MCI values have had a relatively wide range (30 units) at this site, typical of mid to lower reach sites on the ringplain. The historical median value (94 units) also has been typical of mid-reach sites rising outside the National Park elsewhere on the ringplain, and the spring, 2012 (114 units) and summer, 2013 (101 units) scores were a significant (Stark, 1998) 20 units and an insignificant 7 units above the historical median respectively. The spring score was also 5 units higher than the historical maximum. The scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the mid-reaches of a ringplain stream on these occasions. The historical median score (94 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.18.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 79.

Table 79 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Huatoki Stream at Hadley Drive, between 1996 and March 2012 [32 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	9		
ANNELIDA	Oligochaeta	1	21	66		
MOLLUSCA	*Latia*	5	2	6		
	Potamopyrgus	4	21	66		A
CRUSTACEA	*Paracalliope*	5	6	19		
EPHEMEROPTERA	*Austroclima*	7	8	25	A	A
	Coloburiscus	7	17	53	VA	XA
	Deleatidium	8	3	9	XA	XA
	Nesameletus	9	9	28	VA	VA
	Zephlebia group	7	19	59	VA	A
PLECOPTERA	*Zelandobius*	5	8	25	A	
	Zelandoperla	8	1	3		
COLEOPTERA	Elmidae	6	10	31	VA	VA
MEGALOPTERA	*Archichauliodes*	7	2	6	A	A
TRICHOPTERA	*Aoteapsyche*	4	31	97	VA	VA
	Costachorema	7	18	56		
	Hydrobiosis	5	21	66		A
	Neurochorema	6	3	9		
	Oxyethira	2	4	13		
	Pycnocentrodes	5	4	13		
DIPTERA	*Aphrophila*	5	16	50	A	
	Maoridiamesa	3	18	56		
	Orthocladiinae	2	32	100		
	Tanytarsini	3	13	41		
	Empididae	3	1	3		
	Muscidae	3	5	16		
	Austrosimulium	3	14	44		

Prior to the current 2012-2013 period 27 taxa had characterised the community at this site on occasions. These have comprised only three 'highly sensitive', 13 'moderately sensitive' and 11 'tolerant' taxa i.e. a relatively high proportion of 'tolerant' taxa as would be expected in the lower mid-reaches of a ringplain stream rising outside the National Park.

Predominant taxa have included no 'highly sensitive' taxa; five 'moderately sensitive' taxa [mayflies (*Coloburiscus* and *Zephlebia* group), free-living caddisflies (*Hydrobiosis* and *Costachorema*), and cranefly (*Aphrophila*)]; and five 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and midges (orthoclads and *Maoridiamesa*)].

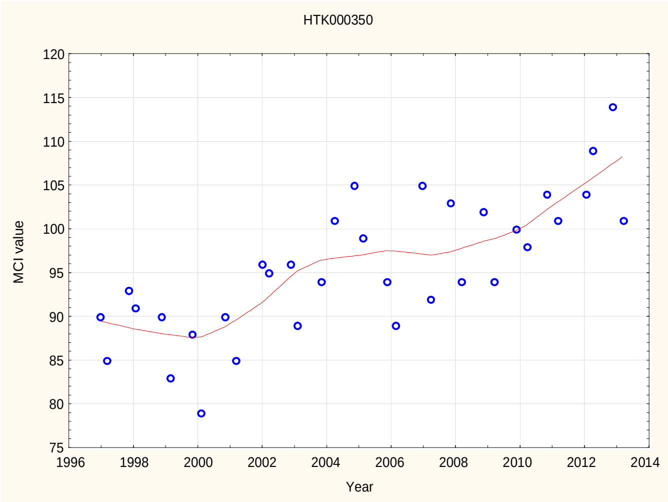
Ten of the historically characteristic taxa were dominant in the spring, 2012 community comprising four of the predominant taxa (above) together with six of the other historically characteristic taxa. The summer, 2013 community was characterised by eight of the taxa dominant in spring, together with an additional one 'moderately sensitive' and one 'tolerant' taxa, all of which previously had been characteristic of this site's communities, but with two fewer of the 'moderately sensitive' taxa earlier characteristic of the spring community. Few significant differences in spring and the summer abundances within the dominant taxa were reflected in the similarity in seasonal $SQMCI_s$ scores which varied by only 0.2 unit (Table 160 and 161). The six taxa which were recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 9% to 97% of past surveys.

3.2.18.1.3 Predicted stream 'health'

The Huatoki Stream rises below the National Park boundary and the site at Hadley Drive is in the lower mid-reaches at an altitude of 60 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 91 units for this site. The historical site median (94 units) is only 3 units above this altitude prediction while the spring survey score (114 units) was higher by a significant (Stark, 1998) 23 units and the summer score (101 units) by an insignificant 10 units than the predictive value. Of the 34 surveys to date at this site, 32% of MCI scores have been less than 91 units. The current spring and summer MCI scores were higher than those typical of historical conditions.

3.2.18.1.4 Temporal trends in 1996 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the seventeen years of SEM results collected to date from the site, in the Huatoki Stream at Hadley Drive. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 75.



N = 34
Kendall tau = +0.543
p level = < 0.0001 [>FDR, p = < 0.001]
Significant at p< 0.05 and p < 0.01;
and after FDR application

Figure 75 LOWESS trend plot of MCI data at the Hadley Drive site

A strong temporal improvement ($p < 0.01$) in MCI scores, particularly since 2000 has been illustrated at this site on the outskirts of New Plymouth. The overall trend has also been statistically significant after FDR application and the wide LOWESS-smoothed range of MCI scores (20 units) has ecological significance and may have been related to improvements in farming practices (including more recent riparian fencing) and/or wastes disposal in the rural catchment between the stream's seepage sources (below the National Park) and urban New Plymouth.

LOWESS-smoothed MCI scores have been indicative of 'fair' generic stream health (Table 1) almost throughout the period improving to 'good' health more recently. In terms of predictive relationships (Table 2) for a site in the lower mid-reaches of a ringplain stream, health has remained in the 'expected' category over the majority of the seventeen year period and more recently entered the 'better than expected' category (Figure 75).

3.2.18.2 Huatoki Domain site (HTK000425)

3.2.18.2.1 Taxa richness and MCI

Thirty-two surveys have been undertaken at this lower reach site in the Huatoki Stream toward the downstream boundary of the Huatoki Domain between 1996 and March 2012. These results are summarised in Table 80, together with the results from the current period, and illustrated in Figure 76.

Table 80 Results of previous surveys performed at Huatoki Stream in Huatoki Domain, together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to March 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		March 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
HTK000425	32	17-32	26	91-115	103	24	115	29	105

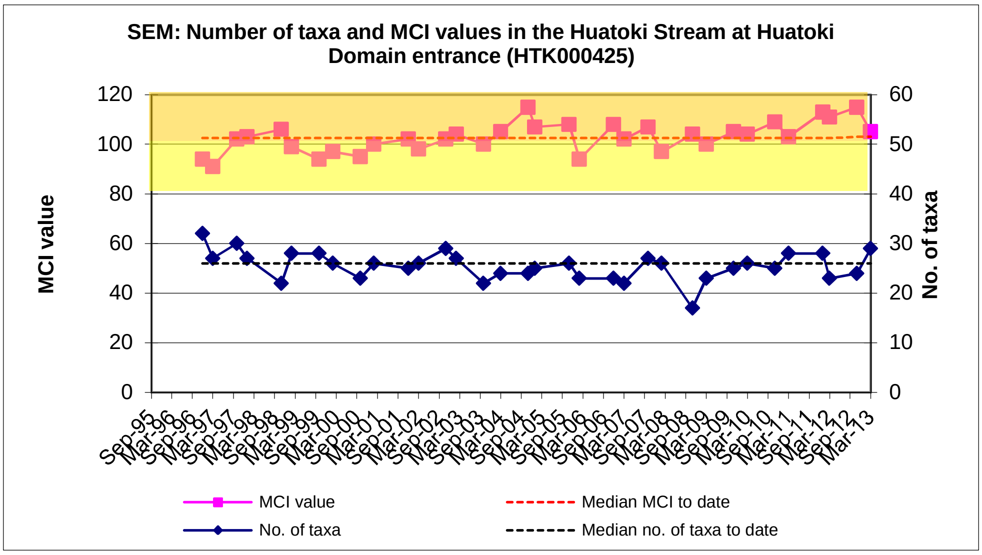


Figure 76 Numbers of taxa and MCI values in the Huatoki Stream at the Huatoki Domain

A moderate range of richnesses (17 to 32 taxa) has been found, with a median richness of 26 taxa (more representative of typical richnesses for the lower reaches of ringplain streams rising outside the National Park boundary). During the 2012-2013 period spring (24 taxa) and summer (29 taxa) richnesses increased in summer but seasonal richnesses were both within three taxa of this median richness.

MCI values have had a moderately wide range (24 units) at this site. The median value (103 units) has been higher than typical of lower reach sites elsewhere on the ringplain however. The spring, 2012 (115 units) and summer, 2013 (105 units) scores were also higher than typical for such a site; significantly 12 and insignificantly 2 units above the historical median respectively. These scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring and summer) health for the lower reaches of a ringplain stream coincident with the extensive riparian cover provided by the Huatoki Domain. The historical median score (103 units) placed this site in the 'good' and 'better than expected' categories for generic and predictive methods of assessment respectively.

3.2.18.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 81.

Table 81 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Huatoki Stream at Huatoki Domain, between 1996 and March 2012 [32 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	9		
ANNELIDA	Oligochaeta	1	28	88		A
MOLLUSCA	*Latia*	5	15	47		
	Potamopyrgus	4	27	84	A	A
CRUSTACEA	*Paracalliope*	5	3	9		
EPHEMEROPTERA	*Austroclima*	7	6	19	A	VA
	Coloburiscus	7	27	84	VA	XA
	Deleatidium	8	4	13	VA	A
	Mauiulus	5	1	3		
	Nesameletus	9	0	0		A
	Zephlebia group	7	31	97	A	A
PLECOPTERA	*Zelandobius*	5	15	47	A	
COLEOPTERA	Elmidae	6	21	66	VA	XA
	Ptilodactylidae	8	3	9		
MEGALOPTERA	*Archichauliodes*	7	14	44	A	A
TRICHOPTERA	*Aoteapsyche*	4	31	97	A	VA
	Costachorema	7	1	3		
	Hydrobiosis	5	6	19		
	Pycnocentrodes	5	19	59		
DIPTERA	*Aphrophila*	5	1	3		
	Orthocladiinae	2	10	31		
	Austrosimulium	3	31	97		A
	Tanyderidae	4	1	3		

Prior to the current 2012-2013 period, 22 taxa had characterised the community at this site on occasions. These have comprised two 'highly sensitive', 13 'moderately sensitive', and seven 'tolerant' taxa i.e. a higher proportion of 'sensitive' taxa than might be expected in the lower reaches of a ringplain stream, coincident with the extensive riparian cover provided by the Huatoki Domain.

Predominant taxa have included no 'highly sensitive' taxa; four 'moderately sensitive' taxa [mayflies (*Zephlebia* group and *Coloburiscus*), elmid beetles, and stony-cased caddisfly (*Pycnocentrodes*)]; and four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and sandfly (*Austrosimulium*)].

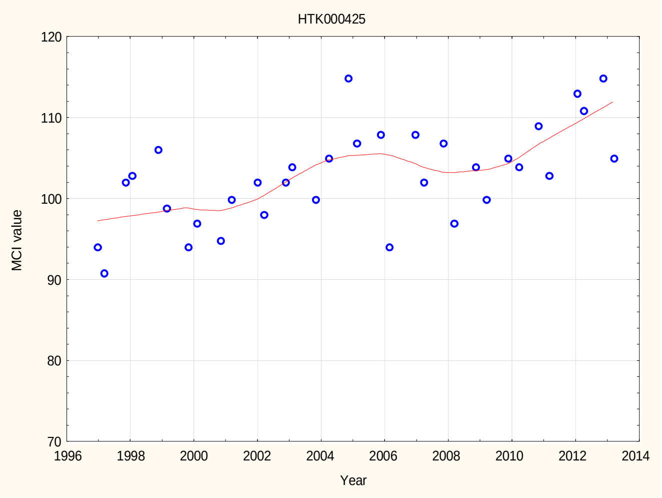
Nine of the historically characteristic taxa were dominant in the spring, 2012 community and were comprised of five of the predominant taxa (above) together with one 'highly sensitive ' and three 'moderately sensitive' taxa. The summer, 2013 community was characterised by eight of the taxa dominant in spring, with one fewer 'moderately sensitive' taxon and two additional 'tolerant' taxa than characteristic of the spring community (Table 81). Despite some increases in numerical abundances within some taxa in summer, plus one 'highly sensitive' taxon [mayfly (*Nesameletus*)] not previously found in abundance at this site, there was minimal change in seasonal $SQMCI_s$ scores (0.3 unit)(Tables 160 and 161). All taxa which were recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 13% to 97% of past surveys.

3.2.18.2.3 Predicted stream 'health'

The Huatoki Stream rises below the National Park boundary and the site at Hadley Domain is in the lower mid-reaches at an altitude of 30 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 87 units for this site. The historical site median (103 units) is well above this altitude prediction coincident with the extensive riparian vegetation cover of the Huatoki Domain and both the spring survey score (115 units) and the summer score (105 units) were significantly higher (Stark, 1998) than the predictive value. Of the 34 surveys to date at this site, no MCI scores have been less than 87 units, indicating that the current spring and summer MCI scores were typical of historical conditions although toward the higher end of the range.

3.2.18.2.4 Temporal trends in 1996 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the seventeen years of SEM results collected to date from the site in the Huatoki Stream at Huatoki Domain. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 77.



N = 34
Kendall tau = +0.448
p level = 0.0002 [>FDR, p < 0.001]
Significant at p< 0.05 and p < 0.01; and after FDR application

Figure 77 LOWESS trend plot of MCI data for the Huatoki Domain site

A similar temporal trend of a marked improvement in MCI scores, but not quite as strong as that found at the upstream site (at Hadley Drive), was identified at this site in the Domain although scores peaked with a small decrease after 2006 prior to a recent steady increasing trend. The overall trend has been very statistically significant after FDR application (p< 0.01) and the LOWESS-smoothed range of scores (15 units) has become ecologically significant. The trend may have been related to the upstream catchment activities noted above (Section 3.2.18.1.4) as no nearby habitat changes have been recorded within the Domain.

The smoothed MCI scores which indicated 'fair' generic stream health (Table 1) much earlier in the monitoring period, improved to 'good' stream health consistently since 2002. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, health has remained 'better than expected' over the entire period, further indication of the value of the extensive riparian cover provided by the Huatoki Domain.

3.2.18.3 Site near coast (HTK000745)

3.2.18.3.1 Taxa richness and MCI

Thirty-two surveys have been undertaken at this lower reach site in the Huatoki Stream between 1996 and March 2012. These results are summarised in Table 82, together with the results from the current period, and illustrated in Figure 78.

Table 82 Results of previous surveys performed in Huatoki Stream at the site near the coast, together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to March 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		March 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
HTK000745	32	14-27	22	69-101	86	17	101	25	88

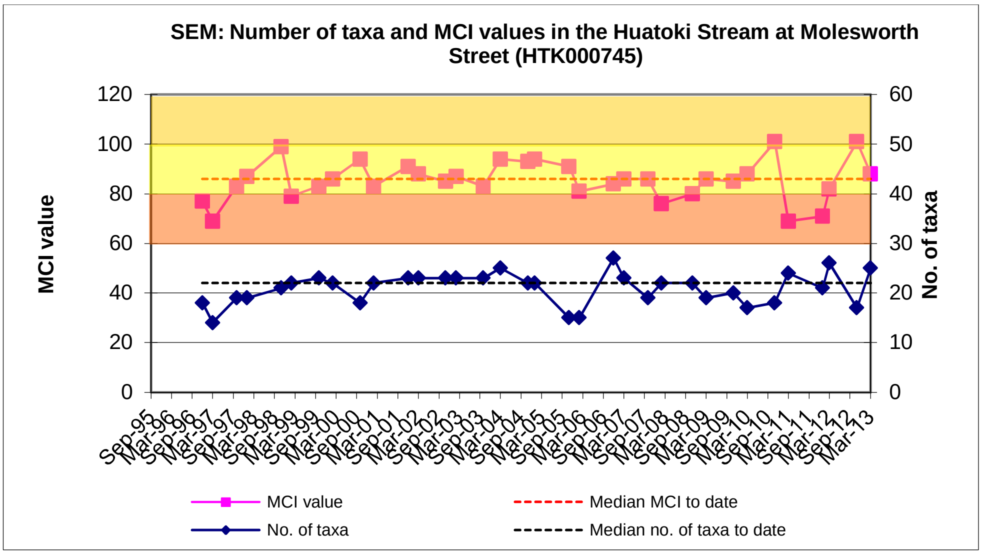


Figure 78 Numbers of taxa and MCI values in the Huatoki Stream at Molesworth Street (near coast)

A moderate range of richnesses (14 to 27 taxa) has been found, with a median richness of 22 taxa (more representative of typical richnesses in the lower reaches of ringplain streams rising outside the National Park boundary). During the 2012-2013 period spring (17 taxa) and summer (25 taxa) richnesses were relatively different and lower in spring and higher in summer than this median richness.

MCI values have had a relatively wide range (32 units) at this site. The median value (86 units) has been typical of lower reach sites elsewhere on the ringplain however, and the spring, 2012 (101 units) and summer, 2013 (88 units) scores varied significantly and were a significant 15 units above (spring) and 2 units above (summer) the historical median. The spring score was equal with the historical maximum score despite pulsed flows a short distance downstream of a relatively recently installed weir and fish pass (for beautification purposes) while the summer score was nearer the historical median score. These scores categorised this site as having 'good (spring) and 'fair' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower reaches of a ringplain stream. The historical median score (86 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.18.3.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 83.

Table 83 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Huatoki Stream at the site near the coast, between 1996 and March 2012 [32 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	3		
ANNELIDA	Oligochaeta	1	32	100	A	VA
MOLLUSCA	*Ferrissia*	3	1	3		
	Latia	5	3	9		
	Potamopyrgus	4	32	100	VA	XA
CRUSTACEA	Ostracoda	1	1	3		
	Paratya	3	2	6		VA
EPHEMEROPTERA	*Coloburiscus*	7	4	13		
	Zephlebia group	7	5	16		
PLECOPTERA	*Zelandobius*	5	3	9		
COLEOPTERA	Elmidae	6	16	50	VA	XA
TRICHOPTERA	*Aoteapsyche*	4	3	9		
	Oxyethira	2	1	3		
	Pycnocentrodes	5	9	28		
	Triplectides	5	2	6		
DIPTERA	*Aphrophila*	5	1	3		
	Orthocladiinae	2	14	44		
	Polypedilum	3	1	3		
	Empididae	3	2	6		
	Austrosimulium	3	1	3		
	Tanyderidae	4	5	16		

Prior to the current 2012-2013 period, 21 taxa had characterised the community at this site on occasions. These have comprised no 'highly sensitive', eight 'moderately sensitive', and 13 'tolerant' taxa i.e. a high proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain stream.

Predominant taxa have included one 'moderately sensitive' taxon (elmid beetles) and two 'tolerant' taxa [oligochaete worms and snail (*Potamopyrgus*); both on every occasion].

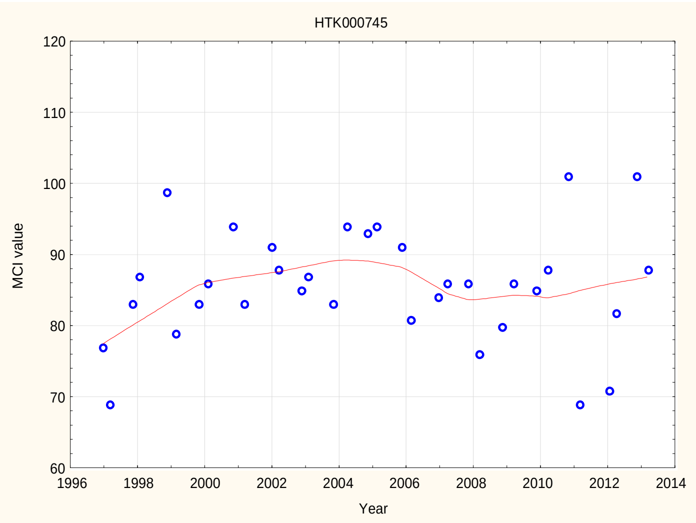
Three of the historically characteristic taxa were dominant in the spring, 2012 community and comprised all of the predominant 'tolerant' taxa (above). The summer, 2013 community was characterised by the same three taxa dominant in spring, together with an additional 'tolerant' taxon [freshwater shrimp (*Paratya*)], which previously had been seldom characteristic of this site's communities (Table 83). Minimal differences in the balances of numerical taxa dominances were reflected in the similarity in seasonal $SQMCI_s$ scores which were within 0.2 unit (Table 160 and 161). The four taxa which were recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 6% to 100% of past surveys.

3.2.18.3.3 Predicted stream 'health'

The Huatoki Stream rises below the National Park boundary and the site near the coast is in the lower reaches at an altitude of 5 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 85 units for this site. The historical site median (86 units) is only one unit above this altitude prediction while the spring score (101 units) was a significant 15 units above, and the summer score (88 units) 3 units above the predictive value. Of the 34 surveys to date at this site, 41% of MCI scores have been less than 85 units, indicating that the current MCI scores were relatively more typical of historical conditions.

3.2.18.3.4 Temporal trends in 1996 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the seventeen years of SEM results collected to date from the site, in the Huatoki Stream near the coast. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 79.



N = 34
Kendall tau = + 0.051
p level = 0.672 [>FDR, p = 0.712]
N/S at $p < 0.05$

Figure 79 LOWESS trend plot of MCI data for the site near the coast

A trend of steady improvement in MCI scores had occurred at this urbanised site until 2005 after which scores trended downward until plateauing more recently. This has resulted in an overall weakly positive and statistically non-significant trend for the seventeen year monitoring period. However, the range of LOWESS-smoothed scores (12 units) has some ecological significance probably related in part to those activities noted for the two sites further upstream in the Huatoki catchment (see above) and more recently to the pulsed flows and subtle habitat changes caused by the beautification project which involved construction of a weir and a fishpass.

Smoothed MCI scores indicative of 'fair' generic stream health (Table 1) have been recorded for all but the first years of the monitoring programme (Figure 79) and, in terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, health has remained in the 'expected' category throughout the monitoring period.

3.2.18.4 Discussion

Seasonal MCI values typically decreased between spring and summer at all three sites by a significant 13 units at the Hadley Drive site, 10 units at the Huatoki Domain site, and by a significant 13 units near the coast. These results may be compared with historical median seasonal data (Appendix II) which indicate typical small summer MCI decreases of 2 and 3 units at Hadley Drive and Huatoki Domain respectively and minimal change near the coast. Seasonal communities shared 50% of the 38 taxa common at the mid-reach Hadley Drive site, 71% of 31 taxa at Huatoki Domain, and 40% of 30 taxa at the furthest downstream site in the lower reaches near the coast indicative of the least dissimilarity in seasonal community composition at the more stable Huatoki Domain site and increased dissimilarity at the furthest downstream site.

Community composition indicated some improvement at the Domain site where proportionately more higher scoring taxa were recorded. Further downstream, near the mouth, urbanisation and habitat modification coincided with a significant variation in community composition. This site's faunal community was characterised by an increase in the 'tolerant' taxa proportion of the community composition.

Community composition varied markedly through the mid reach to lower reach length of the stream surveyed. A total of 32 taxa was recorded in spring of which only eleven taxa were present at all three sites. These included two 'highly sensitive', five 'moderately sensitive', and four 'tolerant' taxa with only one 'moderately sensitive' taxon (elmid beetles) abundant at all three sites. A higher total of 45 taxa was found along the stream's surveyed length by the summer survey when only fourteen taxa were present at all three sites. Seven of these were the same as the widespread taxa in spring with the addition of one 'moderately sensitive' and three 'tolerant taxa and loss of one 'highly sensitive' taxon which was widespread in spring. Two taxa were abundant at all three sites in summer, ['moderately sensitive' elmid beetles and 'tolerant' snail (*Potamopyrgus*)]. Dissimilarities in spatial community structure along the surveyed length (mid to lower reaches) of the Huatoki Stream were slightly more pronounced in summer than in spring.

MCI score increased atypically (for ringplain streams) in a downstream direction by 1 unit in spring and 4 units in summer between the open farmland site (Hadley

Drive) and the Huatoki Domain site, coincident with the improved habitat provided by the riparian vegetation cover in the domain, but relatively typical of increases recorded by most past surveys (e.g. historical median scores have increased by 9 units (spring) and 8 units (summer) between these sites (Appendix II)). MCI score fell significantly by 14 units (spring) and also significantly by 17 units (summer) through the city between the Domain and the coast, despite a change in elevation of only 25 m, representing a relatively high rate of MCI decrease of 3.6 and 4.3 units/km respectively. Both spring and summer decreases were well above the rate expected through the mid to lower reaches of Taranaki ringplain rivers and streams sourced outside of the National Park (TRC, 1999). These MCI rates of decrease were amplified by the presence of the improved habitat within the Huatoki Domain and possibly by recent deterioration in habitat adjacent to the coastal site. There were decreases in MCI between the open farmland site and the coast of 13 units (spring) and 13 units (summer) coincident with some impacts of urbanisation on the stream's macroinvertebrate fauna.

Using the long-term median SEM MCI scores for each site (Appendix II), the rate of decline between mid-reach site and lower reach site near the coast has been 2.1 MCI units/km over the surveyed length. Therefore rates of MCI decline over the 2012-2013 period were higher in spring and in summer than the average rate for the 1995 to 2012 period.

3.2.19 Kaupokonui River

Five sites located along the length of the Kaupokonui River were included in the SEM programme, commencing in the 1999-2000 year for the purpose of long term monitoring of the impacts of riparian vegetation planting initiatives throughout this catchment. Two sites, at Opunake Road (KPK000250) and near the coast (KPK000990), were established specifically for this purpose, while the remaining three sites were components of existing consent monitoring programmes.

The results of the spring, 2012 survey are presented in Table 162 and the summer, 2012-2013 survey in Table 163, Appendix I.

3.2.19.1 Opunake Road site (KPK000250)

3.2.19.1.1 Taxa richness and MCI

Twenty-seven surveys have been undertaken in the Kaupokonui River at this upper mid-reach site at Opunake Road (draining relatively open farmland approximately 3.3 km downstream of the National Park) between 1999 and February 2012. These results are summarised in Table 84, together with the results from the current period, and illustrated in Figure 80.

Table 84 Results of previous surveys performed in the Kaupokonui River at Opunake Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPK000250	27	20-36	27	125-138	129	24	128	23	125

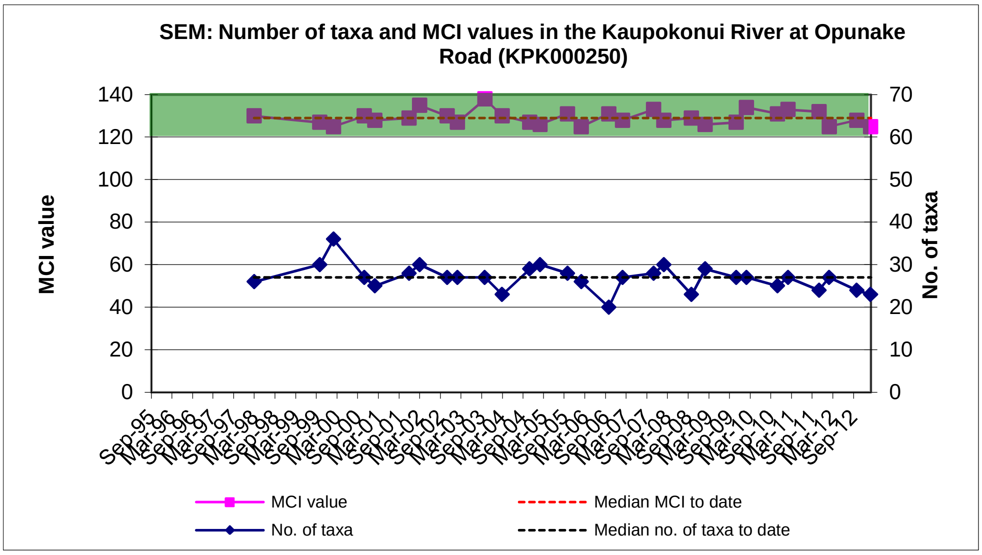


Figure 80 Numbers of taxa and MCI values in the Kaupokonui River at Opunake Road

A relatively wide range of richnesses (20 to 36 taxa) has been found; wider than might be expected, with a median richness of 27 taxa (more representative of typical richnesses in the upper mid-reaches of ringplain streams and rivers). During the 2012-2013 period spring (24 taxa) and summer (23 taxa) richnesses were very similar and slightly below the median taxa number coincidental with minimal substrate periphyton cover on both occasions at this site.

MCI values have had a narrow range (13 units) at this site, more typical of sites in the upper reaches of ringplain rivers. The median value (129 units) has been higher than typical of mid-reach sites elsewhere on the ringplain however, with the spring, 2012 (128 units) and summer, 2013 (125 units) scores above those typical for such a site and within 3 units of the historical median. These scores categorised this site as having 'very good' (spring) and 'very good' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' health on both occasions for the upper mid reaches of a ringplain river. The historical median score (129 units) placed this site in the 'very good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.19.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 85.

Table 85 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kaupokonui River at Opunake Road between 1995 and February 2012 [27 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	2	7		
EPHEMEROPTERA	*Austroclima*	7	2	7		A
	Coloburiscus	7	25	93	VA	VA
	Deleatidium	8	27	100	XA	XA
	Nesameletus	9	13	48		VA
PLECOPTERA	*Acroperla*	5	1	4		
	Megaleptoperla	9	18	67		A
	Zelandoperla	8	25	93	VA	VA
COLEOPTERA	Elmidae	6	27	100	VA	A
	Hydraenidae	8	3	11		
MEGALOPTERA	*Archichauliodes*	7	6	22		
TRICHOPTERA	*Aoteapsyche*	4	21	78	A	A
	Costachorema	7	5	19		
	Hydrobiosis	5	4	15		
	Beraeoptera	8	20	74	A	
	Helicopsyche	10	3	11		
	Olinga	9	15	56	A	VA
	Pycnocentrodes	5	11	41		
DIPTERA	*Aphrophila*	5	25	93	A	A
	Eriopterini	5	6	22		
	Maoridiamesa	3	7	26		
	Orthocladiinae	2	8	30		

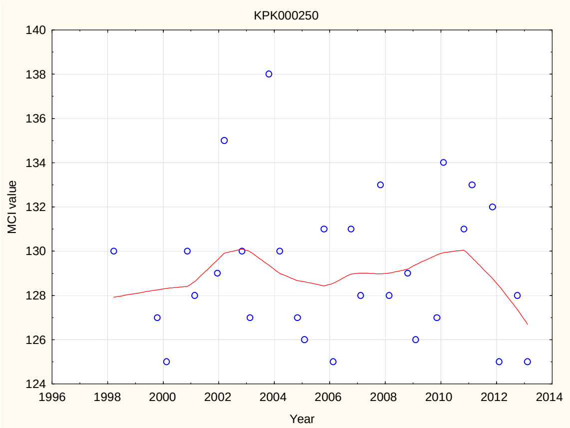
Prior to the current 2012-2013 period, 22 taxa had characterised the community at this site on occasions. These have comprised eight 'highly sensitive', ten 'moderately sensitive', and four 'tolerant' taxa i.e. a majority of 'sensitive' taxa as would be expected in the upper mid-reaches of a ringplain stream. Predominant taxa have included five 'highly sensitive' taxa [mayfly (*Deleatidium*, on every sampling occasion), stoneflies (*Megaleptoperla* and *Zelandoperla*), and cased caddisflies (*Beraeoptera* and *Olinga*)]; three 'moderately sensitive' taxa [mayfly (*Coloburiscus)*, elmid beetles, and cranefly (*Aphrophila*)]; and one 'tolerant' taxon [net-building caddisfly (*Aoteapsyche*)]. Eight of these predominant taxa were dominant in the spring, 2012 community. These were comprised of four 'highly sensitive', three 'moderately sensitive', and one 'tolerant' taxa. All but one of these predominant taxa were again dominant in the summer, 2013 community together with another two 'highly sensitive' taxa. Despite the larger number of numerically dominant taxa in the summer community, there was minimal change in seasonal $SQMCI_s$ values which differed by only 0.4 unit (Tables 162 and 163). All taxa recorded as very or extremely abundant during spring and/or summer had characterized this site's communities on 48% to 100% of past surveys. Two of the taxa recorded as very or extremely abundant during spring and/or summer have characterised this site's communities on every past survey occasion.

3.2.19.1.3 Predicted stream 'health'

The Kaupokonui River site at Opunake Road is 3.3 km downstream of the National Park boundary at an altitude of 380 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 123 (altitude) and 118 (distance) for this site. The historical site median (129) is 6 units higher than the altitude prediction and a significant (Stark, 1998) 11 units higher than the distance predictive value. The spring, 2012 survey score (128 units) was 5 to 10 units higher than these predictive values, whereas the summer, 2013 score (125 units) was higher than both predictive values by 2 to 7 units. Of the 27 surveys to date at this site, no MCI scores have been less than 118 units while all scores have been greater than 123 units, further indicative of the better than predicted health of the river at this site within 3.5 km of the National Park.

3.2.19.1.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the fourteen years of SEM results collected to date from the site in the Kaupokonui River at Opunake Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 81.



N = 29
Kendall tau = -0.015
p level = 0.907 [>FDR, p = 0.907]
N/S at p < 0.05

Figure 81 LOWESS trend plot of MCI data at the Opunake Road site

The temporal trend of a very small decline in MCI scores has not been statistically significant at this site in the upper mid-reaches of the river over the fourteen year monitoring period. The LOWESS-smoothed range of scores (3 units) was extremely narrow and not ecologically significant. Smoothed MCI scores were continuously indicative of 'very good' generic river health (Table 1), while in terms of predictive relationships (Table 2) for a site in the upper mid reaches of a ringplain river, health has been in the 'expected' category for the entire period.

3.2.19.2 Site upstream of the Kaponga oxidation ponds system (KPK000500)

3.2.19.2.1 Taxa richness and MCI

Thirty surveys have been undertaken in the Kaupokonui River at this mid-reach site at the site upstream of the Kaponga oxidation ponds system between 1995 and February 2012. These results are summarised in Table 86, together with the results from the current period, and illustrated in Figure 86.

Table 86 Results of previous surveys performed in the Kaupokonui River at the site upstream of the Kaponga oxidation ponds system together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPK000500	30	21-33	26	98-133	116	22	128	20	120

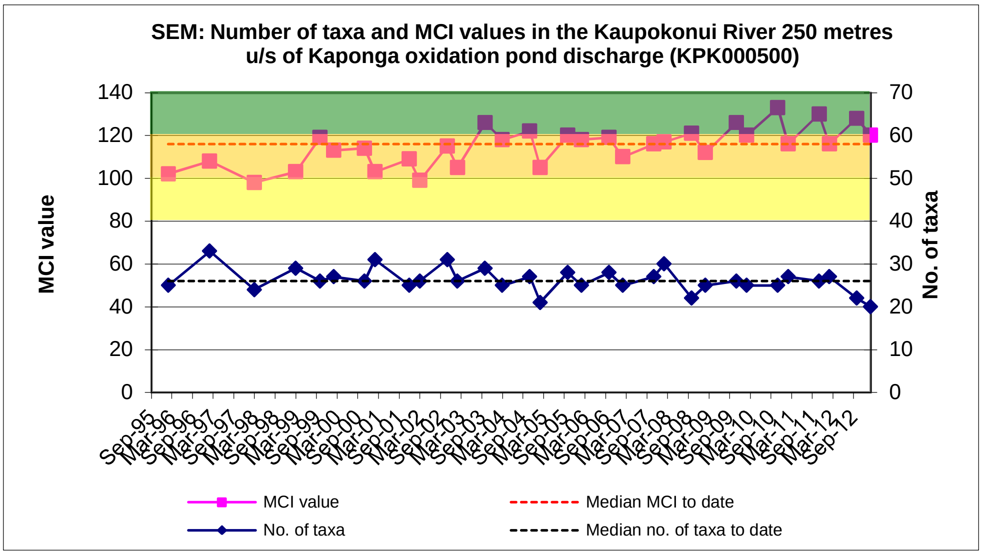


Figure 82 Numbers of taxa and MCI values in the Kaupokonui River upstream of Kaponga oxidation pond system

A moderate range of richnesses (21 to 33 taxa) has been found with a median richness of 26 taxa, typical of richnesses in the mid reaches of ringplain streams and rivers. During the 2012-2013 period, spring (22 taxa) and summer (20 taxa) richnesses were very similar but up to 6 taxa fewer than the median taxa number coincidental with minimal periphyton substrate cover on both occasions.

MCI values have had a relatively wide range (35 units) at this site, slightly wider than typical of sites in the mid-reaches of ringplain rivers. The median value (116 units) has been slightly higher than typical of mid-reach sites elsewhere on the ringplain with the spring, 2012 (128 units) and summer, 2013 (120 units) scores four units higher (summer) and significantly higher by 12 units (spring) than the historical median. These scores categorised this site as having 'very good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' health in spring and summer respectively for the mid-reaches of a ringplain river. The historical median score (116 units) placed this site in the 'good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.19.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 87.

Table 87 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kaupokonui River upstream of the Kaponga oxidation ponds system between 1995 and February 2012 [30 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	2	7		
ANNELIDA	Oligochaeta	1	6	20		
EPHEMEROPTERA	*Austroclima*	7	1	3		
	Coloburiscus	7	29	97	XA	XA
	Deleatidium	8	25	83	XA	XA
	Nesameletus	9	13	43	VA	VA
PLECOPTERA	*Megaleptoperla*	9	1	3		
	Zelandoperla	8	7	23	A	
COLEOPTERA	Elmidae	6	27	90	A	VA
MEGALOPTERA	*Archichauliodes*	7	14	47	A	A
TRICHOPTERA	*Aoteapsyche*	4	26	87	A	VA
	Costachorema	7	18	60		
	Hydrobiosis	5	8	27		
	Beraeoptera	8	14	47	VA	A
	Olinga	9	3	10	A	
	Oxyethira	2	1	3		
	Pycnocentrodes	5	16	53	A	A
DIPTERA	*Aphrophila*	5	29	97	VA	VA
	Eriopterini	5	4	13	A	
	Maoridiamesa	3	21	70		
	Orthocladiinae	2	20	67		
	Tanytarsini	3	5	17		
	Empididae	3	1	3		
	Muscidae	3	3	10		
	Austrosimulium	3	1	3		

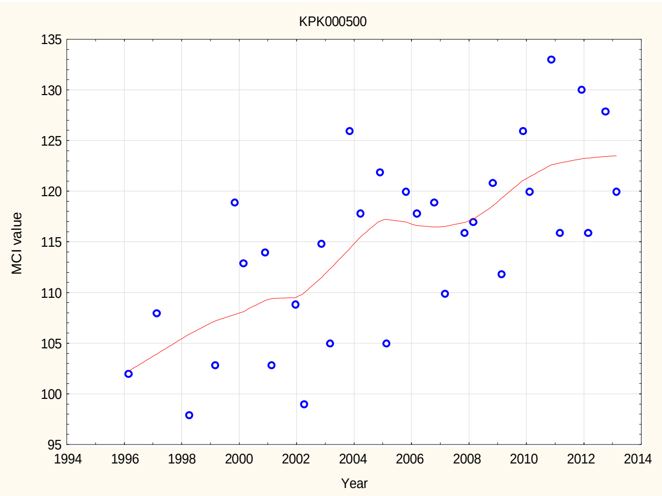
Prior to the current 2012-2013 period, 25 taxa had characterised the community at this site on occasions. These have comprised six 'highly sensitive', nine 'moderately sensitive', and ten 'tolerant' taxa i.e. a majority of 'sensitive' taxa but a small downstream increase in 'tolerant' taxa compared with the Opunake Road site, as would be expected in the mid-reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; five 'moderately sensitive' taxa [mayfly (*Coloburiscus*), elmid beetles, free-living caddisfly (*Costachorema*), stony-cased caddisfly (*Pycnocentrodes*), and cranefly (*Aphrophila*)]; and three 'tolerant' taxa [free-living caddisfly (*Aoteapsyche*) and midges (*Maoridiamesa* and orthoclads)]. Twelve of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised five 'highly sensitive' taxa, six 'moderately sensitive' taxa, and one 'tolerant' taxon. Nine of these taxa were also dominant in the summer community (Table 87) when three fewer 'sensitive' taxa were dominant. Very similar seasonal dominances resulted in a difference of only 0.3 unit in $SQMCI_s$ scores between spring and summer (Tables 162 and 163). The seven taxa recorded as very or extremely abundant during spring and/or summer have characterised this site's communities on 43% to 97% of past survey occasions.

3.2.19.2.3 Predicted stream 'health'

The Kaupokonui River site upstream of the Kaponga oxidation pond system is 9.2 km downstream of the National Park boundary at an altitude of 260 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 111 (altitude) and 107 (distance) for this site. The historical site median (116) is 5 units higher than the altitude prediction and 9 units higher than the distance predictive value. The spring, 2012 survey score (128 units) was significantly 17 to 21 units above predictive values and the summer, 2013 score (120 units) was 9 units above the predictive altitude value and a significant 13 units above the predictive distance value. Of the 32 surveys to date at this site, 22% of MCI scores have been less than 107 units while 69% have been greater than 111 units.

3.2.19.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Kaupokonui River upstream of the Kaponga oxidation ponds system. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 83.



N = 32
Kendall tau = +0.475
p level = 0.0001 [>FDR, p < 0.001]
Significant at p < 0.05 & p < 0.01
after FDR application

Figure 83 LOWESS trend plot of MCI data at the site upstream of the Kaponga oxidation ponds system

A very strong positive temporal trend in MCI scores has been found over the eighteen-year period which has been statistically significant at the 5% and 1% levels after FDR application. This has been more pronounced since 2001 but scores plateaued for about three years before a more recent gradual improvement. The wide range of LOWESS-smoothed range of scores (22 units) has ecological significance over the eighteen year period, particularly over the first nine years of the monitoring period, and may have been related partly to improved dairyshed wastes disposal consents' compliance reported in this catchment. Smoothed MCI scores consistently indicated 'good' generic river health (Table 1) with a tendency toward 'very good' more recently. In terms of predictive relationships (Table 2) for a site in the mid reaches of a ringplain river, health has been in the 'expected' throughout the first thirteen years of the period trending recently into the 'better than expected' category.

3.2.19.3 Site upstream of Kapuni railbridge (KPK000660)

3.2.19.3.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Kaupokonui River at this mid-reach site upstream of the Kapuni railbridge between 1995 and February 2012. These results are summarised in Table 88, together with the results from the current period, and illustrated in Figure 84.

Table 88 Results of previous surveys performed in the Kaupokonui River upstream of Kapuni railbridge, together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPK000660	34	15-32	24	71-122	102	26	128	27	115

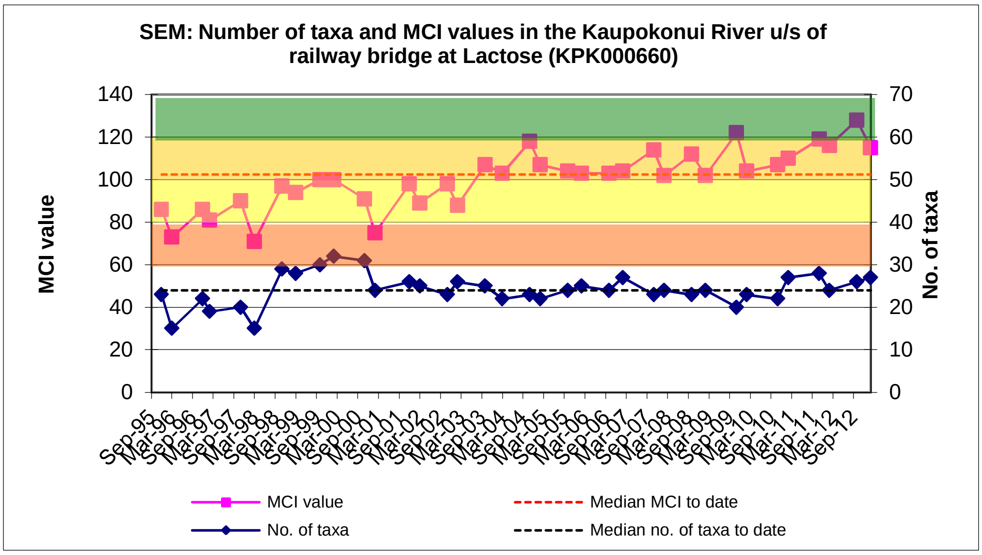


Figure 84 Numbers of taxa and MCI values in the Kaupokonui River upstream of Kapuni railbridge

A wide range of richnesses (15 to 32 taxa) has been found with a median richness of 24 taxa (more representative of typical richnesses in the mid reaches of ringplain streams and rivers). During the 2012-2013 period spring (26 taxa) and summer (27 taxa) richnesses were very similar and within three taxa of the median taxa number coincident with markedly less extensive (minimal) substrate periphyton cover recorded in both seasons.

MCI values have had a very wide range (51 units) at this site, much wider than typical of sites in the mid reaches of ringplain rivers. However, the median value (102 units) has been typical of mid reach sites elsewhere on the ringplain. The spring, 2012 (128 units) and summer, 2013 (115 units) scores were higher than typical of this site on both occasions, with the spring score 6 units higher than previously recorded at this site. These scores categorised this site as having 'very good' (spring) and 'good' (summer) health generically (Table 1) and, in terms of predictive relationships

(Table 2), 'better than expected' in spring and summer for the mid reaches of a ringplain river. The historical median score (102 units) placed this site in the 'good' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.19.3.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 89.

Table 89 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kaupokonui River upstream of Kapuni railbridge between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	8	24		A
ANNELIDA	Oligochaeta	1	18	53		
	Lumbricidae	5	1	3		
MOLLUSCA	*Potamopyrgus*	4	5	15		VA
CRUSTACEA	Ostracoda	1	1	3		
EPHEMEROPTERA	*Austroclima*	7	2	6		A
	Coloburiscus	7	16	47	VA	A
	Deleatidium	8	23	68	XA	XA
	Nesameletus	9	1	3	A	
PLECOPTERA	*Acroperla*	5	1	3		
HEMIPTERA	*Sigara*	3	1	3		
COLEOPTERA	Elmidae	6	27	79	VA	VA
	Hydraenidae	8	2	6		
MEGALOPTERA	*Archichauliodes*	7	14	41	A	VA
TRICHOPTERA	*Aoteapsyche*	4	17	50	XA	XA
	Costachorema	7	4	12		
	Hydrobiosis	5	14	41		
	Beraeoptera	8	2	6	VA	A
	Olinga	9	1	3	A	
	Oxyethira	2	6	18		
	Pycnocentrodes	5	7	21	VA	
DIPTERA	*Aphrophila*	5	18	53	VA	A
	Eriopterini	5	1	3		
	Chironomus	1	1	3		
	Maoridiamesa	3	20	59		
	Orthocladiinae	2	28	82		
	Tanytarsini	3	4	12		
	Empididae	3	2	6		
	Muscidae	3	2	6		
	Austrosimulium	3	5	15		

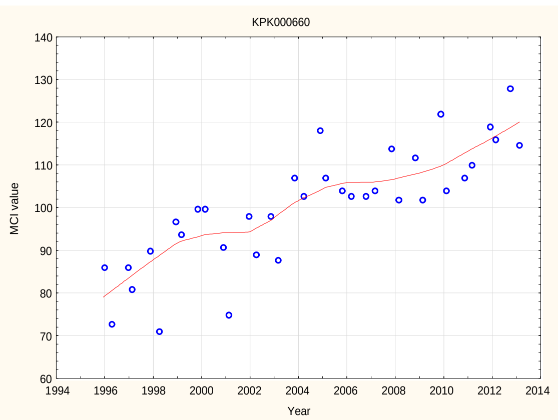
Prior to the current 2012-2013 period, 30 taxa had characterised the community at this site on occasions. These have comprised five 'highly sensitive', eleven 'moderately sensitive', and fourteen 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and a higher proportion of 'tolerant' taxa as might be expected in the mid reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; two 'moderately sensitive' taxa [elmid beetles and cranefly (*Aphrophila*)]; and four 'tolerant' taxa [oligochaete worms, free-living caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]. Ten of the historically characteristic taxa (four of which have been predominant) were dominant in the spring, 2012 community. These comprised four 'highly sensitive', five 'moderately sensitive', and one 'tolerant' taxa. Seven of these taxa plus one 'moderately sensitive' and two 'tolerant' taxa comprised the dominant taxa of the summer community. Therefore, seven of these thirteen taxa were dominant in both spring and summer communities (Table 89). The similarity in the two numerically dominant taxa in spring and summer in particular was reflected in the minimal difference (0.2 unit) in seasonal $SQMCI_s$ scores (Tables 162 and 163). The nine taxa recorded as very or extremely abundant during spring and/or summer have characterised this site's communities on 6% to 79% of past survey occasions.

3.2.19.3.3 Predicted stream 'health'

The Kaupokonui River site upstream of the Kapuni railbridge is 15.5 km downstream of the National Park boundary at an altitude of 170 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 102 (altitude) and 101 (distance) for this site. The historical site median (102) is equal with the altitude prediction and one unit above the distance predictive value. However, the spring, 2012 survey score (128 units) was significantly higher than both predictive values while the summer, 2013 score (115 units) was also significantly higher than both predictive values by 13 to 14 units. Of the 36 surveys to date at this site, 44% of MCI scores have been less than 101 units while 50% have been greater than 102 units.

3.2.19.3.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Kaupokonui River upstream of the Kapuni railbridge. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 85.



N = 36
Kendall tau = +0.659
p level < 0.0001 [>FDR, p = < 0.0001]
Significant at $p < 0.05$ and $p < 0.01$ (after FDR)

Figure 85 LOWESS trend plot of MCI data at the site upstream of Kapuni railbridge

A very strong, statistically significant temporal improvement in MCI scores has been found at this mid-catchment site. This trend has been similar to, but stronger than, that found at the nearest river site upstream and the very wide range of LOWESS-smoothed scores (40 units) has been ecologically very significant. Fonterra factory wastewater irrigation activities nearby in this catchment have been better managed during this period and surveillance monitoring has reported improved dairy shed waste treatment ponds systems compliance upstream of this site.

The trend in generic river health (Table 1) indicated by smoothed MCI scores, has moved from 'poor' to 'fair' during the first half of the period, improving to 'good' where it has remained since 2003. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a ringplain river, health has improved from the 'worse than expected' category (prior to 1999), through 'expected' (from 1997 to 2010), to the 'better than expected' category where it has remained since 2010.

3.2.19.4 Upper Glenn Road site (KPK000880)

3.2.19.4.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in the Kaupokonui River at this lower reach site at Upper Glenn Road between 1995 and February 2012. These results are summarised in Table 90, together with the results from the current period, and illustrated in Figure 86.

Table 90 Results of previous surveys performed in the Kaupokonui River at Upper Glenn Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPK000880	34	14-31	19	66-108	91	21	110	18	100

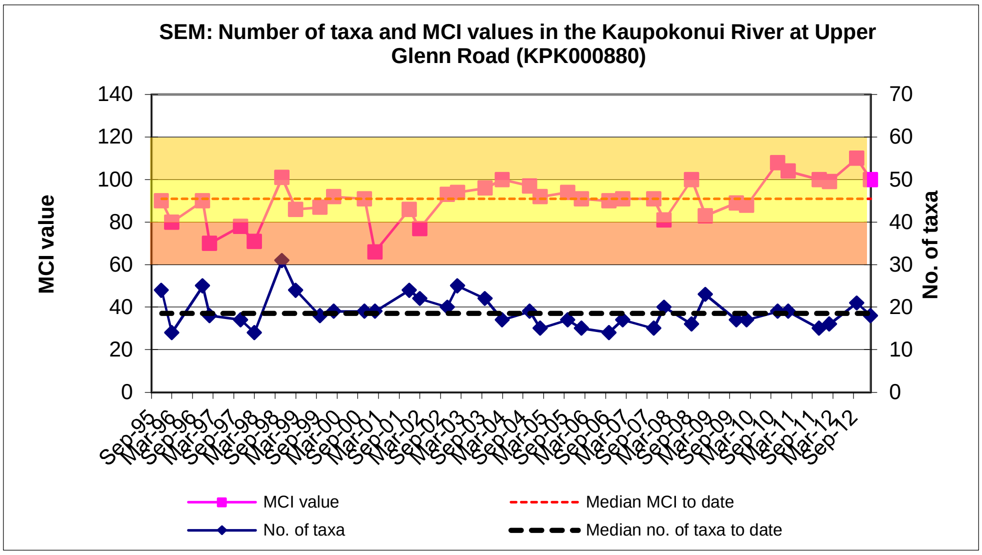


Figure 86 Numbers of taxa and MCI values in Kaupokonui River at Upper Glenn Road

A wide range of richnesses (14 to 31 taxa) has been found with a median richness of 19 taxa (typical of richnesses in the lower reaches of ringplain streams and rivers).

During the 2012-2013 period spring (21 taxa) and summer (18 taxa) richnesses were similar and within two taxa of the median taxa number.

MCI values have had a very wide range (42 units) at this site, typical of sites in the lower reaches of ringplain streams and rivers. The median value (91 units) has been typical of scores at lower reach sites elsewhere on the ringplain (TRC, 1999 (updated, 2013)). The spring, 2012 (110 units) and summer, 2013 (160 units) scores were significantly higher in spring and 9 units higher (summer) than the historical median score at this site with the spring score 2 units above the maximum previously recorded at this site, coincident with less extensive substrate periphyton cover than usual. These scores categorised this site has having 'good' health (spring and summer) generically (Table 1) and, in terms of predictive relationships (Table 2) 'better than expected (spring) and 'expected' (summer) health for the lower reaches of a ringplain river. The historical median score (91 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.19.4.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 91.

Table 91 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kaupokonui River at Upper Glenn Road between 1995 and February 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
PLATYHELMINTHES	*Cura*	3	1	3		
NEMERTEA	Nemertea	3	5	15		A
ANNELIDA	Oligochaeta	1	30	88		
MOLLUSCA	*Latia*	5	1	3		
	Physa	3	2	6		
	Potamopyrgus	4	10	29		A
CRUSTACEA	Ostracoda	1	1	3		
	Paracalliope	5	1	3		
EPHEMEROPTERA	*Austroclima*	7	0	0		A
	Coloburiscus	7	2	6	A	
	Deleatidium	8	17	50	XA	XA
	Nesameletus	9	1	3		
COLEOPTERA	Elmidae	6	27	79		A
MEGALOPTERA	*Archichauliodes*	7	2	6	A	A
TRICHOPTERA	*Aoteapsyche*	4	22	65	A	XA
	Costachorema	7	3	9		
	Hydrobiosis	5	18	53		A
	Oxyethira	2	6	18		
	Pycnocentrodes	5	14	41	A	A
DIPTERA	*Aphrophila*	5	6	18		
	Chironomus	1	1	3		
	Maoridiamesa	3	20	59	A	
	Orthocladiinae	2	31	91	A	VA
	Tanytarsini	3	5	15		
	Ephydridae	4	1	3		
	Muscidae	3	2	6		
	Austrosimulium	3	2	6		

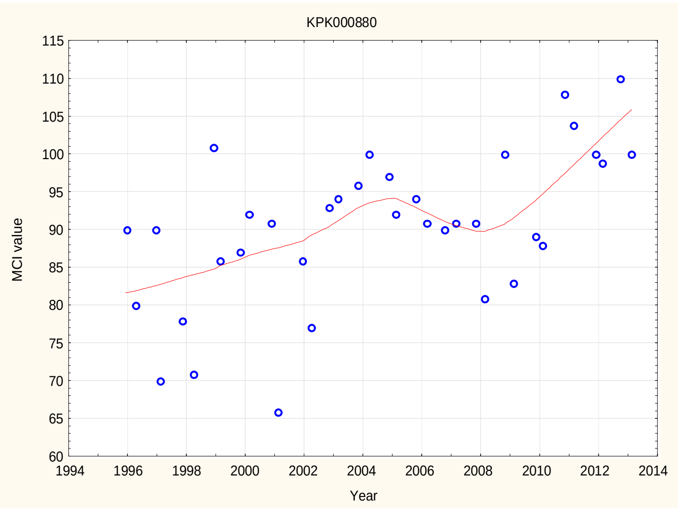
Prior to the current 2012-2013 period, 26 taxa had characterised the community at this site on occasions. These have comprised two 'highly sensitive', nine 'moderately sensitive', and fifteen 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and relatively high proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; two 'moderately sensitive' taxa [elmid beetles and caddisfly (*Hydrobiosis*)]; and four 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]. Seven of the historically characteristic taxa were dominant in the spring, 2012 community. These comprised one 'highly sensitive', three 'moderately sensitive', and three 'tolerant' taxa, whereas five of these taxa (one 'highly sensitive', two 'moderately sensitive', and two 'tolerant' taxa) together with two 'moderately sensitive' and two 'tolerant' additional taxa, comprised the dominant taxa of the summer, 2013 community. Five of these twelve taxa were dominant in both spring and summer communities (Table 91). The proportional increase in summer dominance by several 'tolerant' taxa in particular was reflected in the decrease of 1.7 units in seasonal $SQMCI_s$ scores (Tables 162 and 163). The three taxa recorded as very or extremely abundant during spring and/or summer have characterised this site's communities on 50% to 91% of past survey occasions.

3.2.19.4.3 Predicted stream 'health'

The Kaupokonui River site at Upper Glenn Road is 25.7 km downstream of the National Park boundary at an altitude of 60 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 91 (altitude) and 95 (distance) for this site. The historical site median (91) is equal with the altitude prediction and four units lower than the predictive distance value. The spring, 2012 score (110 units) was significantly higher (by 15 to 19 units) and summer, 2012 score (100 units) by 4 to 9 units above both predictive values. Of the 36 surveys to date at this site, 47% of MCI scores have been less than 91 units while only 31% have been greater than 95 units.

3.2.19.4.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Kaupokonui River at Upper Glenn Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 87.



N = 36
Kendall tau = +0.374
p level = 0.001 [>FDR, p = 0.004]
Significant at $p < 0.05$ and $p < 0.01$ after FDR application

Figure 87 LOWESS trend plot of MCI data at the Upper Glenn Road site

A temporal trend of improvement in MCI scores was found at this site up until 2005 followed by a gradual decline, before a more recent, steady improvement, with the overall trend statistically significant ($p < 0.05$ and $p < 0.01$ after FDR). The LOWESS-smoothed range of MCI scores (24 units) has been ecologically significant but nowhere near as wide as that upstream indicative of some decrease in effects in a downstream direction. The overall positive temporal trend was due to improved wastes management further upstream in the catchment but more particularly in relation to a reduction in heat input (via cooling water) to the river at the Fonterra, Kapuni factory.

Smoothed MCI scores have consistently indicated 'fair' generic river health (Table 1) through the period, until very recently when it has reached 'good' health, although prior to 2003 individual scores varied between 'fair' and 'poor' health. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain river, river health has improved from 'worse than expected' (prior to 1998) to 'expected' where it remained until most recently entering the 'better than expected' category.

3.2.19.5 Kaupokonui Beach site (KPK000990)

3.2.19.5.1 Taxa richness and MCI

Twenty-six surveys have been undertaken in the Kaupokonui River at this lower reach site at Kaupokonui Beach between 1999 and February 2012. These results are summarised in Table 92, together with the results from the current period, and illustrated in Figure 88.

Table 92 Results of previous surveys performed in the Kaupokonui River at the Kaupokonui Beach site, together with spring 2012 and summer 2013 results

Site code	SEM data (1999 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPK000990	26	11-26	19	69-98	91	21	103	17	92

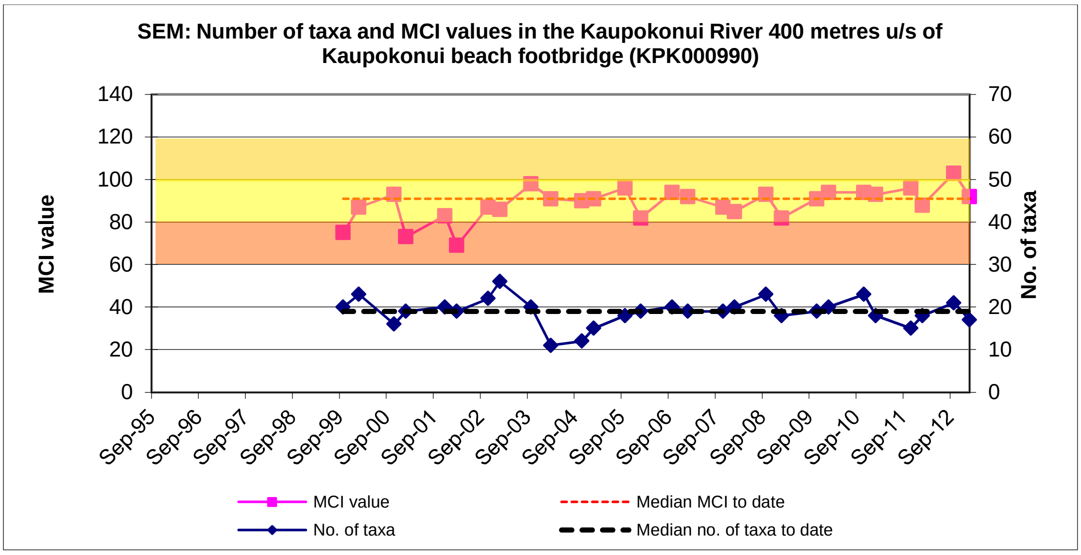


Figure 88 Numbers of taxa and MCI values in the Kaupokonui River at the Kaupokonui Beach site

A wide range of richnesses (11 to 26 taxa) has been found, with a median richness of 19 taxa. During the 2012-2013 period spring (21 taxa) and summer (17 taxa) richnesses were relatively similar and within two taxa of the median richness.

MCI values have had a moderate range (29 units) at this site, typical of sites in the lower reaches of ringplain streams and rivers. The median value (91 units) has been relatively typical of scores at lower reach sites elsewhere on the ringplain (TRC, 1999 (updated, 2013)). The spring, 2012 (103 units) score was 5 units higher than previously recorded at this site while the summer, 2013 (92 units) score was typical for such a site and one unit above the historical median. These scores categorised this site as having 'good health generically (Table 1) in spring and 'fair' health in summer, and in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower reaches of a ringplain river. The historical median score (91 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.19.5.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 93.

Prior to the current 2012-2013 period, 15 taxa had characterised the community at this site on occasions. These have comprised one 'highly sensitive', six 'moderately sensitive', and eight 'tolerant' taxa i.e. a very low proportion of 'highly sensitive' taxa and a higher proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; three 'moderately sensitive' taxa [elmid beetles, free-living caddisfly (*Hydrobiosis*), and stony-cased caddisfly (*Pycnocentrodes*)]; and four 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]

Table 93 Characteristic taxa (abundant, very abundant,, extremely abundant) recorded in the Kaupokonui River at the Kaupokonui Beach site between 1999 and February 2012 [26 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	3	12		
ANNELIDA	Oligochaeta	1	25	96		VA
MOLLUSCA	*Potamopyrgus*	4	12	46		A
EPHEMEROPTERA	*Austroclima*	7	1	4		A
	Coloburiscus	7	0	0	A	
	Deleatidium	8	18	69	XA	
COLEOPTERA	Elmidae	6	18	69		
TRICHOPTERA	*Aoteapsyche*	4	17	65	VA	XA
	Costachorema	7	3	12		
	Hydrobiosis	5	17	65		A
	Pycnocentrodes	5	16	62	VA	
DIPTERA	*Aphrophila*	5	1	4	A	
	Maoridiamesa	3	18	69	VA	
	Orthocladiinae	2	24	92	A	VA
	Tanytarsini	3	7	27		
	Muscidae	3	1	4		

Six of the historically characteristic taxa were dominant in the spring, 2012 community, five of which have been predominant to date together with one 'moderately sensitive' taxon [mayfly (*Coloburiscus*)] not previously found in abundance at this site. These comprised one 'highly sensitive' taxon, three 'moderately sensitive' taxa, and three 'tolerant' taxa, whereas only two of these taxa plus two additional 'moderately sensitive' and two additional 'tolerant' taxa comprised the dominant taxa of the summer, 2013 community. Therefore, only two of these eleven taxa were dominant in both spring and summer communities (Table

93) with a significant decrease in numerical abundances in four 'sensitive' taxa and the increase in some 'tolerant' taxa abundances were reflected in the significant summer decrease of 2.8 units in $SQMCI_s$ score (Tables 162 and 163).

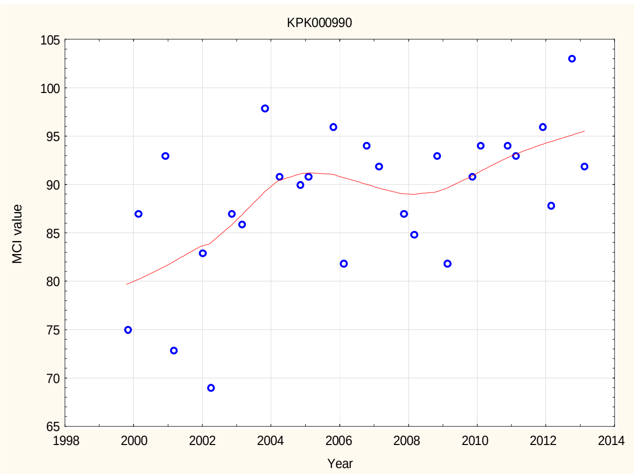
The six taxa recorded as very or extremely abundant during spring and/or summer have characterised this site's communities on 62% to 96% of past survey occasions.

3.2.19.5.3 Predicted stream 'health'

The Kaupokonui River at the Kaupokonui Beach site is 31.1 km downstream of the National Park boundary at an altitude of 5 m asl. Relationships for ringplain streams and rivers developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 85 (altitude) and 93 (distance) for this site. The historical site median (91) is 6 units higher than the altitude and 2 units below the distance predictive values. The spring 2012 survey score (103 units) was 10 to a significant 18 units above the predictive values while the summer score (92 units) was 7 units higher than the predictive altitude value and one unit below the distance value. Of the 28 surveys to date at this site, 21% of MCI scores have been less than 85 units while only 25% have been greater than 93 units.

3.2.19.5.4 Temporal trends in 1999 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the fourteen years of SEM results collected to date from the site in the Kaupokonui River at Kaupokonui Beach. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 89.



N = 28
Kendall tau = +0.354
p level = 0.008 [>FDR, p = 0.018]
Significant at p < 0.05 and p < 0.01
N/S after FDR at p < 0.01

Figure 89 LOWESS trend plot of MCI data at the Kaupokonui Beach site

Although the fourteen year trend in MCI scores has indicated improvement, the overall temporal trend has not been statistically significant (p > 0.01 after FDR), unlike trends further upstream which have a slightly (four year) longer monitoring period, with poorer 'health' prior to 1999. However, an ecologically significant range of LOWESS-smoothed scores (15 units) has been recorded, much narrower than ranges at the two nearest upstream sites, possibly reflecting certain upstream improvements in waste disposal management (documented earlier).

Individual MCI scores have been indicative of generic river health (Table 1) varying between 'poor' and 'fair' prior to 2003 improving to 'fair' where scores have remained consistently since this date. LOWESS-smoothed scores have been indicative of 'fair' generic river health throughout the period. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, river health has remained in the 'expected' category over the entire period, after bordering on the 'worse than expected' category early in the period.

3.2.19.6 Discussion

Seasonal MCI values typically decreased between spring and summer at all sites, from the Opunake Road site (by 3 units) through the Kaponga WWTP site (8 units), the Kapuni railbridge site (13 units), the Upper Glen Road site (10 units), to the Kaupokonui Beach site (11 units). These seasonal decreases may be compared with historical seasonal median decreases of 2, 9, 2, 2, and 5 units respectively (Appendix II). Seasonal communities shared 52% of 31 taxa at the Opunake Road upper mid-reach site, 68% of 25 taxa at the Kaponga mid-reach site, 61% of 33 taxa at the Kapuni Railbridge mid-reach site, 63% of 24 taxa at the Upper Glenn Road lower reach site, and 52% of 25 taxa at the furthest downstream site (Kaupokonui Beach) in the lower reaches. Seasonal community compositions have generally been more variable with increasing distance downstream from the National Park, particularly nearer the coast, but during the 2012-2013 monitoring period greater variability occurred in the upper mid-reaches and near the coast, but there was similar seasonal variability than usual at three (mainly mid-reach) sites..

Community composition varied markedly through the length of the river surveyed. A total of 40 taxa was recorded in spring of which only nine taxa were present at all five sites. These included two 'highly sensitive' taxa, five 'moderately sensitive' taxa, and two 'tolerant' taxa with only the 'highly sensitive' mayfly (*Deleatidium*), 'moderately sensitive' mayfly (*Coloburiscus*), and 'tolerant' caddisfly (*Aoteapsyche*) abundant at all five sites. A lower number of taxa (35) was found along the river's length by the summer survey of which nine taxa were present at all five sites. These were very similar to the ten widespread taxa in spring with the overall loss of one 'highly sensitive' taxon and addition of one 'moderately sensitive taxon'. Only the one 'tolerant' net-building caddisfly was abundant at all five sites. These dissimilarities in spatial community structure along the length of the Kaupokonui River were slightly more pronounced in spring than in summer.

The MCI scores fell in a downstream direction between the upper site and the furthest downstream lower reaches site by 25 units in spring and 33 units in summer, over a river distance of 27.8 km. These seasonal falls in MCI scores equated to rates of decline of 0.9 unit/km (spring) and 1.2 units/km (summer), compared with a predicted rate of 0.9 unit/km for the equivalent length and reach of a National Park-sourced river (Stark and Fowles, 2009). Although small, this was typical of the trend of most past summers' seasonal increases in rates of decline. This may be compared with a rate of 0.6 unit/km derived for the nearby Waingongoro River (over the 18 year period) which, although a ringplain National Park-sourced river, has an atypical meandering nature over more than twice the length of the Kaupokonui River.

Between the upper mid-reach (Opunake Road) site and Kapuni mid-reach site, the spring (0 units/km) and summer (0.8 unit/km) rates of decline were dissimilar and much lower than the predicted rate (1.4 units/km) for the equivalent river reach. For the mid-reach Kapuni site to Kaupokonui Beach lower reach site, spring (1.6 units/km) and summer (1.5 units/km) rates of decline were very similar with both rates well above the predicted rate of 0.5 unit/km.

Using the long-term median SEM MCI scores for each site (Appendix II), the rates of decline between upper mid-reach site (Opunake Road) and mid-catchment (Kapuni) site, and between this mid-catchment site and lower river site at Kaupokonui Beach, have been about 2.3 and 0.6 units per km respectively with an overall rate of decline of 1.4 MCI units/km over the river's length. Spring and summer overall rates of decline have been 1.3 and 1.4 units per kilometre. Therefore overall rates of decline over the 2012-2013 period were slightly lower in both spring (by 0.4 unit/km) and summer (by 0.2 unit/km) than the median rates prior to 2013.

3.2.20 Katikara Stream

Two sites in the Katikara Stream, one located near the headwaters (just inside the National Park) and the other near the coast, were included in the SEM programme in the 2000-2001 year, for the purpose of long term monitoring of the impacts of riparian vegetation planting initiatives within this north-western Taranaki catchment. In the 2008-2009 period severe headwater erosion events impacted upon the macroinvertebrate communities of the upper reaches of this stream (TRC, 2009). The results found in the 2012-2013 surveys are presented in Tables 164 and 165, Appendix I.

3.2.20.1 Carrington Road site (KTK000150)

3.2.20.1.1 Taxa richness and MCI

Twenty-six surveys have been undertaken at this upper reach site in the Katikara Stream inside the National park boundary at Carrington Road between 1999 and March 2012. These results are summarised in Table 94, together with the results from the current period, and illustrated in Figure 90.

Table 94 Results of previous surveys performed in the Katikara Stream at Carrington Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to March 2012)						2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values			Nov 2012		Mar 2013	
		Range	Median	Range	Median		Taxa no	MCI	Taxa no	MCI
KTK000150	26	11-38	30	112-148	137		18	143	23	124

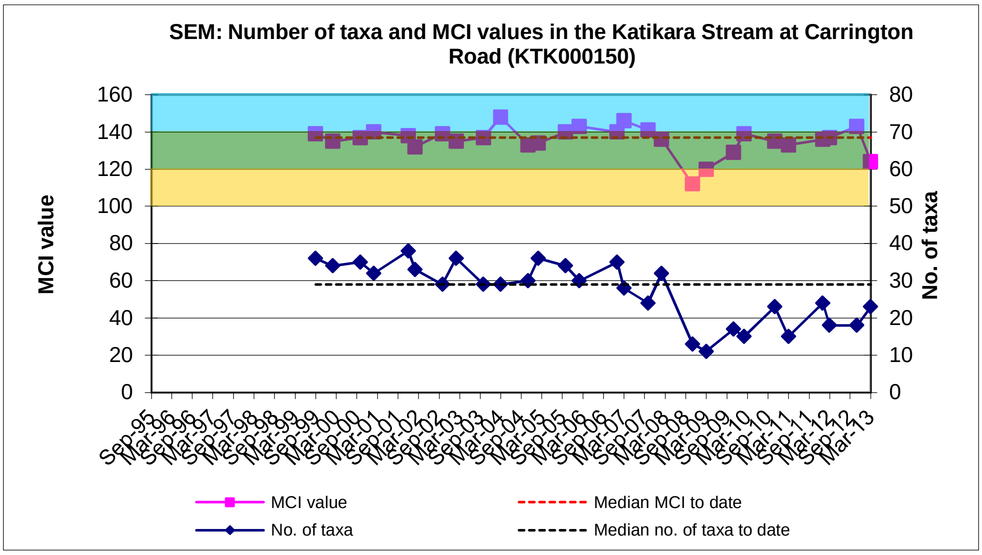


Figure 90 Numbers of taxa and MCI values in the Katikara Stream at Carrington Road

A very wide range of richnesses (11 to 38 taxa) has been found; wider than might be expected, due to significant headwater erosion over the 2008-2009 period and subsequent effects, with a median richness of 30 taxa (far more representative of typical richnesses in ringplain streams and rivers near the National Park boundary). However, median richness since 2008-2009 has been 18 taxa. During the 2012-2013 period spring (18 taxa) and summer (23 taxa) richnesses were well below this median richness indicative of a continuing post-headwater erosion recovery phase and/or long term degradation of the physical habitat.

MCI values at this site have had a wider range (36 units) than typical of a National Park boundary site, due in part to atypically lower values for a short period and on other occasions after the 2008-2009 headwater erosion event. The median value (137 units) has been typical of upper reach sites (near or within the National Park) elsewhere on the ringplain however, whereas the spring, 2012 (143 units) score was relatively typical of the range for such a site, and six units above of the historical median. The summer score (124 units) was a significant (Stark, 1998) 13 units lower than the historical median, coincident with extremely low flow conditions and limited riffle 'habitat' at this site. These scores categorised this site as having 'excellent' (spring) and 'very good' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' health for the upper reaches of a ringplain stream in spring and 'expected' health in summer although taxa richnesses were not indicative of typical pre-erosion community compositions. The historical median score (137 units) placed this site in the 'very good' category for the generic, and 'expected' category for the predictive methods of assessment.

3.2.20.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 95.

Table 95 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Katikara Stream at Carrington Road between 1999 and March 2012 [26 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
EPHEMEROPTERA	*Ameletopsis*	10	1	4		
	Austroclima	7	15	58		
	Coloburiscus	7	19	73		A
	Deleatidium	8	23	88	A	A
	Nesameletus	9	17	65		A
PLECOPTERA	*Acroperla*	5	2	8		
	Austroperla	9	6	23		
	Zelandobius	5	18	69		
	Zelandoperla	8	13	50	A	
COLEOPTERA	Elmidae	6	7	27		
MEGALOPTERA	*Archichauliodes*	7	1	4		A
TRICHOPTERA	*Costachorema*	7	1	4		
	Hydrobiosis	5	1	4		
	Hydrobiosella	9	7	27		
	Orthopsyche	9	8	31		
	Beraeoptera	8	1	4		
	Oxyethira	2	0	0		A
DIPTERA	*Aphrophila*	5	5	19		
	Orthocladiinae	2	14	54		
	Polypedilum	3	1	4		

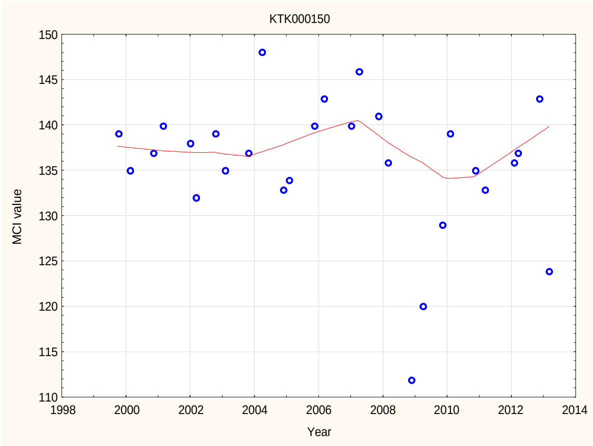
Prior to the current 2012-2013 period, 19 taxa had characterised the community at this site on occasions. These have comprised eight 'highly sensitive', nine 'moderately sensitive', and two 'tolerant' taxa i.e a majority of 'sensitive' taxa as would be expected near the National Park boundary of a ringplain stream. Predominant taxa have included three 'highly sensitive' taxa [mayflies (*Deleatidium* and *Nesameletus*) and stonefly (*Zelandoperla*)]; three 'moderately sensitive' taxa [mayflies (*Coloburiscus* and *Austroclima*), and stonefly (*Zelandobius*)]; and one 'tolerant' taxon [orthoclad midges]. Only two of these characteristic taxa (both 'highly sensitive' taxa) were dominant in the spring, 2012 community. One of these taxa was again dominant in the summer, 2013 community together with one other 'highly sensitive' and two other 'moderately sensitive' historically characteristic taxa of this site together with one 'tolerant' taxon [algal-piercing caddisfly (*Oxyethira*)] previously not characteristic at this site. Mainly 'sensitive' taxa were dominant coincident with minimal periphyton substrate cover at this site on both occasions. The presence of one 'tolerant' taxon in the summer characteristic community composition was reflected in the decrease of 0.9 unit in the summer $SQMCI_s$ value (Tables 164 and 165). No taxa were recorded as very or extremely abundant at the time of either of the surveys.

3.2.20.1.3 Predicted stream 'health'

The Katikara Stream at Carrington Road is within the National Park boundary at an altitude of 420 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009) predict MCI values of 127 (altitude) and 132 (distance) for this site. The historical site median (137 units) is 10 units higher than the altitude prediction and 5 units higher than the distance predictive value. The spring (143 units) score was significantly higher than both predictive scores while the summer (124 units) score was within 3 to 8 units lower than both predictive scores. Of the 28 surveys to date at this site, only 11% of MCI scores have been less than 127 units while 86% have been greater than 132 units.

3.2.20.1.4 Temporal trends in 1999 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the fourteen years of SEM results collected to date from the site in the Katikara Stream at Carrington Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 91.



N = 28
Kendall tau = -0.105
p level = 0.432 [>FDR, p = 0.535]
N/S at p < 0.05

Figure 91 LOWESS trend plot of MCI data at the Carrington Road site

Relatively stable MCI scores over the first four years of the period at this pristine site inside the National Park were followed by a very gradual rise. The more recent downward trend has been due to significant headwater erosion effects during 2008, with some recovery after 2010 However, the overall temporal trend of deterioration has not been statistically significant and the range of LOWESS-smoothed scores (7 units) over the period has not been ecologically significant, the range having widened since the erosion event. Smoothed scores have been indicative of 'very good' generic stream health (Table 1) throughout the period and, in terms of predictive relationships (Table 2) for a site in the upper reaches of a ringplain stream at the boundary of the National Park, stream health has been in the 'expected' category throughout the period, bordering on the 'better than expected' category just prior to the headwater erosion impacts during 2008.

3.2.20.2 Coastal site (KTK000248)

3.2.20.2.1 Taxa richness and MCI

Twenty-five surveys have been undertaken in the Katikara Stream at this lower reach site near the coast between 2000 and March 2012. These results are summarised in Table 96, together with the results from the current period, and illustrated in Figure 92.

Table 96 Results of previous surveys performed in the Katikara Stream near the coast together with spring 2012 and summer 2013

Site code	SEM data (1995 to March 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Nov 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KTK000248	25	20-31	26	81-116	102	22	118	26	99

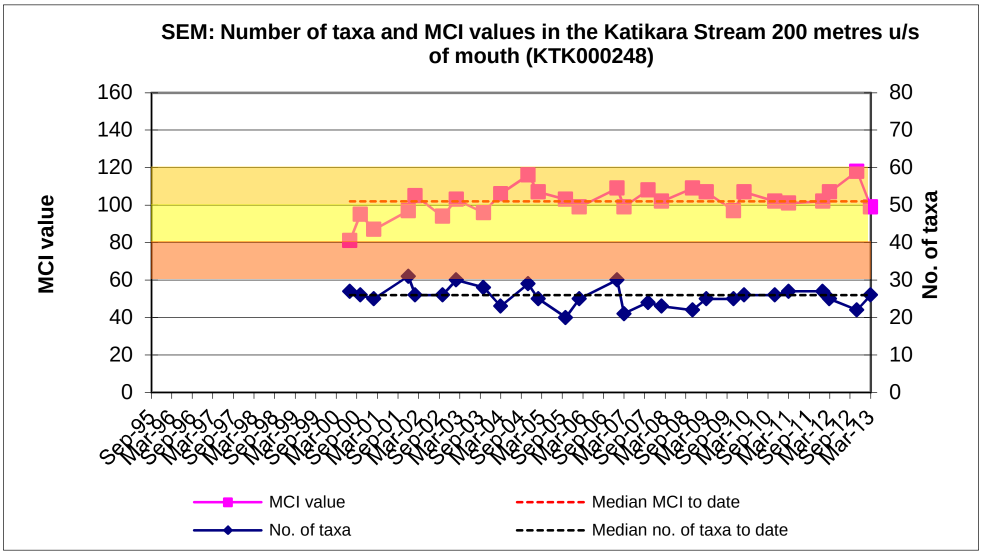


Figure 92 Numbers of taxa and MCI values in the Katikara Stream 200m u/s of the coast

A moderate range of richnesses (20 to 31 taxa) has been found with no indication of the effects of headwater erosion events that have been noted at the upstream site. The median richness of 26 taxa has been more representative of typical richnesses in the lower reaches of ringplain streams and rivers. During the 2012-2013 period, spring (22 taxa) and summer (26 taxa) richnesses were relatively similar and within four taxa of the median taxa number.

MCI values have had a relatively wide range (35 units) at this site, typical of sites in the lower reaches of ringplain streams. The median value (102 units) has been higher than typical of lower reach sites elsewhere on the ringplain however, with the spring, 2012 (118 units) score well above those typical for such a site and a significant (Stark, 1998) 16 units higher than the historical median and also 2 units above the previous maximum score. The summer score (99 units) was within 3 units of the historical median coincident with some increase in periphyton substrate cover. These scores categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower reaches of a ringplain stream. The historical median score (102 units) placed this site in the 'good' category for the generic and 'expected' category for the predictive methods of assessment.

3.2.20.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 97.

Table 97 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Katikara Stream near the mouth between 2000 and March 2012 [25 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	4	16		
ANNELIDA	Oligochaeta	1	17	68		
MOLLUSCA	*Latia*	5	2	8		
	Potamopyrgus	4	22	88		VA
CRUSTACEA	*Paratya*	3	2	8		
EPHEMEROPTERA	*Austroclima*	7	15	60		
	Coloburiscus	7	11	44	A	A
	Deleatidium	8	16	64	XA	VA
	Rallidens	9	1	4		
PLECOPTERA	*Zelandobius*	5	1	4		
	Zelandoperla	8	1	4		
COLEOPTERA	Elmidae	6	21	84	VA	VA
	Ptilodactylidae	8	2	8		
MEGALOPTERA	*Archichauliodes*	7	12	48	A	A
TRICHOPTERA	*Aoteapsyche*	4	22	88	VA	VA
	Costachorema	7	7	28		
	Hydrobiosis	5	16	64	A	VA
	Pycnocentrodes	5	21	84	A	
DIPTERA	*Aphrophila*	5	16	64	A	VA
	Maoridiamesa	3	8	32		
	Orthocladiinae	2	19	76		VA
	Tanytarsini	3	3	12		
	Austrosimulium	3	7	28		VA

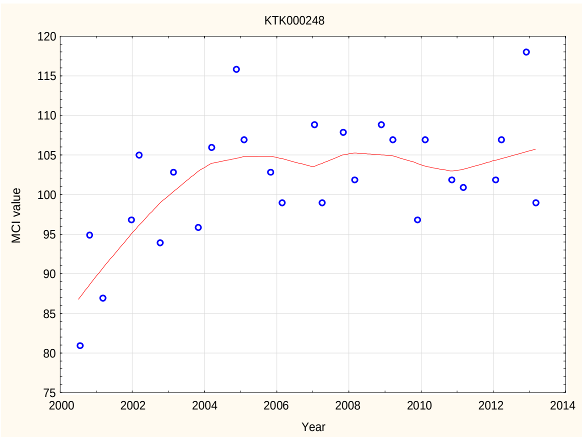
Prior to the current 2012-2013 period, 23 taxa had characterised the community at this site on occasions. These have comprised four 'highly sensitive', ten 'moderately sensitive', and nine 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and an increased proportion of 'tolerant' taxa as would be expected in the lower reaches of a ringplain stream. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; five 'moderately sensitive' taxa [mayfly (*Austroclima*), elmid beetles, free-living caddisfly (*Hydrobiosis*), stony-cased caddisfly (*Pycnocentrodes*), and cranefly (*Aphrophila*)]; and four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and orthoclad midges]. Eight of the historically characteristic taxa were dominant in the spring, 2012 community comprising one 'highly sensitive', six 'moderately sensitive', and one 'tolerant' taxa, whereas seven of the taxa plus another three 'tolerant' taxa comprised the dominant taxa of the summer community. Seven of these 11 taxa were dominant in both spring and summer communities (Table 97) but an increase in numerical abundances in some 'tolerant' taxa and a decrease in abundance within one 'highly sensitive' taxon were reflected in the seasonal decrease in $SQMCI_s$ scores (Table 164 and 165) which were lower by 2.2 units in summer. All taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 28% to 87% of past survey occasions.

3.2.20.2.3 Predicted stream 'health'

The Katikara Stream at the site near the coast is 18.1 km downstream of the National Park boundary at an altitude of 5 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 85 (altitude) and 99 (distance) for this site. The historical site median (102) is a significant 17 units higher than the altitude prediction and 3 units higher than the distance predictive value. The spring survey score (118 units) was significantly higher than both the predictive values while the summer score (99 units) was also significantly higher than the predictive altitude value but equal with the predictive distance value. Of the 27 surveys to date at this site, only 4% of MCI scores have been less than 85 units while 63% have been greater than 99 units, confirmation of the 'better than expected' stream health at this site.

3.2.20.2.4 Temporal trends in 2000 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the thirteen years of SEM results collected to date from the site in the Katikara Stream near the coast. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 93.



N = 27
Kendall tau = +0.309
p level = 0.024 [>FDR, p = 0.048]
Significant at p <0.05;
N/S at p < 0.01 or after FDR application

Figure 93 LOWESS trend plot of MCI data at the coastal site

A relatively strong temporal improvement in MCI scores has been recorded, particularly during the first five years of the thirteen year monitoring period. This trend has levelled off over the most recent eight year period with a minimal trend very recently. Whereas previously the overall trend was statistically significant, it now has statistical significance after FDR application at p > 0.05 but not at p < 0.01. The wide range of LOWESS-smoothed MCI scores (19 units) has particular ecological significance and has occurred coincidentally with retirement and riparian planting of the margins of the lower reaches of this stream.

Smoothed MCI scores indicative of 'fair' generic stream health (Table 1) have improved to 'good' health after 2003 where they have remained (Figure 93). In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, stream health has improved from ‘expected’ to ‘better than expected’ category where it has remained since 2003.

3.2.20.3 Discussion

Atypically seasonal MCI values decreased significantly between spring and summer (19 units difference) at the National Park site but more typically decreased (also by 19 units) in summer at the coastal site with the percentage composition of ‘tolerant’ taxa increasing (by up to 23%) in the summer communities. Seasonal median scores (Appendix II) have remained very similar at the National Park site (within 2 units) and at the coastal site (within 1 unit); an atypical seasonal trend compared with lower reach sites elsewhere on the ringplain. Seasonal communities at the upper site shared only 13 common taxa (46% of the 28 taxa found at this site in 2012-2013) compared with 17 shared common taxa (55% of the 31) at the lower reaches site near the coast; an atypically more pronounced seasonal change in community composition at the upstream site coincidental with extremely low flow summer conditions. The two sites shared only 10 common taxa (33% of the 30 taxa found at upper and lower reach sites) in spring and only 12 common taxa (32% of 37 taxa) in summer, indicative of little change in dissimilarities in spatial community structures in spring and in summer.

MCI score typically fell in a downstream direction in spring (by 25 units) and identically in summer (by 25 units), over a stream distance of 18.1 km downstream from the National Park boundary. These falls equated to rates of decline of 1.4 units/km in spring and in summer, slightly lower than the predicted rate of 1.8 units/km over the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009).

Using the long-term median SEM MCI scores for both sites (Appendix II), the rate of decline between the upper reach site and lower reach site near the coast has been 1.9 MCI units/km over the surveyed length. Therefore rates of MCI decline over the 2012-2013 period were lower in spring and in summer than the long term median rate for the 1995 to 2012 period.

3.2.21 Kapoaiaia Stream

Three established sites in the Kapoaiaia Stream, located at Wiremu Road (in open farmland nearly 6 km below the National Park boundary), Wataroa Road bridge (nearly 8 km further downstream), and about 0.8 km from the coast (in open farmland about 8 km further downstream, i.e. 25 km below the National Park boundary), were included in the SEM programme commencing in the 2000-2001 year. This stream was selected for the purpose of monitoring a western Taranaki ringplain catchment with minimal existing riparian vegetation cover. Biological sampling had been undertaken previously in this catchment as a component of the Taranaki ringplain survey (TCC, 1984) and on various occasions in relation to the periodic operation of the Pungarehu Dairy Factory. This factory has been closed to dairy operations since 1995.

The results of the spring, 2012 and summer, 2012-2013 surveys are presented in Tables 166 and 167 Appendix I.

3.2.21.1 Wiremu Road site (KPA000250)

3.2.21.1.1 Taxa richness and MCI

Twenty-six surveys have been undertaken in the Kapoaiaia Stream between 1995 and February 2012 at this open, upper mid-reach site in farmland, 5.7 km downstream of the National Park. These results are summarised in Table 98 together with the results from the current period, and illustrated in Figure 94.

Table 98 Results of previous surveys performed in the Kapoaiaia Stream at Wiremu Road together with the spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPA000250	26	19-31	25	83-130	109	25	130	28	126

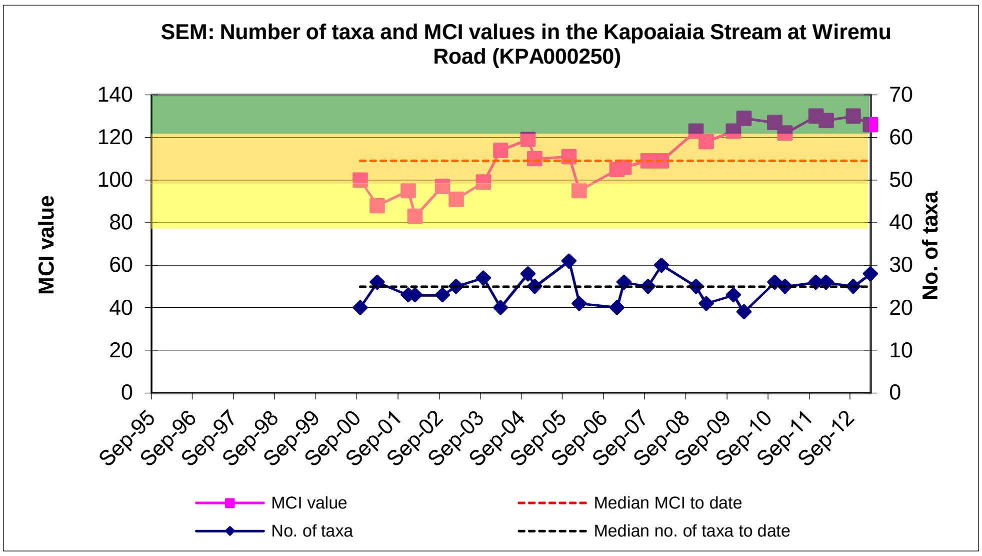


Figure 94 Numbers of taxa MCI values in the Kapoaiaia Stream at Wiremu Road

A moderate range of richnesses (19 to 31 taxa) has been found with a median richness of 25 taxa (more typical of richnesses in the mid-reaches of ringplain streams and rivers). During the 2012-2013 period, spring (25 taxa) and summer (28 taxa) richnesses were similar and within three taxa of this median richness.

MCI values have had a relatively wide range (47 units) at this site, wider than typical of a site in the upper mid-reaches of a ringplain stream although this site is in a reach of very open farmland. The median value (109 units) has been lower than typical of upper mid-reach sites elsewhere on the ringplain. The spring, 2012 (130 units) and summer, 2013 (126 units) scores were significantly 21 and 17 units above the historical median respectively and in spring the score was equal with the historical maximum recorded (in the previous spring) at this site. These scores categorised this site as having 'very good' generic health (Table 1) in spring and summer and, in terms of predictive relationships (Table 2), 'well above expected' health for the mid-reaches of a ringplain stream on both of these occasions. The historical median score (109 units) placed this site in the 'good' and 'expected' categories for the generic and predictive methods of assessment respectively.

3.2.21.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 99.

Table 99 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kapoaiaia Stream at Wiremu Road between 1995 and February 2012 [26 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	12	46		
MOLLUSCA	*Potamopyrgus*	4	3	12		
CRUSTACEA	*Paracalliope*	5	1	4		
EPHEMEROPTERA	*Austroclima*	7	5	19		
	Coloburiscus	7	16	62	VA	A
	Deleatidium	8	18	69	XA	XA
	Nesameletus	9	7	27	A	A
PLECOPTERA	*Acroperla*	5	5	19	A	
	Zelandoperla	8	8	31	VA	
COLEOPTERA	Elmidae	6	24	92	A	XA
	Hydraenidae	8	1	4		
MEGALOPTERA	*Archichauliodes*	7	3	12		A
TRICHOPTERA	*Aoteapsyche*	4	22	85		VA
	Costachorema	7	16	62		
	Hydrobiosis	5	9	35		
	Beraeoptera	8	4	15	VA	
	Helicopsyche	10	0	0	A	
	Olinga	9	1	4		
	Oxyethira	2	4	15		
	Pycnocentrodes	5	6	23	VA	
DIPTERA	*Aphrophila*	5	20	77		A
	Eriopterini	5	1	4		
	Maoridiamesa	3	20	77		A
	Orthocladiinae	2	22	85		A
	Tanytarsini	3	2	8		
	Muscidae	3	3	12		
	Austrosimulium	3	5	19		

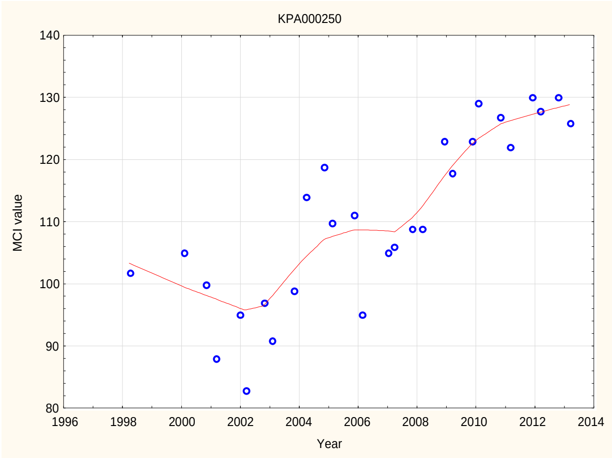
Prior to the current 2012-2012 period, 26 taxa have characterised the community at this site on occasions. These have comprised six 'highly sensitive', eleven 'moderately sensitive', and nine 'tolerant' taxa i.e. a predominance of 'sensitive' taxa as would be expected in the upper mid-reaches of a ringplain stream but a relatively higher number of 'tolerant' taxa for a site within 6km of the National Park boundary. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; four 'moderately sensitive' taxa [mayfly (*Coloburiscus*), elmid beetles, free-living caddisfly (*Costachorema*), and cranefly (*Aphrophila*)]; and three 'tolerant' taxa [net-building caddisfly (*Aoteapsyche*), and midges (orthoclads and *Maoridiamesa*)]. Three of these predominant taxa were dominant in the spring, 2012 community together with three other 'highly sensitive' taxa, two other 'moderately sensitive' taxa, and another 'highly sensitive' taxon [spiral-cased caddisfly (*Helicopsyche*)] not previously characteristic at this site. Some of these taxa, previously found only in low numbers at this site, were abundant coincident with a very sparse periphyton substrate cover which seldom had been a feature of this habitat previously. The summer, 2012 community was characterised by only four of the taxa dominant in spring, together with an additional two 'moderately sensitive' and three 'tolerant' taxa all of which had been characteristic of this site's communities previously (Table 99). Differences between the seasonally most dominant taxa compositions and some increase in 'tolerant' taxa abundances in summer resulted in a 0.9 unit $SQMCI_s$ decrease between spring and summer scores (Tables 166 and 167). Taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 15% to 92% of the past surveys.

3.2.21.1.3 Predicted stream 'health'

The Kapoaiaia Stream site at Wiremu Road is 5.7 km downstream of the National Park boundary at an altitude of 240 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 109 (altitude) and 112 (distance) for this site. The historical site median (109 units) is equal with the altitude prediction and 3 units lower than the distance predictive values. However, the spring, 2012 survey score (130 units) was significantly (Stark, 1998) higher than both predictive values while the summer, 2013 score (126 units) was also significantly higher (by 14 to 17 units) than both predictive values. Of the 28 surveys to date at this site, 43% of MCI scores have been less than 109 units while 43% have been greater than 112 units indicating that scores in the 2012-2013 period were better than most previous scores.

3.2.21.1.4 Temporal trends 1995 to 2013

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the SEM results collected to date from the site in the Kapoaiaia Stream at Wiremu Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 95.



N = 28
Kendall tau = +0.636
p level < 0.0001 [>FDR, p < 0.001]
Significant at p <0.05 and after FDR application

Figure 95 LOWESS trend plot of MCI data at the Wiremu Road site

A statistically significant temporal trend of improvement in MCI scores has been found over the fifteen year duration of this monitoring period (particularly over the period since 2003) and this trend was also significant at p<0.01 after FDR application. There has been an ecologically significant variability in the wide (32 units) range of LOWESS-smoothed scores at this site also. This appears to have been related to farming practices, particularly variations in fertiliser usage, through the open reach between the National Park boundary and this upper site, which may have been exacerbated by the lack of riparian vegetation along this reach.

Smoothed MCI scores, indicative of generic stream health (Table 1) varying between 'fair' and 'very good' have been slightly lower than might be expected at times (particularly prior to 2004) at this site approximately 6 km below the National Park. A strong improvement has been obvious since 2007 which has been maintained over the 2012-2013 period.

In terms of predictive relationships (Table 2) for a site in the upper mid-reaches of a ringplain stream, stream health had been mainly in the 'worse than expected' category prior to 2004 improving to 'expected' until 2010 and most recently to the 'better than expected' category.

3.2.21.2 Wataroa Road site (KPA000700)

3.2.21.2.1 Taxa richness and MCI

Twenty-six surveys have been undertaken in the Kapoaiaia Stream at this mid-reach site at Wataroa Road between 1995 and February 2012. These results are summarised in Table 100, together with the results from the current period, and illustrated in Figure 96.

Table 100 Results of previous surveys performed in the Kapoaiaia Stream at Wataroa Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPA000700	26	12-30	21	78-108	93	16	118	25	100

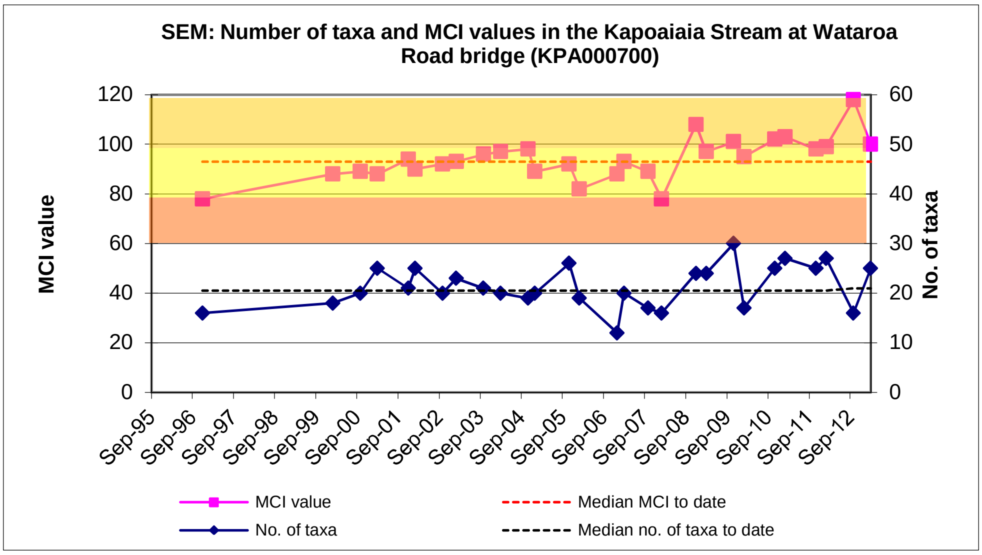


Figure 96 Numbers of taxa and MCI values in the Kapoaiaia Stream at Wataroa Road

A wide range of richnesses (12 to 30 taxa) has been found, with a median richness of 21 taxa, relatively typical of richnesses in the mid-reaches of ringplain streams and rivers. During the 2012-2013 period, spring (16 taxa) and summer (25 taxa) richnesses were very different, well below median taxa number in spring and above the median richness in summer; the latter coincident with much more extensive substrate periphyton cover and warmer water temperatures. MCI values have had a relatively wide range (30 units) at this site, more so than typical of many sites in the mid-reaches of ringplain rivers. The median value (93 units) is lower than values typical of mid-reach sites elsewhere on the ringplain however. The spring, 2012 (118 units) and summer, 2013 (100 units) scores were higher than the historical median by 25 and 7 units respectively with the spring score much higher (by 10 units) than the previous maximum. These scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected'(summer) health for the mid-reaches of a ringplain river. The historical median score (93 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.21.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 101.

Table 101 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kapoaiaia Stream at Wataroa Road between 1995 and February 2012 [26 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
PLATYHELMINTHES	*Cura*	3	1	4		
NEMATODA	Nematoda	3	1	4		
ANNELIDA	Oligochaeta	1	16	62		
	Lumbricidae	5	1	4		
MOLLUSCA	*Potamopyrgus*	4	6	23		
EPHEMEROPTERA	*Austroclima*	7	2	8		A
	Coloburiscus	7	3	12	VA	VA
	Deleatidium	8	12	46	XA	VA
	Nesameletus	9	0	0	A	
PLECOPTERA	*Acroperla*	5	2	8		
COLEOPTERA	Elmidae	6	23	88	VA	VA
MEGALOPTERA	*Archichauliodes*	7	8	31		A
TRICHOPTERA	*Aoteapsyche*	4	18	69		VA
	Costachorema	7	14	54		
	Hydrobiosis	5	16	62		A
	Oxyethira	2	2	8		
	Pycnocentrodes	5	7	27	A	
DIPTERA	*Aphrophila*	5	15	58		
	Maoridiamesa	3	16	62	A	VA
	Orthocladiinae	2	24	92		VA
	Tanytarsini	3	3	12		A
	Empididae	3	3	12		
	Muscidae	3	2	8		A
	Austrosimulium	3	10	38		

Prior to the current 2012-2013 period, 23 taxa had characterised the community at this site on occasions. These have comprised only one 'highly sensitive', ten 'moderately sensitive', and twelve 'tolerant' taxa i.e. a minority of 'highly sensitive' taxa and a downstream increase in the number of 'tolerant' taxa to a higher proportion than might be expected in the mid reaches of a ringplain stream. Predominant taxa have included no 'highly sensitive' taxa; four 'moderately sensitive' taxa [elmid beetles, free-living caddisflies (*Costachorema* and *Hydrobiosis*), and cranefly (*Aphrophila*)]; and four 'tolerant' taxa [oligochaete worms, net-building caddisfly (*Aoteapsyche*), and midges (*Maoridiamesa* and orthoclads)]. Five of the historically characteristic taxa were dominant in the spring, 2012 community together with one 'highly sensitive' mayfly taxon (*Nesameletus*) not previously characteristic at this site. These taxa comprised two 'highly sensitive' taxa, three 'moderately sensitive' taxa, and one 'tolerant' taxon, whereas one 'highly sensitive', five 'moderately sensitive', and five 'tolerant' taxa comprised the dominant taxa of the summer community. However, only four of these thirteen taxa were dominant in both spring and summer communities (Table 101). Increased summer seasonal abundances within five 'tolerant' taxa in particular were reflected in the significant decrease of 2.4 units in $SQMCI_s$ scores between spring and summer (Tables 166 and 167).

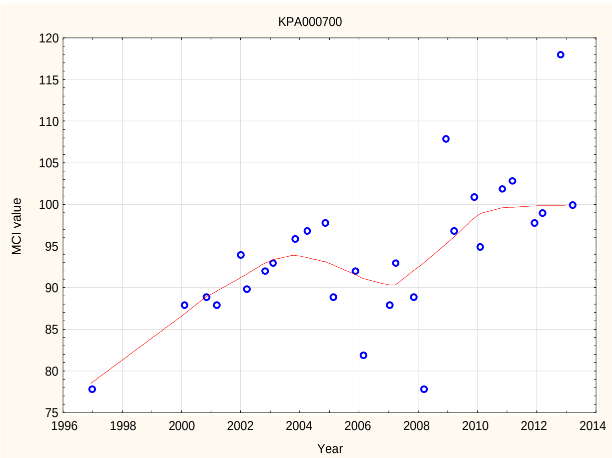
The six characteristic taxa found as very and/or extremely abundant by the seasonal surveys have characterised this site's communities on 12% to 92% of past survey occasions.

3.2.21.2.3 Predicted stream 'health'

The Kapoaiaia Stream site at Wataroa Road, is 13.5 km downstream of the National Park boundary at an altitude of 140 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 99 (altitude) and 103 (distance) for this site. The historical site median (93) is 6 units lower than the altitude prediction and 10 units lower than the distance predictive value. However, the spring, 2012 survey score (118 units) was significantly 19 units higher than the predictive altitude value and 15 units higher than the predictive distance value. The summer, 2013 score (100 units) was one unit above the altitude value and three units below the distance predictive value. Of the 28 surveys to date at this site, 79% of MCI scores have been less than 99 units while only 11% have been greater than 103 units, confirmation of the poorer than predicted historical biological 'health' at this site.

3.2.21.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the SEM results collected to date from the site in the Kapoaiaia Stream at Wataroa Road. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of the LOWESS plot of trends in MCI data and the Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 97.



N = 28
Kendall tau = +0.497
p level = 0.0002 [>FDR, p < 0.001]
Significant at p <0.01 and after FDR application

Figure 97 LOWESS trend plot of MCI data at the Wataroa Road site

Although the initial six years of the monitoring programme indicated a significant temporal improvement in MCI scores, this tended to reverse between 2004 and 2007. Future recent improvement has resulted in an overall sixteen year trend which has been statistically significant (p< 0.01 after FDR). The range of LOWESS-smoothed scores (21 units) has been ecologically significant but has been influenced by an initial very low score. From 2000 to date this range has been 12 units which also has ecological significance.

Smoothed MCI scores have consistently indicated 'fair' generic stream health (Table 1) at this mid-catchment site, improving to border on 'good' over the 2010-2013 period. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a ringplain stream health was in the 'worse than expected' category until 2001 and has been in the 'expected' category since then.

3.2.21.3 Upstream of coast site (KPA000950)

3.2.21.3.1 Taxa richness and MCI

Twenty-six surveys have been undertaken at this lower reach site near the coast in the Kapoaiaia Stream between 1995 and February 2012. These results are summarised in Table 102, together with the results from the current period, and illustrated in Figure 98.

Table 102 Results of previous surveys performed in the Kapoaiaia Stream at the site upstream of the coast together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KPA000950	26	15-24	19	76-98	85	17	101	18	80

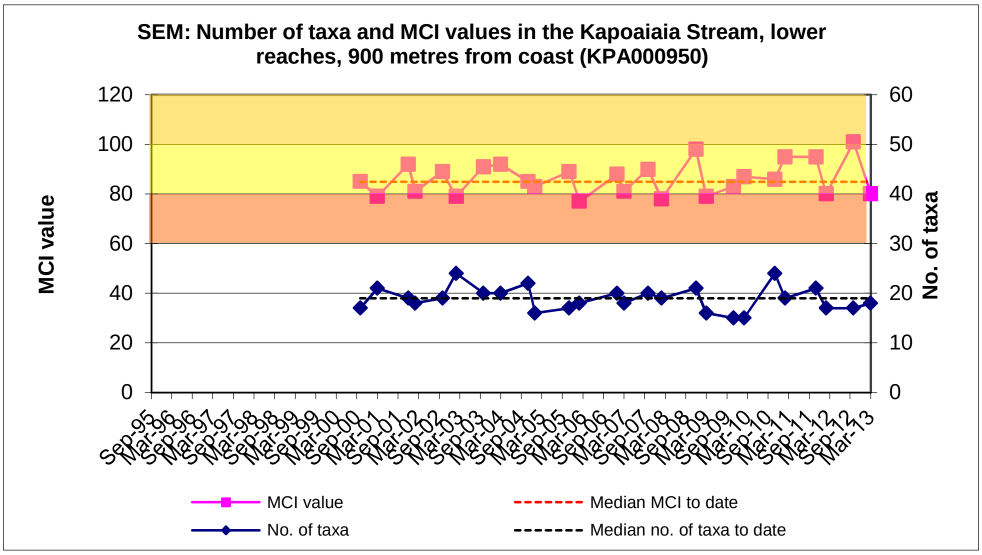


Figure 98 Numbers of taxa and MCI values in the Kapoaiaia Stream at the Cape Egmont (upstream of coast) site

A moderate range of richnesses (15 to 24 taxa) has been found with a median richness of 19 taxa relatively typical of richnesses in the lower reaches of ringplain streams and rivers. During the 2012-2013 period, spring (17 taxa) and summer (18 taxa) richnesses were very similar, but slightly below (by up to two taxa) the median taxa number coincident with patchy (spring) to widespread (summer) substrate periphyton cover and higher water temperatures at the time of the summer survey.

MCI scores have had a moderate range (22 units) at this site, slightly narrower than typical of sites in the lower reaches of ringplain streams. However, the median value (85 units) has been relatively typical of lower reach sites elsewhere on the ringplain

(TRC, 1999b (updated 2012)). The spring, 2012 (101 units) score was higher than typical for such a site and 3 units above the previous maximum for this site, whereas the summer, 2013 (80 units) score was relatively typical for such a site. These scores varied between 13 units above (spring) and 5 units below (summer) the historical median. These scores categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' (spring) and 'worse than expected' (summer) health for the lower reaches of a ringplain stream. The historical median score (85 units) placed this site in the 'fair' category for generic and 'expected' category for predictive methods of assessment.

3.2.21.3.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 103.

Table 103 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kapoaiaia Stream at the site upstream of the coast between 1995 and February 2012 [26 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
PLATYHELMINTHES	*Cura*	3	1	4		
NEMERTEA	Nemertea	3	1	4		
ANNELIDA	Oligochaeta	1	24	92	A	A
	Lumbricidae	5	1	4		
MOLLUSCA	*Potamopyrgus*	4	18	69	A	A
EPHEMEROPTERA	*Austroclima*	7	2	8		
	Deleatidium	8	2	8	A	
COLEOPTERA	Elmidae	6	18	69		
MEGALOPTERA	*Archichauliodes*	7	1	4		
TRICHOPTERA	*Aoteapsyche*	4	21	81	VA	VA
	Costachorema	7	1	4	A	
	Hydrobiosis	5	18	69		A
	Oxyethira	2	5	19		
	Pycnocentrodes	5	13	50	VA	
DIPTERA	*Aphrophila*	5	6	23	A	
	Chironomus	1	1	4		
	Maoridiamesa	3	15	58	VA	
	Orthocladiinae	2	25	96	A	VA
	Tanytarsini	3	5	19		A
	Muscidae	3	3	12		
	Austrosimulium	3	6	23		

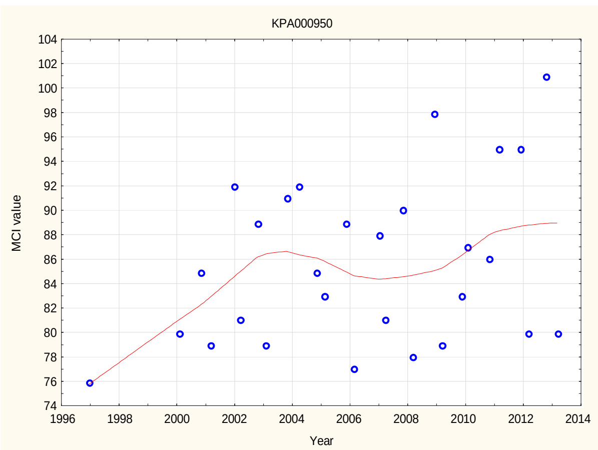
Prior to the current 2012-2013 period 21 taxa have characterised the community at this site on occasions. These have comprised one 'highly sensitive', eight 'moderately sensitive', and twelve 'tolerant' taxa i.e. a high proportion of 'tolerant' taxa as might be expected in the lower reaches of a ringplain stream. Predominant taxa have included no 'highly sensitive' taxa; three 'moderately sensitive' taxa [elmid beetles, free-living caddisfly (*Hydrobiosis*), and stony-cased caddisfly (*Pycnocentrodes*)]; and five 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and midges (orthoclads and *Maoridiamesa*)]. Nine of the historically characteristic taxa were dominant in the spring 2012 community. These comprised one 'highly sensitive', three 'moderately sensitive', and five 'tolerant' taxa. No 'highly sensitive' , one 'moderately sensitive', and five 'tolerant' taxa comprised the dominant taxa in the summer community. Only four of these eleven taxa were dominant in both spring and summer communities (Table 103). An increase in seasonal proportional dominances by 'tolerant' taxa and decrease in 'sensitive' taxa abundances in summer resulted in a decrease of 1.0 unit in seasonal $SQMCI_s$ scores (Table 166 and 167). The four taxa recorded as very abundant during spring and summer had characterised this site's communities on 50% to 96% of past survey occasions.

3.2.21.3.3 Predicted stream 'health'

The Kapoaiaia Stream site near the coast is 25.2 km downstream of the National Park boundary at an altitude of 20 m asl. Relationships for ringplain streams developed between MCI and site altitude and distance from the National Park boundary (Stark and Fowles, 2009), predict MCI values of 86 (altitude) and 96 (distance) for this site. The historical site median (85 units) is within one unit of the altitude prediction but a significant (Stark, 1998) 11 units lower than the distance predictive value. The spring, 2012 survey score (101 units) was a significant 15 units above the altitude predictive value and five units more than the predictive distance value. The summer, 2013 score (80 units) was lower by 6 to a significant 16 units than predictive values. Of the 28 surveys to date at this site, 54% of MCI scores have been less than 86 units while only 7% have been greater than 96 units.

3.2.21.3.4 Temporal trends in 1995 to 2013

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the SEM results collected to date from the site in the Kapoaiaia Stream at near the coast. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 99.



N = 28
Kendall tau = +0.188
p level = 0.160 [>FDR, p = 0.239]
N/S at $p < 0.05$

Figure 99 LOWESS trend plot of MCI data for the site upstream of the coast

No statistically significant temporal trend has been found for the overall monitoring period despite a steady increase in MCI scores over the initial seven year period followed by a small increase recently. There has been a similar, although more pronounced, trend at the mid-catchment site at Wataroa Road. However, there has been an ecologically significant range (of 13 units) in LOWESS-smoothed MCI scores, influenced by the low initial score, but not as wide as the range at the nearest upstream site. Subsequent to the 1997 survey, no usage of the Pungarehu Dairy Factory (between the two sites) has occurred and since 2000 there has been a narrower, ecologically insignificant, range of MCI scores (8 units). In more recent years, there has been an increase in water abstraction in the lower reaches for irrigation purposes.

Smoothed MCI scores have consistently been indicative of 'fair' generic stream health (Table 1) although individual scores have occasionally indicated 'poor' health, invariably under summer, warmer and lower flow conditions. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, health remained in the 'expected' category almost throughout the period, after improving from 'worse than expected' prior to 2000.

3.2.21.4 Discussion

Seasonal MCI values were relatively similar between spring and summer at one site (Wiremu Road) while at the Wataroa Road site and site near the coast more typical summer decreases in MCI scores (18 and 21 units respectively) were recorded. This seasonal variability may be compared with median historical seasonal decreases of 2, 2, and 9 units for these three sites (in a downstream direction, Appendix I). Seasonal communities shared 71% of the 31 taxa found at the upper mid-reach (Wiremu Road) site, 52% of 27 taxa at Wataroa Road, and 52% of 23 taxa at the furthest downstream site in the lower reaches near the coast, indicative of least dissimilarity in seasonal community compositions at the upper mid-reach site.

Community composition varied markedly through the upper mid-reach to lower reach length of the stream surveyed. A total of 30 taxa was recorded in spring of which only 11 taxa were present at all three sites (Table 166). These included two 'highly sensitive', seven 'moderately sensitive', and two 'tolerant' taxa with only one 'highly sensitive' taxon [mayfly (*Deleatidium*)] and one 'moderately sensitive' taxon [stony-cased caddisfly (*Pycnocentrodes*)], abundant at all three sites. A higher total of 37 taxa was found along the river's length by the summer survey (Table 167) of which slightly fewer (10) taxa were present at all three sites. Most of these were also widespread taxa in spring with a loss of one 'highly sensitive' taxon and two 'moderately sensitive' taxa and addition of three 'tolerant' taxa. Only two 'tolerant' taxa were abundant at all three sites. These dissimilarities in spatial community structure along the surveyed length (upper mid-reaches to lower reaches) of the Kapoaiaia Stream were slightly more pronounced in summer.

The MCI scores fell in a downstream direction between the upper mid-reach (Wiremu Road) site and the lower reaches site near the coast by 29 units in spring and typically, to a greater degree, by 46 units in summer, over a river distance of 19.5 km. These seasonal falls in MCI scores equated to rates of decline of 1.5 units/km (spring) and 2.4 units/km (summer), much greater than the predicted rate of 0.8

unit/km for the equivalent length and reach of a National Park-sourced stream (Stark and Fowles, 2009). In terms of seasonal rates, this was more typical of the trend of past summer increases in rates of decline.

Between the upper mid-reach site (Wiremu Road) and Wataroa Road mid-reach site, both the spring (1.5 units/km) and particularly the summer (3.3 units/km) rates of decline were higher than the predicted rate (1.2 units/km) for the equivalent stream reach. For the Wataroa Road mid-reach site to lower reach site near the coast, spring (1.5 units/km) and summer (1.7 units/km) rates of decline were both well above the predicted rate of 0.6 unit/km.

Using the long-term median SEM MCI scores for each site (Appendix II), the rates of decline between upper mid-reach (Wiremu Road) and Wataroa Road mid-reach sites, and the Wataroa Road mid-reach site and lower reach site near the coast have been about 2.1 and 0.8 units per km respectively with an overall average rate of decline of 1.3 MCI units/km over the surveyed length of the stream. Therefore rates of MCI decline over the 2012-2013 period were generally higher than the historical median rates for the 1995 to 2012 period.

3.2.22 Kurapete Stream

Two sites in this small ringplain seepage-sourced stream, one located immediately upstream of the Inglewood Wastewater Treatment (WWTP) and the other nearly 6 km downstream, were included in the SEM programme for the purposes of long term monitoring of the impacts of the removal of the treated wastewater discharge from the stream and also, riparian vegetation planting initiatives in the catchment.

The results of the spring (2012) and summer (2012–2013) surveys are presented in Table 168 and Table 169, Appendix 1.

3.2.22.1 Site upstream of Inglewood WWTP (KRP000300)

3.2.22.1.1 Taxa richness and MCI

Thirty-five surveys have been undertaken, between 1995 and March 2012, at this mid-reach, partly shaded site, draining developed farmland, downstream of Inglewood, but immediately upstream of the WWTP. These results are summarised in Table 104, together with the results from the current period, and illustrated in Figure 100.

Table 104 Results of previous surveys performed in the Kurapete Stream upstream of Inglewood WWTP 2009, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KRP000300	35	13-32	22	80-103	93	25	101	21	101

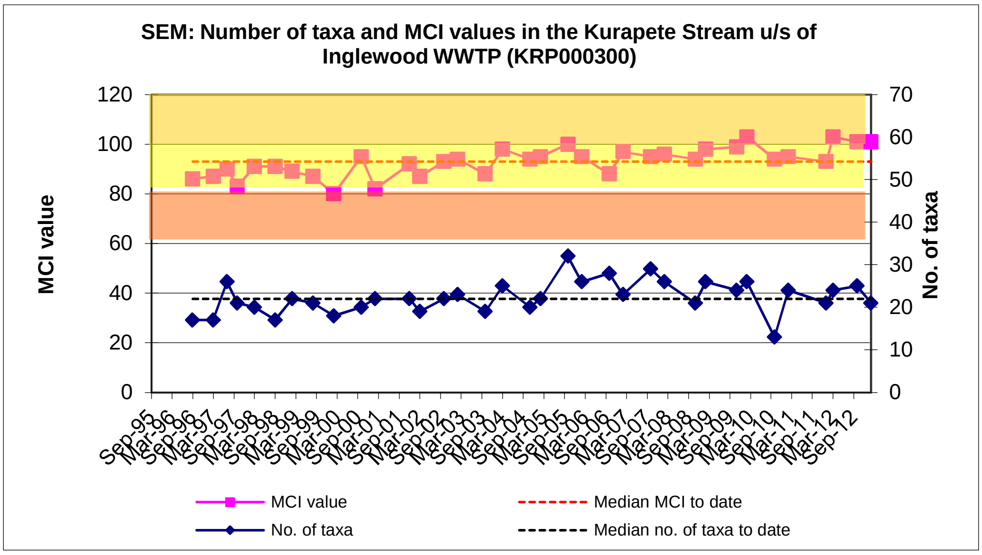


Figure 100 Numbers of taxa and MCI values in the Kurapete Stream upstream of the Inglewood WWTP

A relatively wide range of richnesses (13 to 32 taxa) has been found with a moderate median richness of 22 taxa, relatively typical of richnesses in the mid reaches of ringplain streams rising outside the National Park boundary. During the 2012-2013 period spring (25 taxa) and summer (21 taxa) richnesses were relatively similar with the spring and summer richnesses within three taxa of this median richness coincident with no more than patchy periphyton layers on the predominantly stony-bouldery substrate of this partially shaded site.

MCI values have had a moderate range (23 units) at this site, typical of mid-reach sites in seepage streams on the ringplain. The historical median value (93 units) also has been typical of mid-reach sites rising outside the National Park elsewhere on the ringplain. The spring, 2012 (101 units) and summer, 2013 (101 units) scores were identical and 8 units above the historical median score. Both scores were only two units below the historical maximum score for this site found by the previous (summer 2012) survey. The scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'expected' (spring and summer) health for the mid-reaches of a ringplain stream on these occasions. The historical median score (93 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.22.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 105.

Table 105 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kurapete Stream upstream of Inglewood WWTP, between 1996 and March 2012 [35 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
PLATYHELMINTHES	*Cura*	3	1	3		
NEMERTEA	Nemertea	3	1	3		
ANNELIDA	Oligochaeta	1	26	74	A	A
MOLLUSCA	*Potamopyrgus*	4	21	60	VA	VA
CRUSTACEA	Paraleptamphopidae	5	3	9		
EPHEMEROPTERA	*Austroclima*	7	11	31		A
	Deleatidium	8	2	6	VA	VA
	Zephlebia group	7	14	40	VA	VA
PLECOPTERA	*Acroperla*	5	2	6		
COLEOPTERA	Elmidae	6	17	49	A	VA
MEGALOPTERA	*Archichauliodes*	7	11	31		A
TRICHOPTERA	*Aoteapsyche*	4	25	71		VA
	Hydrobiosis	5	3	9		
DIPTERA	*Aphrophila*	5	21	60	A	
	Maoridiamesa	3	3	9		
	Orthocladiinae	2	24	69	A	
	Tanypodinae	5	0	0		A
	Austrosimulium	3	23	66	A	

Prior to the current 2012-2013 period 17 taxa had characterised the community at this site on occasions. These have comprised only one 'highly sensitive', eight 'moderately sensitive', and eight 'tolerant' taxa i.e. a balance between 'sensitive' and 'tolerant' taxa as would be expected in the mid-reaches of a ringplain stream rising outside the National Park.

Predominant taxa have included only one 'moderately sensitive' taxon [cranefly (*Aphrophila*)] and five 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), orthoclad midges, and sandfly (*Austrosimulium*)].

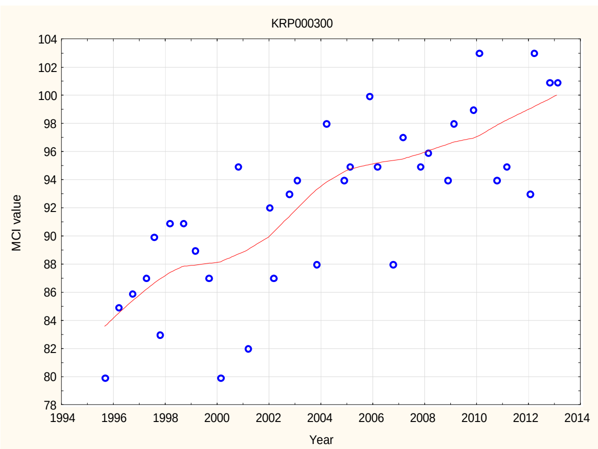
Eight of the historically characteristic taxa were dominant in the spring, 2012 community (Table 105) and comprised one 'highly sensitive', three 'moderately sensitive', and four 'tolerant' taxa. The summer, 2013 community was characterised by five of the spring dominant taxa, plus two 'moderately sensitive' taxa and two fewer 'moderately sensitive' taxa together with one 'moderately sensitive' taxon (tanypod midges) not previously found in abundance at this site. Despite these seasonal differences the relative similarity amongst the most abundant taxa resulted in $SQMCI_s$ scores which varied by only 0.1 unit (Tables 168 and 169). All taxa which were recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 6% to 71% of past survey occasions.

3.2.22.1.3 Predicted stream 'health'

The Kurapete Stream rises below the National Park boundary and the site upstream of the Inglewood WWTP is in the mid-reaches at an altitude of 180 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 103 units for this site. The historical site median (93 units) is 10 units lower than this altitude prediction while the spring survey scores (101 units) were only two units lower than the predictive value. Of the 37 surveys to date at this site, virtually all (95%) of MCI scores have been less than 103 units, indicating that the current spring and summer MCI scores, while typical of historical conditions, were amongst highest scores to date.

3.2.22.1.4 Temporal trends in 1995 to 2012 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site, in the Kurapete Stream upstream of the Inglewood WWTP. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 101.



N = 37
Kendall tau = +0.610
p level < 0.0001 [>FDR, p = < 0.0001]
Significant at p < 0.05 and p < 0.01
and after FDR application

Figure 101 LOWESS trend plot of MCI data at the site upstream of the Inglewood WWTP

The very strong positive temporal trend in MCI scores has been statistically significant at this site immediately upstream of the Inglewood WWTP discharge but below the tributary inflow draining the old Inglewood landfill. This improvement has followed the diversion of the iron-oxide laden drainage out of the stream and into the WWTP system which markedly reduced sediment deposition on the streambed. The strong earlier trend has tended to ease since 2004 with subsequent gradual improvement, while the overall range of LOWESS-smoothed MCI scores (16 units) has been ecologically significant.

LOWESS-smoothed MCI scores have been indicative of 'fair' generic stream health (Table 1) throughout the period bordering on 'good' very recently. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a ringplain stream, health has remained in the 'expected' category over the period subsequent to being in the 'worse than expected' category prior to 2002 (Figure 101).

3.2.22.2 Site approximately 6km downstream of the Inglewood WWTP outfall (KRP000660)

3.2.22.2.1 Taxa richness and MCI

Thirty-five surveys have been undertaken at this lower reach site in the Kurapete Stream 6 km downstream of the Inglewood WWTP outfall (KRP000660) between 1995 and March 2012. These results are summarised in Table 106, together with the results from the current period, and illustrated in Figure 102.

Table 106 Results of previous surveys performed in the Kurapete Stream at the site 6km downstream of the Inglewood WWTP outfall together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to March 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
KRP000660	35	14-30	25	70-103	91	25	112	27	97

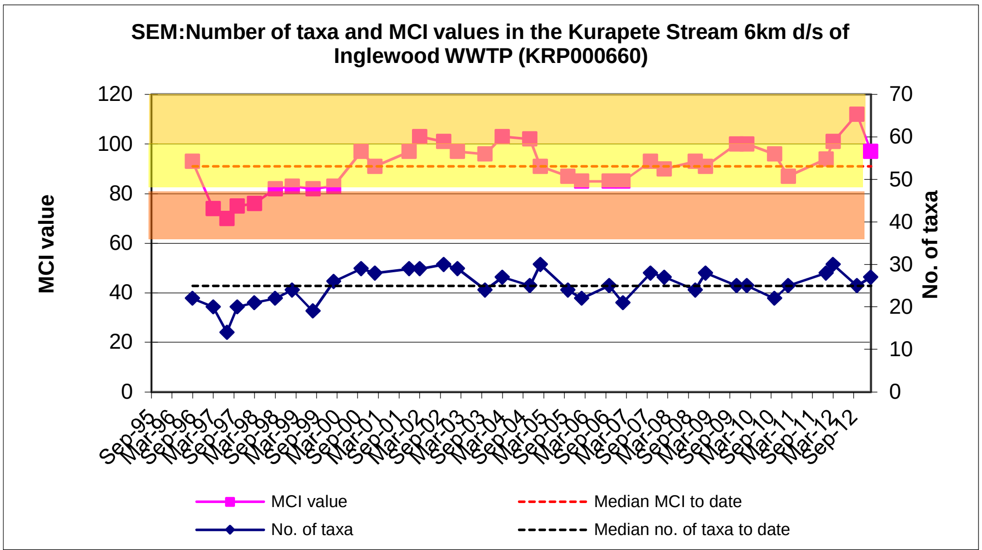


Figure 102 Numbers of taxa and MCI values in the Kurapete Stream, 6 km downstream of the Inglewood WWTP outfall

A moderate range of richnesses (14 to 30 taxa) has been found, with a median richness of 25 taxa (more representative of typical richnesses for the lower mid-reaches of ringplain streams rising outside the National Park boundary (TRC, 1999 (updated, 2013)). During the 2012-2013 period spring (25 taxa) and summer (27 taxa) richnesses were similar and up to two taxa higher than this median richness.

MCI values have had a moderately wide range (33 units) at this site. The median value (91 units) has been typical of lower mid-reach sites in similar streams elsewhere on the ringplain. The spring, 2012 (112 units) and summer, 2012 (97 units) scores were well above (spring) and higher than typical (summer) scores for such a site and a significant (Stark, 1998) 21 units above (spring), and 6 units higher (summer) than the historical median score. The spring score was also 9 units higher than the historical maximum score for this site. These scores categorised this site as having 'good' (spring) and 'fair' (summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer) health for the lower mid-reaches of a seepage-fed ringplain stream coincident with improved physicochemical water quality following the diversion of Inglewood WWTP discharges out of the catchment. The historical median score (91 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.22.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 107.

Table 107 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Kurapete Stream at the site 6 km downstream of Inglewood WWTP outfall, between 1996 and March 2012 [35 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
PLATYHELMINTHES	*Cura*	3	1	3		
NEMERTEA	Nemertea	3	3	9		
NEMATODA	Nematoda	3	1	3		
ANNELIDA	Oligochaeta	1	31	89	A	A
MOLLUSCA	*Potamopyrgus*	4	21	60	A	
CRUSTACEA	Ostracoda	1	1	3		
EPHEMEROPTERA	*Austroclima*	7	6	17	A	
	Coloburiscus	7	5	14	VA	A
	Deleatidium	8	4	11	VA	VA
	Zephlebia group	7	8	23	A	
PLECOPTERA	*Zelandobius*	5	7	20	VA	
COLEOPTERA	Elmidae	6	20	57	A	VA
MEGALOPTERA	*Archichauliodes*	7	10	29	A	A
TRICHOPTERA	*Aoteapsyche*	4	21	60	VA	VA
	Costachorema	7	2	6		
	Hydrobiosis	5	15	43		A
	Oxyethira	2	13	37		
	Pycnocentrodes	5	6	17	A	
DIPTERA	*Aphrophila*	5	25	71	VA	A
	Maoridiamesa	3	8	23		
	Orthocladiinae	2	34	97	A	A
	Tanytarsini	3	4	11		
	Empididae	3	2	6		
	Muscidae	3	3	9		
	Austrosimulium	3	18	51		

Prior to the current 2012-2013 period, 25 taxa had characterised the community at this site on occasions. These have comprised one ‘highly sensitive’, ten ‘moderately sensitive’, and fourteen ‘tolerant’ taxa i.e. a higher proportion of ‘tolerant’ taxa, which is typical of the lower mid-reaches of a ringplain stream.

Predominant taxa have included two ‘moderately sensitive’ taxa [elmid beetles and cranefly (*Aphrophila*)] and five ‘tolerant’ taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), orthoclad midges, and sandfly (*Austrosimulium*)].

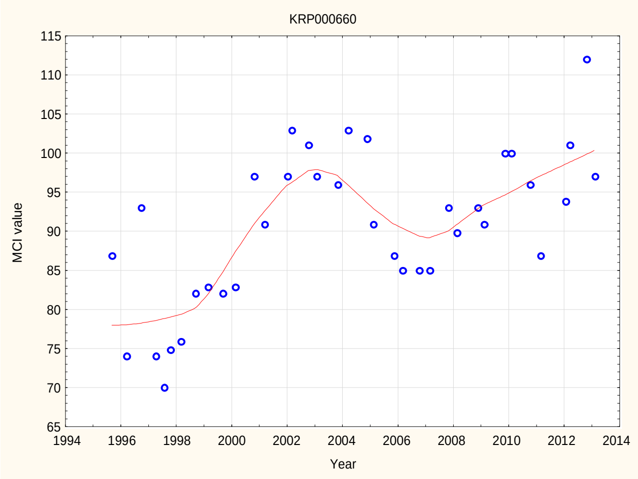
A high number (13) of taxa were dominant in the spring, 2012 community comprising six of the predominant taxa (above) together with one of the other ‘highly sensitive’ and six of the ‘moderately sensitive’ historically characteristic taxa. The summer, 2013 community was characterised by eight of the taxa dominant in spring (including the one ‘highly sensitive’ mayfly taxon), together with one additional ‘moderately sensitive’ taxon all of which previously had been characteristic of this site’s communities (Table 107). Despite some differences in characteristic taxa, there were very similar seasonal $SQMCI_s$ scores which differed by only 0.2 unit (Tables 168 and 169). The six taxa which were recorded as very abundant during spring and/or summer had characterised this site’s communities on 11% to 71% of past surveys.

3.2.22.2.3 Predicted stream ‘health’

The Kurapete Stream rises below the National Park boundary and the site 6 km downstream of the Inglewood WWTP outfall is in the lower mid-reaches at an altitude of 120 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 97 units for this site. The historical site median (91 units) is 6 units lower than altitude prediction and the spring survey score (112 units) was a significant 21 units above the predictive value while the summer score (97 units) was equal with the predictive value. Of the 37 surveys to date at this site, 68% of MCI scores have been less than 97 units, indicating that the current spring and summer MCI scores were less typical of historical conditions although 54% of scores have equalled or exceeded 97 units since wastewater discharges were directed out of the stream in 2000.

3.2.22.2.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the eighteen years of SEM results collected to date from the site in the Kurapete Stream at the site 6 km downstream of the Inglewood WWTP outfall. The MCI has been chosen as the preferable indicator of ‘stream/river health’ for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 103.



N = 37
Kendall tau = +0.409
p level <0.001 [>FDR, p = 0.001]
Significant at p < 0.05 and p < 0.01, and after FDR application

Figure 103 LOWESS trend plot of MCI data for the site 6 km downstream of the Inglewood WWTP outfall

A strong positive temporal trend of MCI score improvement, particularly after 2000 (following diversion of all Inglewood WWTP wastes out of the stream (to the New Plymouth WWTP)), which was emphasised by an ecologically significant increase in LOWESS-smoothed score of 17 units over a 5 year period. Subsequently, a decreasing trend in scores has been followed by a steady recovery (since 2007) while the overall statistical significance of the eighteen-year trend has been significant after FDR application coincident with relatively few consented municipal wastes short-duration discharge overflows to the stream during recent years.

Overall, the trend of LOWESS-smoothed scores indicated improving stream health from 'poor' through 'fair' approaching 'good' in 2003 and again more recently including the 2012-2013 monitoring period. In terms of predictive relationships (Table 2) for a site in the lower mid-reaches of a seepage ringplain stream, health has improved from 'worse than expected' prior to 2000 to the 'expected' category over the remainder of the period indicative of the positive effects of diversion of the WWTP discharge out of the stream.

3.2.22.3 Discussion

Seasonal MCI values atypically remained the same between spring and summer at the site upstream of the Inglewood WWTP outfall but typically decreased at the site 6km downstream where the summer score was 15 units lower than the spring score. These seasonal differences may be compared with historical seasonal medians (Appendix II) which indicate a summer increase of 2 units at the upstream site and a summer decrease of 4 units at the lower site. Seasonal communities shared 53% of total of 30 taxa found at the mid-reach site and 53% of the total of 34 taxa found at the downstream lower mid-reach site indicative of marked seasonal community dissimilarities which were equivalent at the two sites.

MCI score increased atypically by eleven units in spring and decreased by 4 units in summer in a downstream direction between the two sites coincident with the diversion of wastewater discharges from the Inglewood WWTP out of the stream and slightly more frequent but short-duration consented overflow discharges during the 2012-2013 period. These results were more typical of minimal downstream deterioration recorded by most surveys since 2000. These rates of decline in MCI (0 to 0.6 MCI unit/km) were below the rate expected through the mid reaches of Taranaki ringplain rivers and streams sourced outside of the National Park (TRC, 1999).

Using the long-term median SEM MCI scores for each site (Appendix II), the rate of decline between these mid-reach and lower mid-reach sites has been only 0.5 MCI unit/km over the surveyed length. Therefore rates of MCI decline over the 2012-2013 period were lower in spring but very similar in summer to the median historical rate for the 1995 to 2012 period.

Community composition varied markedly through the mid-reach to lower mid-reach length of the stream surveyed. A total of 34 taxa was recorded in spring of which 16 taxa (47%) were present at both sites. These included one 'highly sensitive', nine 'moderately sensitive', and six 'tolerant' taxa with only one 'highly sensitive', three 'moderately sensitive', and three 'tolerant' taxa abundant at both sites. A lower total of 32 taxa was found along the stream's surveyed length by the summer survey of which sixteen taxa (50%) were present at both sites. They were very similar to the widespread taxa in spring with two fewer 'moderately sensitive' taxa, and two additional 'tolerant' taxa. Five taxa were abundant at both sites in summer; two 'tolerant', two 'moderately sensitive', and one 'highly sensitive' taxa. Dissimilarities in spatial community structure along the surveyed length of the Kurapete Stream were slightly less pronounced in spring than in summer, more similar to most seasonal structures to date.

3.2.23 Waiokura Stream

Two sites in this small, intensively dairy-farmed, ringplain seepage-sourced stream, were included in the SEM programme in recognition of a long-term collaborative study of the effects of best-practice dairy-farming initiatives being evaluated in five dairying catchments throughout the country (Wilcock et al, 2009). Fonterra, Kapuni lactose factory also irrigates wastewater to land in the mid reaches of this catchment. The two sites are located upstream of the irrigation area (in mid-catchment) and approximately 10 km further downstream toward the lower reaches of the stream. Some consent monitoring data have been collected from the upper site since 2003 whereas the downstream site was established for biological trend monitoring purposes in the 2008-2009 period.

The results of spring (2012) and summer (2012-2013) surveys are summarised in Tables 170 and 171, Appendix I.

3.2.23.1 Skeet Road site (WKR000500)

3.2.23.1.1 Taxa richness and MCI

Thirteen surveys have been undertaken, between 1996 and February 2012, at this mid-reach, partially shaded site, draining open developed farmland upstream of the Fonterra, Kapuni wastewater irrigation area. These results are summarised in Table 108, together with the results form the current period, and illustrated in Figure 104.

Table 108 Results of previous surveys performed in the Waiokura Stream at Skeet Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1996 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2012	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WKR000500	15	20-28	23	88-107	98	18	114	29	101

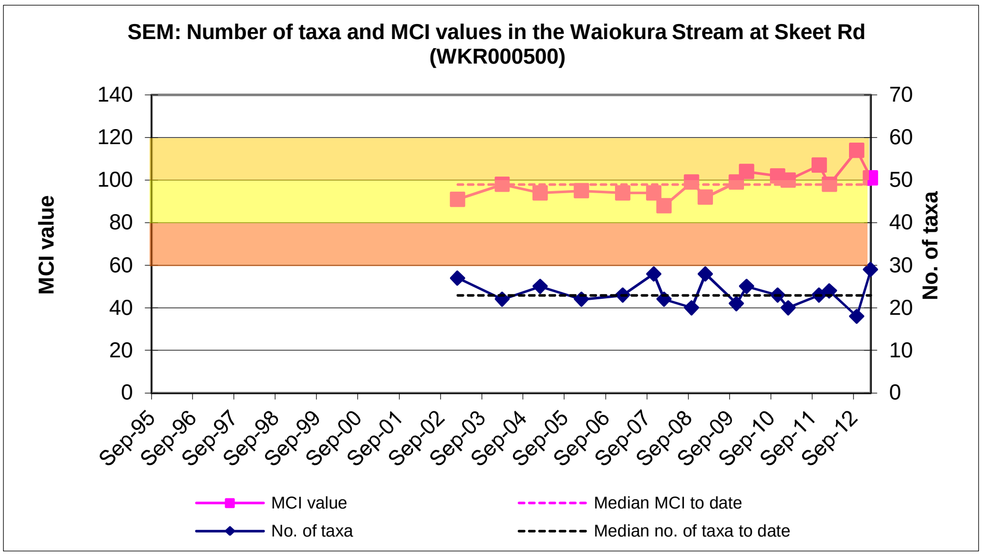


Figure 104 Numbers of taxa and MCI values in the Waiokura Stream at Skeet Road

A relatively narrow range of richnesses (20 to 28 taxa) has been found to date with a median richness of 23 taxa more typical of richnesses in the mid reaches of ringplain streams rising outside the National park boundary. During the 2012-2013 period spring (18 taxa) and summer (29 taxa) richnesses were dissimilar varying from 5 taxa fewer to 6 taxa more than this median richness coincident with minimal periphyton on the predominantly gravel-cobble substrate of this site in spring and in summer following relatively short flow recession periods on both occasions.

MCI values have had a relatively narrow range (19 units) at this site, atypical of mid reach sites on the ringplain, but the monitoring period has been relatively short to date. The historical median value (98 units) has been typical of mid-reach sites in streams rising outside the National Park elsewhere on the ringplain (TRC, 1999b (updated, 2012)). The spring, 2012 (114 units) and summer, 2012 (101 units) scores were a significant 16 units and 3 units above the historical median respectively. The spring score was also 7 units higher than the previous maximum for this site. The scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' and 'expected' health respectively for the mid-reaches of a ringplain stream on these occasions. The historical median score (98 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.23.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2011-2013 period are listed in Table 109.

Table 109 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiokura Stream at Skeet Road, between 2002 and February 2012 [15 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	7		
ANNELIDA	Oligochaeta	1	9	60		
MOLLUSCA	*Potamopyrgus*	4	5	33		VA
CRUSTACEA	*Paracalliope*	5	1	7		
	Paraleptamphopidae	5	1	7		
EPHEMEROPTERA	*Austroclima*	7	15	100	VA	XA
	Coloburiscus	7	2	13	A	A
	Deleatidium	8	8	53	A	VA
	Zephlebia group	7	4	27		VA
PLECOPTERA	*Zelandobius*	5	1	7		
COLEOPTERA	Elmidae	6	15	100	VA	XA
MEGALOPTERA	*Archichauliodes*	7	8	53	A	A
TRICHOPTERA	*Aoteapsyche*	4	15	100	A	VA
	Costachorema	7	1	7		
	Hydrobiosis	5	4	27		
	Pycnocentrodes	5	7	47	VA	
DIPTERA	*Aphrophila*	5	1	7		
	Maoridiamesa	3	3	20		
	Orthocladiinae	2	7	47		
	Tanytarsini	3	1	7		
	Austrosimulium	3	0	0		A

Prior to the current 2012-2013 period 20 taxa had characterised the community at this site on occasions. These have comprised only one 'highly sensitive', 12 'moderately sensitive' and seven 'tolerant' taxa i.e. a moderately high proportion of 'sensitive' taxa as would be expected in the mid-reaches of a ringplain stream rising outside the National Park.

Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; three 'moderately sensitive' taxa [mayfly (*Austroclima*), elmid beetles, and dobsonfly (*Archichauliodes*)]; and two 'tolerant' taxa [oligochaete worms and net-building caddisfly (*Aoteapsyche*].

Seven of the historically characteristic taxa were dominant in the spring, 2012 community comprising five of the predominant taxa (above) together with two other 'moderately sensitive' taxa. The summer, 2013 community was characterised by two additional taxa; six of the taxa dominant in spring, with one additional 'moderately sensitive' taxon and one 'tolerant' taxon and one other 'tolerant' taxon [sandfly (*Austrosimulium*)] not previously characteristic of this site"s community. Increased summer abundances both in 'sensitive' and 'tolerant 'dominant taxa resulted in minimal change in the seasonal $SQMCI_s$ scores (0.2 unit) (Tables 170 and 171). The seven taxa which were recorded as very/extremely abundant during spring and/or summer had characterised this site's communities on 27% to 100% of past survey occasions (Table 109).

3.2.23.1.3 Predicted stream 'health'

The Waiokura Stream rises below the National Park boundary and the site at Skeet Road is in the mid-reaches at an altitude of 150 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 100 units for this site. The historical site median (98 units) is 2 units below this altitude prediction while the spring survey score (114 units) and the summer score (101 units) were a significant 14 units (spring) and within 1 unit (summer) of the predictive value. Of the 17 surveys to date at this site, 59% of MCI scores have been less than 100 units, indicating that the current spring and summer MCI scores were atypical of historical conditions (but the shorter monitoring period to date should be noted for this site).

3.2.23.1.4 Temporal trends in 2002 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has not been performed as the duration and frequency of data collection has been insufficient to date from this site in the Waiokura Stream at Skeet Road. The MCI has been chosen as the preferable indicator' of stream/river health' for SEM trend reporting purposes. However, a graphical presentation of LOWESS plot of trends in MCI data is provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 105.

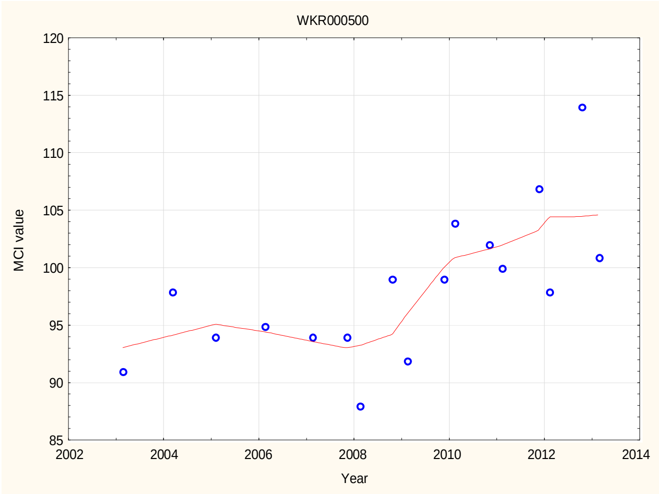


Figure 105 LOWESS trend plot of MCI data at the Skeet Road site

More recently there has been relatively strong temporal improvement in MCI scores at this site. The LOWESS-smoothed range of MCI scores (11 units) has bordered on ecological significance and increases in scores may have been related to improvements in farming practices and/or wastes disposal in the rural catchment between the stream's seepage sources (below the National Park) and mid reaches at Skeet Road, although the shorter duration and less frequent initial monitoring must be noted.

LOWESS-smoothed MCI scores have been indicative of 'fair' generic stream health (Table 1) almost throughout the period but entering the 'good' health category more recently. In terms of predictive relationships (Table 2) for a site in the mid-reaches of a seepage-fed ringplain stream, health has remained in the 'expected' category over the entire ten year period (Figure 105).

3.2.23.2 Manaia golf course site (WKR000700)

3.2.23.2.1 Taxa richness and MCI

Ten surveys have been undertaken at this more recently established lower reach site in the Waiokura Stream at Manaia between 2007 and February 2012. These results are summarised in Table 110 together with the results from the current period, and illustrated in Figure 106.

Table 110 Results of previous surveys performed at Waiokura Stream at Manaia golf course, , together with spring 2012 and summer 2013 results

Site code	SEM data (2007 to Feb 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
WKR000700	10	16-26	23	92-103	99	24	105	27	100

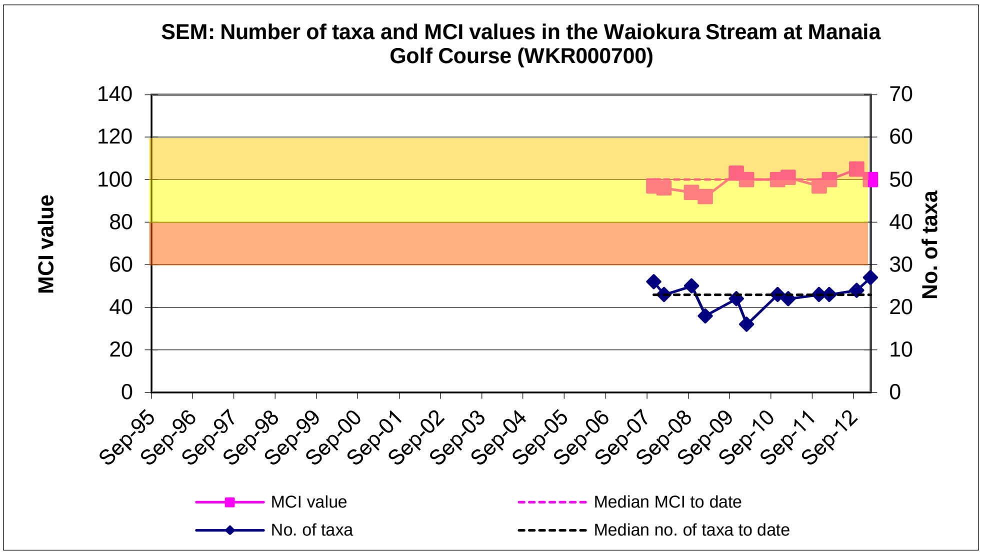


Figure 106 Numbers of taxa and MCI values in the Waiokura Stream at Manaia Golf course

A moderate range of richnesses (16 to 26 taxa) has been found, with a median richness of 23 taxa (more representative of typical richnesses for the lower reaches of ringplain streams rising outside the National Park boundary). During the 2012-2013 period spring (24 taxa) and summer (27 taxa) richnesses were relatively similar and up to 4 taxa higher than this median richness.

MCI values have had a narrow range (11 units) at this site partly as a result of the short duration of the monitoring period to date. The median value (99 units) has been slightly higher than typical of similar lower reach sites elsewhere on the ringplain (TRC, 1999b (updated 2012)). The spring, 2012 (105 units) and summer, 2013 (100 units) scores were within 6 units of the historical median with the spring score 2 units higher than previously recorded over the relatively short monitoring period to date. These scores categorised this site as having 'good' (spring and summer) health generically (Table 1) and, in terms of predictive relationships (Table 2), 'better than expected' (spring) and 'expected' (summer), health for the lower reaches of a ringplain stream coincident with some riparian cover within the golf course reaches. The historical median score (99 units) placed this site in the 'fair' and 'expected' categories for generic and predictive methods of assessment respectively.

3.2.23.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site for the short period prior to the 2012-2013 period are listed in Table 111.

Table 111 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Waiokura Stream at the Manaia golf course, between 2007 and February 2012 [10 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMATODA	Nematoda	3	1	10		
ANNELIDA	Oligochaeta	1	9	90	A	A
MOLLUSCA	*Potamopyrgus*	4	6	60		
CRUSTACEA	*Paracalliope*	5	1	10		
EPHEMEROPTERA	*Austroclima*	7	10	100	VA	VA
	Coloburiscus	7	6	60		VA
	Deleatidium	8	1	10		
	Zephlebia group	7	10	100	A	XA
PLECOPTERA	*Zelandobius*	5	2	20		
COLEOPTERA	Elmidae	6	10	100	VA	VA
MEGALOPTERA	*Archichauliodes*	7	8	80	A	A
TRICHOPTERA	*Aoteapsyche*	4	8	80		VA
	Hydrobiosis	5	1	10		
	Pycnocentria	7	0	0		A
	Pycnocentrodes	5	2	20	VA	

Prior to the current 2012-2013 period, 14 taxa had characterised the community at this site on occasions. These have comprised one 'highly sensitive', nine 'moderately sensitive', and four 'tolerant' taxa i.e. a higher proportion of 'sensitive' taxa than might be expected in the lower reaches of a ringplain stream, coincident with the riparian cover provided within the Manaia golf course.

Predominant taxa have included five 'moderately sensitive' taxa [mayflies (*Austroclima*, *Zephlebia* group, and *Coloburiscus*), elmid beetles and dobsonfly (*Archichauliodes*)] and three 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), and net-building caddisfly (*Aoteapsyche*)].

Six of these historically characteristic taxa were dominant in the spring, 2012 community comprising five of the predominant taxa (above). The summer, 2013 community was characterised by all but one of the taxa dominant in spring, with one additional 'moderately sensitive' and one 'tolerant' taxa (Table 111) plus one 'moderately sensitive' taxon [caddisfly (*Pycnocentrodes*)] previously not found in abundance at this site. Increased summer abundances within two 'sensitive' taxa in particular resulted in a small increase of 0.6 unit in seasonal $SQMCI_s$ scores (Tables 170 and 171). All taxa which were recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 20% to 100% of past surveys.

3.2.23.2.3 Predicted stream 'health'

The Waiokura Stream rises below the National Park boundary and the site at the Manaia golf course is in the lower reaches at an altitude of 70 m asl. A relationship for ringplain streams developed between MCI and site altitude (Stark and Fowles, 2009), predicts a MCI value of 92 units for this site. The short-term historical site median (99 units) is 7 units above this altitude prediction coincident with patchy riparian vegetation cover in the reaches through the Manaia golf course. Both the spring survey score (105 units) and the summer score (100 units) were higher than this predictive value by 8 to a significant 13 units. Of the twelve surveys to date at this site, no MCI scores have been less than the predicted 92 units. The summer MCI score was typical of historical conditions, and the spring score was better than all past scores although the monitoring period has been relatively short (six years).

3.2.23.2.4 Temporal trends in 2007 to 2013 data

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has not been performed on the six years of SEM results collected to date from the site in the Waiokura Stream at Manaia golf course due to the short duration of the programme to date. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. However a graphical presentation of LOWESS plot of trends in MCI data is provided for this site despite the short period. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 107.

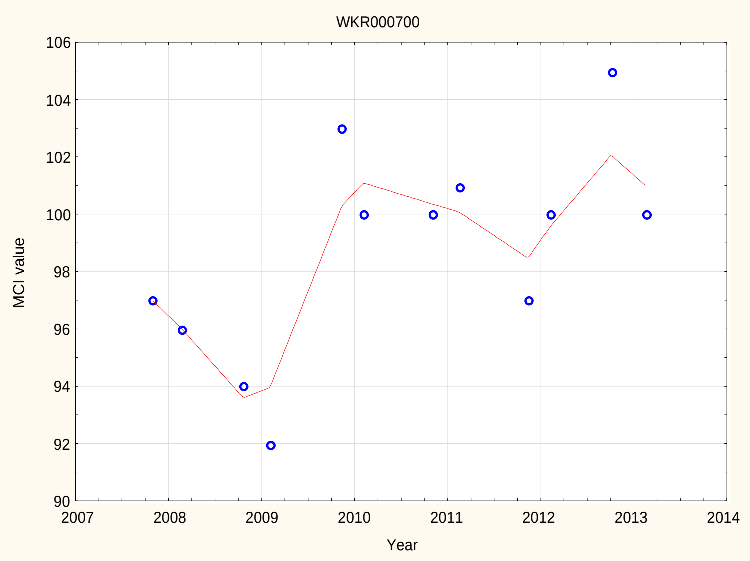


Figure 107 LOWESS trend plot of MCI data for the Manaia golf course

A similar temporal trend of a marked improvement in MCI scores to that found at the upstream site (at Skeet Road) was identified at this site at the Manaia golf course but the short duration of the data record must be noted at this stage. The LOWESS-smoothed range of scores (8 units) has no ecological significance.

The smoothed MCI scores which indicated 'fair' generic stream health (Table 1) for two years of the monitoring period, improved to 'good' stream health for two years before returning to 'fair' stream health and then 'good' health most recently. In terms of predictive relationships (Table 2) for a site in the lower reaches of a ringplain stream, health has remained in the 'expected' category over the period and approaching the 'better than expected' category very recently, further indication of the value of the riparian cover present through the golf course reach of this stream.

3.2.23.3 Discussion

Seasonal MCI values typically decreased (by 13 units) at the mid-reach site and decreased (by 5 units) between spring and summer at the site in the lower reaches. These seasonal differences may be compared with the historical median seasonal summer decrease of 7 units at the Skeet Road site and summer increase of one unit at the Manaia Golf Course site (Appendix II). Seasonal communities shared only 42% of the 33 taxa found at the mid-reach site and 58% of 31 taxa at the downstream site in the lower reaches at Manaia indicative of increased similarity in seasonal community composition in a downstream direction within the riparian covered reaches.

MCI score decreased by 9 units in spring but atypically (for a ringplain stream) improved, but only by one unit in summer in a downstream direction, between the more open farmland mid-reach site (Skeet Road) and the lower reach Manaia golf course site, coincident with some improvement in habitat provided by patches of riparian vegetation cover through the golf course. These differences in MCI scores between sites represented a rate of MCI decrease of 0.9 unit/km (spring) and 0.1 unit/km (summer); below the rate expected through the mid to lower reaches of Taranaki ringplain rivers and streams sourced outside of the National Park (TRC, 1999).

Using the longer-term median SEM MCI scores for each site (for the short period 2007 to date), there has been no decline (but an improvement of 2 units) between the mid-reach site at Skeet Road and the lower reach site near Manaia over the surveyed length. Therefore rate of MCI decline (and improvement) in the 2012-2013 period for spring was atypically higher but for summer the rate was relatively typical in terms of the median historical rate.

Community composition varied through the mid reach to lower reach length of the stream surveyed. A total of 26 taxa was recorded in spring of which 16 taxa (62%) were present at both sites. These included one ‘highly sensitive’, twelve ‘moderately sensitive’, and three ‘tolerant’ taxa with only four ‘moderately sensitive’ taxa abundant at both sites. A higher total (37 taxa) was found along the stream’s surveyed length by the summer survey of which 19 taxa (51%) were present at both sites. They were generally similar to the widespread taxa in spring with a small decrease in ‘moderately sensitive’ taxa and increase of three ‘tolerant’ taxa. Only six taxa were abundant at both sites in summer; one ‘tolerant’ and five ‘moderately sensitive’ taxa. Dissimilarities in spatial community structure along the surveyed length (mid to lower reaches) of the Waiokura Stream were more pronounced in summer.

3.2.24 Tangahoe River

Three sites in this eastern hill country river were included in the SEM programme in 2007 for the purpose of monitoring long-term land use changes (afforestation) particularly in the upper-mid catchment. The Fonterra, Hawera dairy factory abstracts water from the river in the lower catchment for processing purposes. Two of the three sites are in the upper to mid, shallow gradient, reaches of the river (the upstream site within 4 km of the headwaters) with the third site in the lower reaches.

The results of the spring, 2012 survey are presented in Table 172 and the summer, 2012–2013 survey in Table 173, Appendix I.

3.2.24.1 Upper Tangahoe Valley Road site (TNH000090)

3.2.24.1.1 Taxa richness and MCI

Ten surveys have been undertaken at this upper reach site in the Tangahoe River between 2007 and March 2012. These results are summarised in Table 112, together with the results from the current period, and illustrated in Figure 108.

Table 112 Results of previous surveys performed in the Tangahoe River at upper Tangahoe Valley Road, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)						2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
TNH000090	10	17-30	23	90-106	97	20	93	27	107

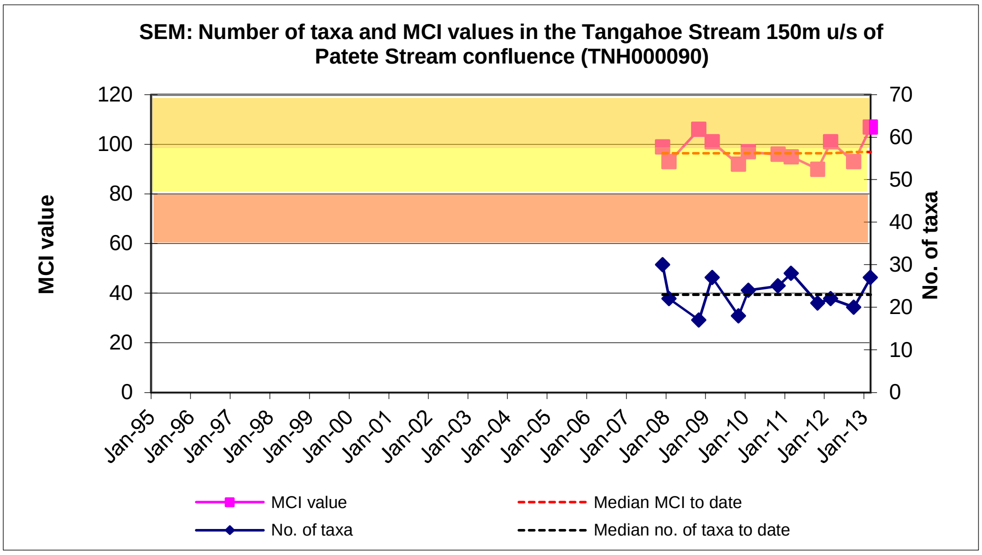


Figure 108 Numbers of taxa and MCI values in the Tangahoe River at Upper Tangahoe Valley Road

A relatively wide range of richnesses (17 to 30 taxa) has been found with a moderate median richness of 23 taxa (lower than richnesses which might be anticipated in the upper reaches of eastern hill country rivers) but higher than the median richness (19 taxa) for sites at this relatively low altitude (85 m asl) (TRC, 1999b (updated, 2012)). During the 2012-2013 period, spring (20 taxa) and summer (27 taxa) richnesses were three taxa less (spring) to four taxa more (summer) than this median richness.

MCI values have had a relatively narrow range (16 units) at this site, more typical of a site in the upper reaches of streams and rivers. However, the median value (97 units) has been more typical of mid reach sites elsewhere and 5 units above the median score recorded by 47 previous surveys at 'control' sites located at similar altitudes (to the upper Tangahoe Valley Road site) in eastern hill country rivers and streams (TRC, 1999b (updated 2012)). The spring, 2012 (93 units) and summer, 2013 (107 units) scores were 4 units lower to 10 units higher than the historical median. This summer score was one unit higher than previously found at this site over the relatively brief monitoring period to date. These scores categorised this site as having 'fair' health generically (Table 1) in spring and 'good' health in summer. The historical median score (97 units) place this site in the 'fair' category for the generic method of assessment.

3.2.24.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 113.

Table 113 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Tangahoe River at upper Tangahoe Valley Road between 2007 and March 2012 [10 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	7	70		A
MOLLUSCA	*Potamopyrgus*	4	9	90		VA
EPHEMEROPTERA	*Austroclima*	7	10	100	A	A
	Deleatidium	8	8	80	XA	XA
	Zephlebia group	7	5	50		
PLECOPTERA	*Megaleptoperla*	9	1	10		A
COLEOPTERA	Elmidae	6	9	90	A	VA
MEGALOPTERA	*Archichauliodes*	7	1	10		A
TRICHOPTERA	*Aoteapsyche*	4	2	20		
	Hydrobiosis	5	3	30		
DIPTERA	Orthocladiinae	2	2	20	A	
	Austrosimulium	3	8	80	VA	

Prior to the current 2012-2013 period, 12 taxa have characterised the community at this site on occasions. These have comprised two 'highly sensitive', five 'moderately sensitive', and five 'tolerant' taxa i.e. a higher proportion of 'tolerant' taxa than would be expected in the upper reaches of an eastern hill-country river, reflecting the relatively flat gradient of this river. Predominant taxa have included one 'highly sensitive' taxon (mayfly (*Deleatidium*)); three 'moderately sensitive' taxa [mayflies (*Austroclima* and *Zephlebia* group) and elmid beetles]; and three 'tolerant' taxa [snail (*Potamopyrgus*), oligochaete worms, and sandfly (*Austrosimulium*)]. Four of these predominant taxa were dominant in the spring, 2012 community together with one other historically characteristic taxon. The summer, 2013 community was characterised by three of the taxa dominant in spring, together with four additional taxa which previously had been characteristic of this site's communities (Table 113). Despite several seasonal differences in characteristic taxa, the numerical dominance taxa in spring and summer surveys by the one 'highly sensitive' taxon was reflected in the very similar seasonal $SQMCI_s$ scores which were within 0.1 unit (Table 172 and 173). The four taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 80% to 90% of past survey occasions.

3.2.24.1.3 Predicted stream 'health'

The Tangahoe River site at upper Tangahoe Valley Road, at an altitude of 85 m asl, is in the upper reaches of this low gradient river draining an eastern hill country catchment. A relationship for ringplain streams and river developed between MCI and altitude (Stark and Fowles, 2009) is therefore not appropriate for this river.

3.2.24.1.4 Temporal trends in 2007 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has not been performed on the six years of SEM results collected to date from the site in the Tangahoe River at upper Tangahoe Valley Road due to the short duration of the data record. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. However, a graphical presentation of LOWESS plot trends in MCI data is provided for this site. The LOWESS (tension 0.14) trend plot of MCI data is presented in Figure 109.

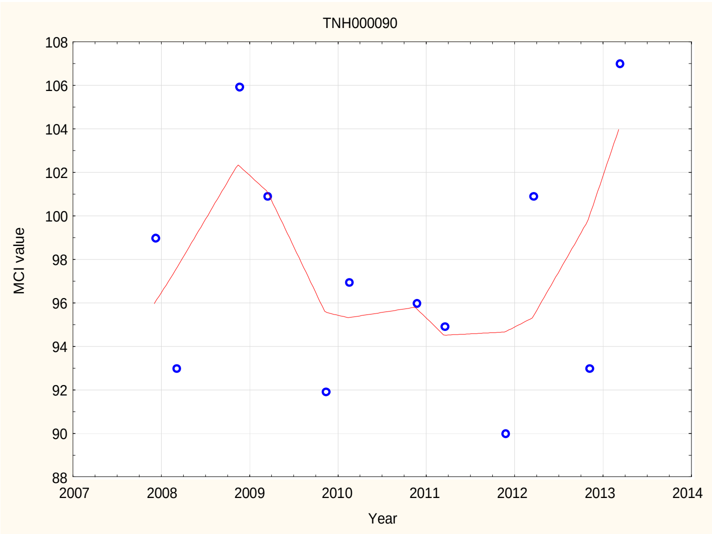


Figure 109 LOWESS trend plot of MCI data for the upper Tangahoe Valley site

No temporal trend in MCI scores may be interpreted for this upper river reach, eastern hill country site due to the short monitoring period to date. The range of smoothed MCI scores (9 units) has no ecological significance but cannot be fully assessed until the monitoring period is of sufficient duration.

Smoothed MCI scores ranging from 'fair' over the majority of the period, to 'good' generic river health (Table 1) have been recorded over the six year period (Figure 109).

3.2.24.2 Tangahoe Valley Road bridge site (TNH000200)

3.2.24.2.1 Taxa richness and MCI

Ten surveys have been undertaken at this mid reach site in the Tangahoe River between 2007 and March 2012. These results are summarised in Table 114, together with the results from the current period, and illustrated in Figure 110.

Table 114 Results of previous surveys performed in the Tangahoe River at Tangahoe Valley Road bridge, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
TNH000200	10	20-33	25	92-108	105	23	104	27	97

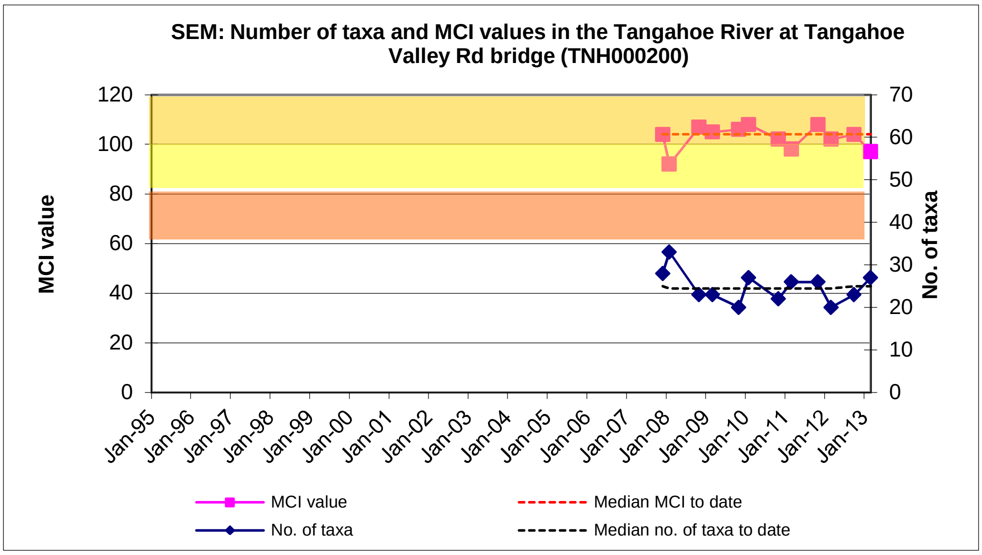


Figure 110 Numbers of taxa and MCI values in the Tangahoe River at Tangahoe Valley Road bridge

A moderate range of richnesses (20 to 33 taxa) has been found with a relatively good median richness of 25 taxa (typical of richnesses in the mid-reaches of eastern hill country rivers). During the 2012-2013 period, spring richness (23 taxa) was slightly below the median, while summer richness (27 taxa) was above this median taxa number.

MCI values have had a moderate range (16 units) at this site, typical of a site in the mid-reaches of eastern hill country streams and rivers. The median value (105 units) has also been typical of mid-reach sites elsewhere and three units above the median score recorded by 16 previous surveys at 'control' sites located at similar altitudes in eastern hill country rivers and streams (TRC, 1999b (updated, 2012)). The spring, 2012 (104 units) and summer, 2013 (97 units) scores were an insignificant 8 units lower than the historical median. These scores categorised this site as having 'good' health generically (Table 1) in spring and 'fair' health in summer. The historical median score (105 units) placed this site in the 'good' category for the generic assessment of health.

3.2.24.2.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 115.

Table 115 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Tangahoe River at Tangahoe Valley Road bridge between 2007 and March 2012 [10 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	2	20		
MOLLUSCA	*Potamopyrgus*	4	6	60		VA
EPHEMEROPTERA	*Austroclima*	7	10	100	A	VA
	Coloburiscus	7	3	30		A
	Deleatidium	8	8	80	VA	VA
	Rallidens	9	1	10		
	Zephlebia group	7	7	70		
PLECOPTERA	*Acroperla*	5	2	20		
	Zelandobius	5	4	40	A	
COLEOPTERA	Elmidae	6	10	100	A	VA
MEGALOPTERA	*Archichauliodes*	7	3	30		
TRICHOPTERA	*Aoteapsyche*	4	9	90		VA
	Hydrobiosis	5	5	50		
	Oxyethira	2	2	20		
	Pycnocentrodes	5	1	10		
DIPTERA	*Aphrophila*	5	5	50	A	VA
	Orthocladiinae	2	5	50	A	
	Tanytarsini	3	4	40		A
	Austrosimulium	3	3	30		

Prior to the current 2012-2013 period, 19 taxa have characterised the community at this site on occasions. These have comprised two 'highly sensitive', ten 'moderately sensitive', and seven 'tolerant' taxa i.e. a relatively high proportion of 'sensitive' taxa as would be expected in the mid-reaches of an eastern hill-country river. Predominant taxa have included one 'highly sensitive' taxon [mayfly (*Deleatidium*)]; five 'moderately sensitive' taxa [mayflies (*Austroclima* and *Zephlebia* group), elmid beetles, caddisfly (*Hydrobiosis*), and cranefly (*Aphrophila*)]; and three 'tolerant' taxa [snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and orthoclad midges]. Five of these predominant taxa were dominant in the spring, 2012 community together with one other taxon which had been characteristic previously. The summer, 2013 community was characterised by four of the taxa dominant in spring, together with an additional four taxa; all of which previously had been characteristic of this site's communities (Table 115). Increases in summer abundances of several 'tolerant' taxa principally were responsible for the decrease of 0.6 unit in seasonal $SQMCI_s$ scores (Tables 172 and 173). The six taxa recorded as very abundant during spring and/or summer had characterised this site's communities on 50% to 100% of the past surveys.

3.2.24.2.3 Predicted stream 'health'

The Tangahoe River site at Tangahoe Valley Road bridge, at an altitude of 65 m asl, is in the mid reaches of a river draining an eastern hill country catchment. A relationship for ringplain streams and river developed between MCI and altitude (Stark and Fowles, 2009) is therefore not appropriate for this river.

3.2.24.2.4 Temporal trends in 2007 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has not been performed on the six years (2007-2013) of SEM results collected to date from the site in the Tangahoe River at the Tangahoe Valley Road bridge site due to the short period of data record. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. However, a graphical presentation of LOWESS plot trends in MCI data is provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 111.

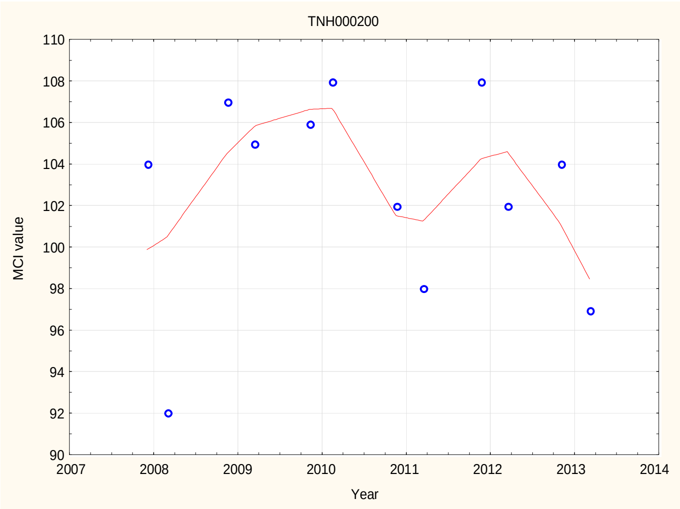


Figure 111 LOWESS trend plot of MCI data for the Tangahoe Valley Road bridge site

No temporal trend in MCI scores may yet be interpreted for this mid river reach, eastern hill country site. The range of smoothed MCI scores (8 units) over the period has no ecological significance, but cannot be accurately assessed until the monitoring period is of sufficient duration.

Smoothed MCI scores have indicated 'good' generic river health (Table 1) over almost the entire six year period, falling recently into the 'fair' category.

3.2.24.3 Site downstream of railbridge (TNH000515)

3.2.24.3.1 Taxa richness and MCI

Nine surveys have been undertaken at this lower reach site in the Tangahoe River between 1995 and March 2012 with eight of these surveys since 2007. These results are summarised in Table 116, together with the results from the current period, and illustrated in Figure 112.

Table 116 Results of previous surveys performed in the Tangahoe River d/s of railbridge, together with spring 2012 and summer 2013 results

Site code	SEM data (1995 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Mar 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
TNH000515	11	13-26	20	84-104	92	19	100	15	87

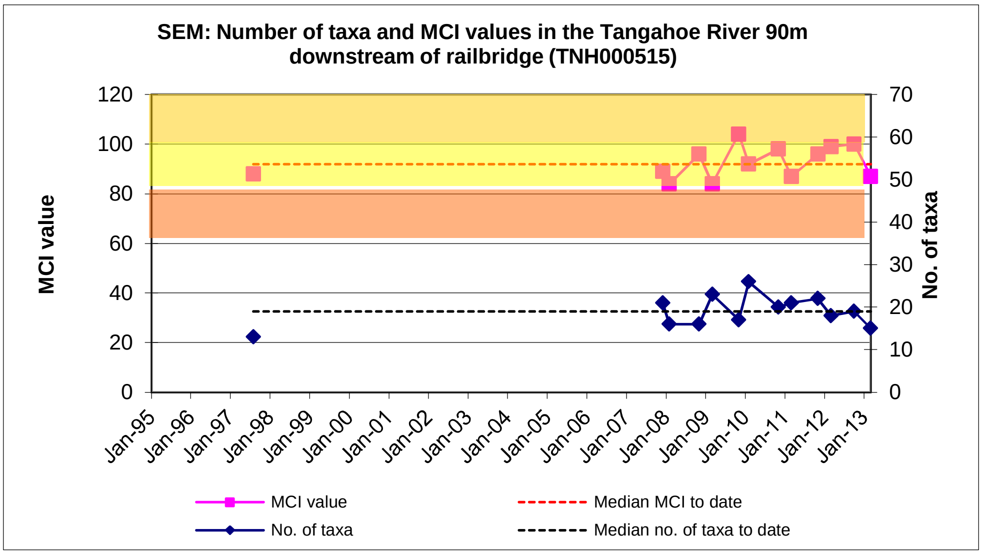


Figure 112 Numbers of taxa and MCI values in the Tangahoe River downstream of the railbridge

A moderate range of richnesses (13 to 26 taxa) has been found with a slightly higher than typical median richness of 20 taxa for a site in the lower reaches of an eastern hill country river (TRC 1999b (updated, 2012)). During the 2012-2013 period, spring (19 taxa) and summer (15 taxa) richnesses were relatively similar and from one to five taxa lower than this median richness.

MCI values also have had a moderate range (20 units) at this site, slightly narrower than typical of a site in the lower reaches of streams and rivers but reference is made to the relatively short monitoring period at this site. The median value (92 units) has been more typical of lower reach sites elsewhere and a significant 16 units higher than the median score recorded by 216 previous surveys at 'control' sites located at similar altitudes (to this site) in eastern hill country rivers and streams (TRC, 1999b (updated, 2012)). The spring, 2012 (100 units) and summer, 2013 (87 units) scores were dissimilar, ranging from 8 units above to 5 units below the historical median respectively. These scores categorised this site as having 'good' health generically (Table 1) in spring and 'fair' health in summer. The historical median score (92 units) placed this site in the 'fair' category for the generic method of assessment.

3.2.24.3.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 117.

Table 117 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Tangahoe River d/s of the railbridge between 1995 and March 2012 [11 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
NEMERTEA	Nemertea	3	1	9		
ANNELIDA	Oligochaeta	1	10	91		A
MOLLUSCA	*Latia*	5	3	27		
	Potamopyrgus	4	8	73		
CRUSTACEA	*Paracalliope*	5	1	9		
EPHEMEROPTERA	*Deleatidium*	8	2	18		
	Zephlebia group	7	1	9		
PLECOPTERA	*Zelandobius*	5	1	9		
COLEOPTERA	Elmidae	6	10	91	A	A
TRICHOPTERA	*Aoteapsyche*	4	10	91	VA	XA
	Pycnocentrodes	5	3	27	A	
DIPTERA	*Aphrophila*	5	4	36	A	A
	Maoridiamesa	3	3	27	A	
	Orthocladiinae	2	9	82	A	VA
	Austrosimulium	3	1	9		

Prior to the current 2012-2013 period, a moderate number of taxa (15) have characterised the community at this site on occasions due in part to the short duration of monitoring at this site. These have comprised one 'highly sensitive', seven 'moderately sensitive', and seven 'tolerant' taxa i.e. a relatively high proportion of 'tolerant' taxa as would be expected in the lower reaches of an eastern hill-country river. Predominant taxa have included one 'moderately sensitive' taxon [elmid beetles] and four 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), net-building caddisfly (*Aoteapsyche*), and orthoclad midges]. Three of these predominant taxa were dominant in the spring, 2012 community together with three other historically characteristic taxa. The summer, 2013 community was characterised by four of the taxa dominant in spring, together with one additional taxon ('tolerant' oligochaete worms) which previously had been characteristic of this site's communities (Table 117). Increased abundances within two 'tolerant' taxa resulted in the lower summer SQMCI$_s$ score by 0.7 unit (Tables 172 and 173). The two taxa recorded as very or extremely abundant during spring and/or summer had characterised this site's communities on 82% to 91% of the past surveys.

3.2.24.3.3 Predicted stream 'health'

The Tangahoe River site downstream of the railbridge, at an altitude of 15 m asl, is in the lower reaches of a river draining an eastern hill country catchment. A relationship for ringplain streams and river developed between MCI and altitude (Stark and Fowles, 2009) is therefore not appropriate for this river.

3.2.24.3.4 Temporal trends in 1995 to 2013 data

Non-parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has not been performed on the mainly six years of SEM results collected to date from the site in the Tangahoe River downstream of the railbridge due to the limited data record. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. However, a graphical presentation of LOWESS plot trends in MCI data is provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 113.

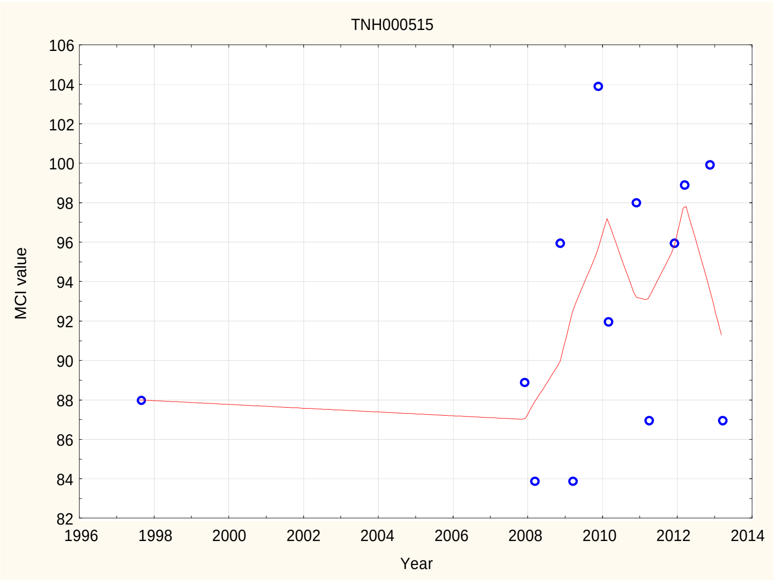


Figure 113 LOWESS trend plot of MCI data for the Tangahoe River site downstream of the railbridge

No temporal trend in MCI scores may yet be inferred for this lower river reach, eastern hill country site. The range of smoothed MCI scores (11 units) has bordered on ecologically significant but this significance cannot be properly assessed until the monitoring period is of sufficient duration and frequency for valid interpretation.

Smoothed MCI scores have indicated 'fair' generic river health (Table 1) over the short period to date (Figure 113).

3.2.24.4 Discussion

Seasonal MCI values atypically increased between spring and summer (by a significant 14 units) at the upper site (Upper Tangahoe Valley Road) where historical median seasonal values (Appendix II) have differed by only 4 units. At the Tangahoe Valley Road bridge site there was a typical summer decrease (7 units) which was 3 units more than the historical seasonal median decrease. At the railbridge site in the lower reaches, a typical summer decrease in MCI score (13 units) was recorded in comparison with the historical seasonal median decrease of 10 units for this site (Appendix II). Seasonal communities shared only 41% of the 34 taxa found at the upper reach (Upper Tangahoe Valley Road) site, 56% of 32 taxa at the Tangahoe Valley Road bridge site, and 48% of 23 taxa at the furthest downstream site in the lower reaches (railbridge), indicative of greater similarity in seasonal community composition at the mid-reach site, atypical of downstream trends of decreasing seasonal similarity found elsewhere.

The spring MCI scores atypically increased by a significant 11 units in a downstream direction over the 8.9 km reach between the upper and mid sites and also increased (by 7 units) between the upper and lower sites over a distance of 30.2 km (and decrease in elevation of 70 m). This trend was atypical of the average rate of decrease of 1.7 units/10 m assessed for large hill country catchments (TRC, 2006c). Summer MCI scores decreased overall between the upper and lower reach sites (by 20 units) representing a rate of decrease of 0.7 MCI unit/km or 2.9 MCI unit/10 m, the latter higher than the average of 1.7 units/10 m assessed for large hill country catchments (TRC, 2006c).

Using the long-term median SEM MCI scores for each site (Appendix II), there has been no decline, rather an improvement (0.8 unit/km), between the upper reach (Upper Tangahoe Valley Road) and the mid-reach (Tangahoe Valley Road bridge) sites. The rate of decline between the mid-reach site and lower reach (railbridge) site has been about 0.6 unit per km with an overall average rate of decline of 0.2 MCI unit/km over the surveyed length of the river. Therefore rates of MCI decline for the entire river length surveyed over the 2012-2013 period were lower (spring) and higher (summer) than the median rate for the short monitoring period prior to 2012.

Community composition varied markedly through the upper reach to lower reach length of the stream surveyed. A total of 31 taxa was recorded in spring of which only 12 taxa (39%) were present at all three sites (Table 110). These included one ‘highly sensitive’, seven ‘moderately sensitive’, and four ‘tolerant’ taxa with only one ‘moderately sensitive’ taxon (elmid beetles) and one ‘tolerant taxon (orthoclad midges) abundant at all three sites. A higher total of 42 taxa was found along the river’s length by the summer survey (Table 111) of which only nine taxa (21%) were present at all three sites. These included six of the widespread taxa in spring. Only one ‘moderately sensitive’ taxon (elmid beetles) was abundant at all three sites. These dissimilarities in spatial community structure along the surveyed length (upper reaches to lower reaches) of the Tangahoe River were more pronounced in summer than in spring.

3.2.25 Herekawe Stream

One site in this small coastal ringplain stream on the western perimeter of New Plymouth City was incorporated into the SEM programme in 2008 for the purpose of monitoring a newly-developed walkway and associated riparian planting initiatives in the lower reaches of the stream. Consent monitoring also has been performed at this 'control' site in spring and summer throughout the period from 1995 to 2008 (and dates back to 1986).

The results found by the 2012-2013 surveys are presented in Table 174 and Table 175, Appendix I for this small lowland stream.

3.2.25.1 Centennial Drive site (HRK000085)

3.2.25.1.1 Taxa richness and MCI

Thirty-four surveys have been undertaken in this lower-reach site in the Herekawe Stream between February 1995 and March 2012. These results are summarised in Table 118, together with the results from the current period, and illustrated in Figure 114.

Table 118 Results of previous surveys performed in Herekawe Stream at Centennial Drive, together with spring 2012 and summer 2013 results

Site code	SEM data (1998 to Mar 2012)					2012-2013 surveys			
	No of surveys	Taxa numbers		MCI values		Oct 2012		Feb 2013	
		Range	Median	Range	Median	Taxa no	MCI	Taxa no	MCI
HRK000085	34	13-23	18	68-97	89	21	99	23	93

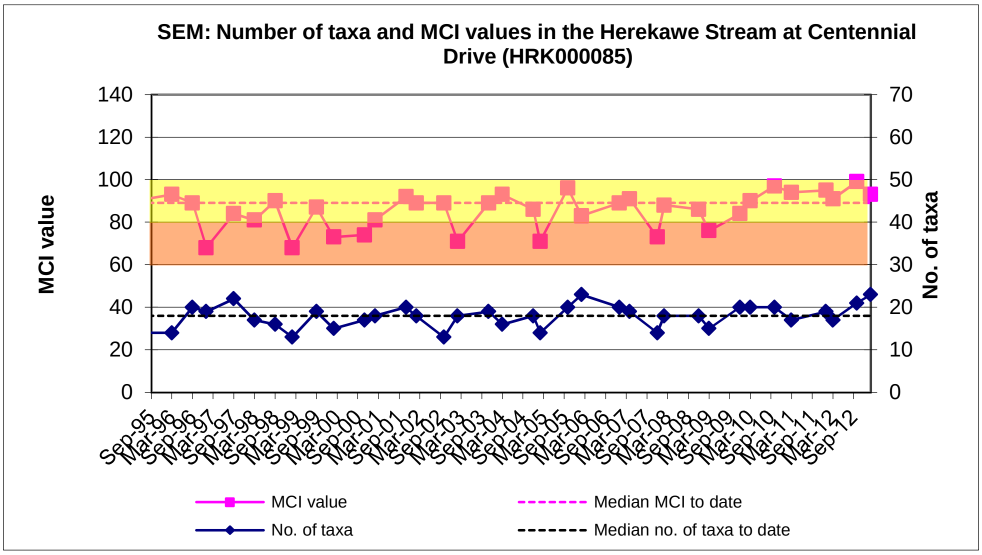


Figure 114 Numbers of taxa and MCI values in the Herekawe Stream upstream of Centennial Drive

A moderate range of richnesses (13 to 23 taxa) has been found, with a median richness of 18 taxa (more representative of typical richnesses in small coastal streams where a median richness of 17 taxa has been recorded from 182 previous surveys of 'control' sites at similar altitudes (TRC, 1999b (updated, 2013)). During the 2012-2013 period, spring (21 taxa) and summer (23 taxa) richnesses were similar and from 3 to 5 taxa higher than this median richness.

MCI values have had a relatively wide range (29 units) at this site. The median value (89 units) is above scores typical of lower reach sites elsewhere in lowland coastal streams however, and the spring, 2012 (99 units) and summer, 2013 (93 units) scores were higher than typical for such a site. These were 10 and 4 units higher than the historical median in spring and summer respectively, with the spring score two units higher than all scores previously recorded at this site.

These scores categorised this site as having 'fair' (spring and summer) health generically (Table 1) and both scores were significantly higher (Stark, 1998) than the median MCI score (78 units) recorded by 182 previous surveys of 'control' sites below 25 m asl in small, coastal ringplain streams in Taranaki (TRC, 1999b (updated, 2012)). The historical median score (89 units) placed this site in the 'fair' category for the generic method of assessment and was also significantly higher than the median score recorded at similar sites elsewhere in lowland coastal streams.

3.2.25.1.2 Community composition

Characteristic macroinvertebrate taxa in the communities at this site prior to the 2012-2013 period are listed in Table 119.

Table 119 Characteristic taxa (abundant, very abundant, extremely abundant) recorded in the Herekawe Stream at Centennial Drive between 1998 and March 2012 [34 surveys], and by the spring 2012 and summer 2013 surveys

Taxa List		MCI Score	Total abundances	% of Surveys	Surveys	
					Spring 2012	Summer 2013
ANNELIDA	Oligochaeta	1	23	68	A	A
MOLLUSCA	*Potamopyrgus*	4	34	100	VA	XA
CRUSTACEA	Ostracoda	1	2	6		
	Paracalliope	5	28	82	A	XA
EPHEMEROPTERA	*Austroclima*	7	2	6		
	Coloburiscus	7	4	12	A	
PLECOPTERA	*Acroperla*	5	1	3		
TRICHOPTERA	*Oxyethira*	2	9	26		VA
	Triplectides	5	12	35		
DIPTERA	*Aphrophila*	5	2	6		
	Orthocladiinae	2	20	59		
	Austrosimulium	3	13	38		

Prior to the current 2012-2013 period, 12 taxa had characterised the community at this site on occasions. These have comprised six 'moderately sensitive' and six 'tolerant' taxa i.e. an absence of 'highly sensitive' taxa and a relatively high proportion of 'tolerant' taxa as would be expected in the lower reaches of a small coastal stream.

Predominant taxa have included only the one 'moderately sensitive' taxon [amphipod (*Paracalliope*)] and three 'tolerant' taxa [oligochaete worms, snail (*Potamopyrgus*), and orthoclad midges].

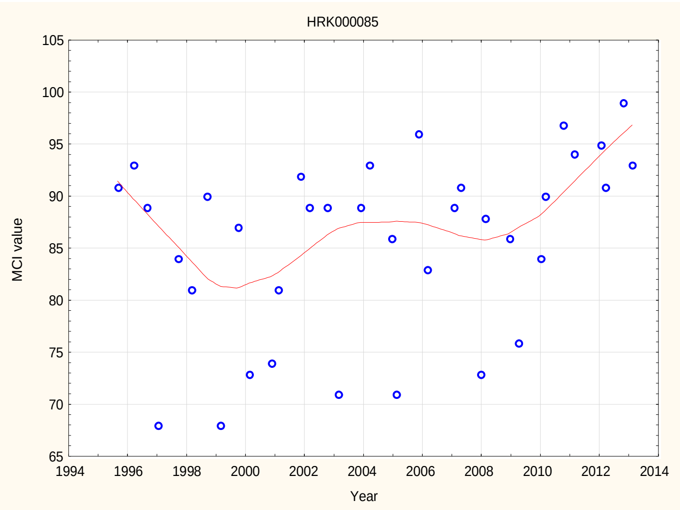
Four of the historically characteristic taxa were dominant in the spring, 2012 community and comprised three of the predominant taxa (above) together with another one 'moderately sensitive' taxon which previously had been characteristic of this site's communities (Table 119). The summer, 2012 community was characterised by three of the taxa dominant in spring together with one 'tolerant' taxon previously characteristic of this community on occasions. The summer increases in abundances within both 'moderately sensitive' and 'tolerant' taxa resulted in identical $SQMCI_s$ scores between seasons (Tables 174 and 175). The three taxa which were recorded as very or extremely abundant during spring and summer had characterised this site's communities on 26% to 100% of past surveys.

3.2.25.1.3 Predicted stream 'health'

The Herekawe Stream rises as seepage near the coast on the ringplain and the site at Centennial Drive, Omata is in the lower reaches near the mouth at an altitude of 5 m asl. Relationships for ringplain streams developed between MCI and distance from the National Park or site altitude (Stark and Fowles, 2009), therefore are not applicable in this type of small lowland coastal stream.

3.2.25.1.4 Temporal trends in 1995 to 2013 data

Non- parametric statistical trend analysis of MCI data (Stark and Fowles, 2006) has been performed on the 18 years of results collected to date from the site in the Herekawe Stream at Centennial Drive. The MCI has been chosen as the preferable indicator of 'stream/river health' for SEM trend reporting purposes. A graphical presentation of LOWESS plot of trends in MCI data and Mann-Kendall test of significance are provided for this site. The LOWESS (tension 0.4) trend plot of MCI data is presented in Figure 115.



N = 36
Kendall tau = +0.264
p level = 0.024 [>FDR, p = 0.048]
Significant at $p < 0.05$ level
N/S at $p < 0.01$

Figure 115 LOWESS trend plot of MCI data at the Centennial Drive site

The positive temporal trend in MCI scores over the monitoring period has become statistically significant at $p < 0.05$ (but not at $p < 0.01$)at this site in the lower reaches of the stream immediately downstream of the more recently constructed walkway. Trends have varied at this site over the eighteen year period with wide variation in individual MCI scores although the range of LOWESS-smoothed scores (16 units) has been ecologically significant.

Smoothed MCI scores have consistently remained indicative of 'fair' stream health throughout the monitoring period.

3.2.25.2 Discussion

Seasonal MCI values typically decreased between spring and summer (by 6 units) at this lower reach site which may be compared with the median seasonal summer decrease of 3 units for the eighteen year period (Appendix II). The percentage composition of 'tolerant' taxa typically increased by 14% in the summer community when periphyton substrate cover was slightly greater under lower flow and warmer temperature conditions. Seasonal communities at this site shared 14 common taxa (44% of the 30 taxa found at this site in 2012-2013)), a relatively low percentage of common taxa.

4. General discussion and conclusions

The detection of trends in the biological data requires a data set of suitable period and collected using rigid, acceptable protocols, to be statistically valid e.g., 10 years of spring and summer surveys. With eighteen years of data available for most sites, temporal trend analyses have been updated further within this report. Other comments in relation to the data collected in the period 1995 to 2013, are presented briefly below. These data are summarised in Appendix II and illustrated in Figures 116 to 123.

4.1 Macroinvertebrate fauna communities

In general terms, data have indicated that the macroinvertebrate communities at sites in upper reaches of catchments have been comprised of a greater proportion of taxa that are 'sensitive' to the effects of organic pollution than proportions which comprised the sites' communities in the mid and lower reaches of catchments. These changes in community composition have resulted from the effects of organic enrichment, higher temperatures, increased algal growth (a consequence of the former), and finer substrate (sedimentation), coincident with poorer physicochemical water quality in the lower reaches of streams and rivers.

Taxa richnesses at most sites in these streams and rivers more often showed higher richnesses in the upper reaches of catchments (with the exception of those affected by preceding headwater erosion events) but more seasonal variability in richnesses further downstream. Summer richnesses have tended to be higher than spring richnesses particularly at lower reach sites.

Over the eighteen year period, sites in the middle and the lower reaches of streams and rivers generally have had lower summer MCI scores than spring MCI scores as evidenced by decreases in median scores by 4 and 5 units respectively, whereas median seasonal scores at upper reach sites have only differed by one unit. This difference has been coincident with summer warmer water temperatures and increased periphyton substrate cover, resulting in the loss or replacement of certain 'sensitive' taxa by lower scoring 'tolerant' taxa.

Furthermore, the results from the 2012-2013 period have shown that:

- over all sites, spring MCI scores were higher than summer scores and t-testing of the mean seasonal MCI difference (7 MCI units) showed that this was very significant ($p = 0.01$)
- at upper reach sites there was a decrease in average MCI score of nearly 3 units in summer which was statistically insignificant ($p = 0.19$)
- at mid reach sites, a decrease in average MCI score of 7 units in summer was significant ($p = 0.03$)
- at lower reach sites, a more marked decrease in average MCI score of 10 units in summer was also significant ($p < 0.01$), unlike the insignificant decrease (2 units) recorded in the 2011-2012 period
- at all sites, spring 2012 MCI scores were on average 11.5 units higher than long term (seventeen year) median scores, and this difference was significant at $p = 0.01$ ($p < 0.001$)

- at all sites, summer 2013 MCI scores were on average 3.2 units higher than long term (seventeen year) median scores, but t-tests showed that this difference was insignificant ($p = 0.30$).

It was noticeable that there were several (20) new maximum MCI scores recorded during the 2012-2013 period particularly during spring months which was coincident with the statistical significance of 2012-2013 results in relation to historical data as referenced above. None of these new maxima were recorded at upper reach sites, with ten at both middle and lower reach sites. These new maximum scores ranged from 1 to 10 units (median: 6 units) at mid reach sites and from 1 to 9 units (median: 2 units) at lower reach sites. Rivers and streams with multiple sites exceeding past maxima were the Punehu, Kaupokonui, Kapoaiaia, and Waiokura Streams and the Waingongoro River. No decreases in historical minimum MCI scores were recorded during the 2012-2013 period.

The trend of higher MCI scores during the 2012-2013 period resulted in increases (of up to 2 units) in historical median values at one upper reach site, eleven mid-reach sites, and two lower reach sites whereas there were two decreases in historical medians at upper reach sites and one at a mid reach site (of one MCI unit).

4.1.1 Spring surveys

4.1.1.1 Historical SEM

Fifty-six (of the 57) sites' faunal communities' spring 2012 MCI scores were either similar to, or better than, historical SEM medians for those sites (Figure 116). Significantly higher scores were found at a high proportion (twenty-nine) of sites coincident with most of these sites having reduced periphyton cover in comparison with many past surveys. Significantly lower scores were not found at any sites at the time of these spring surveys.

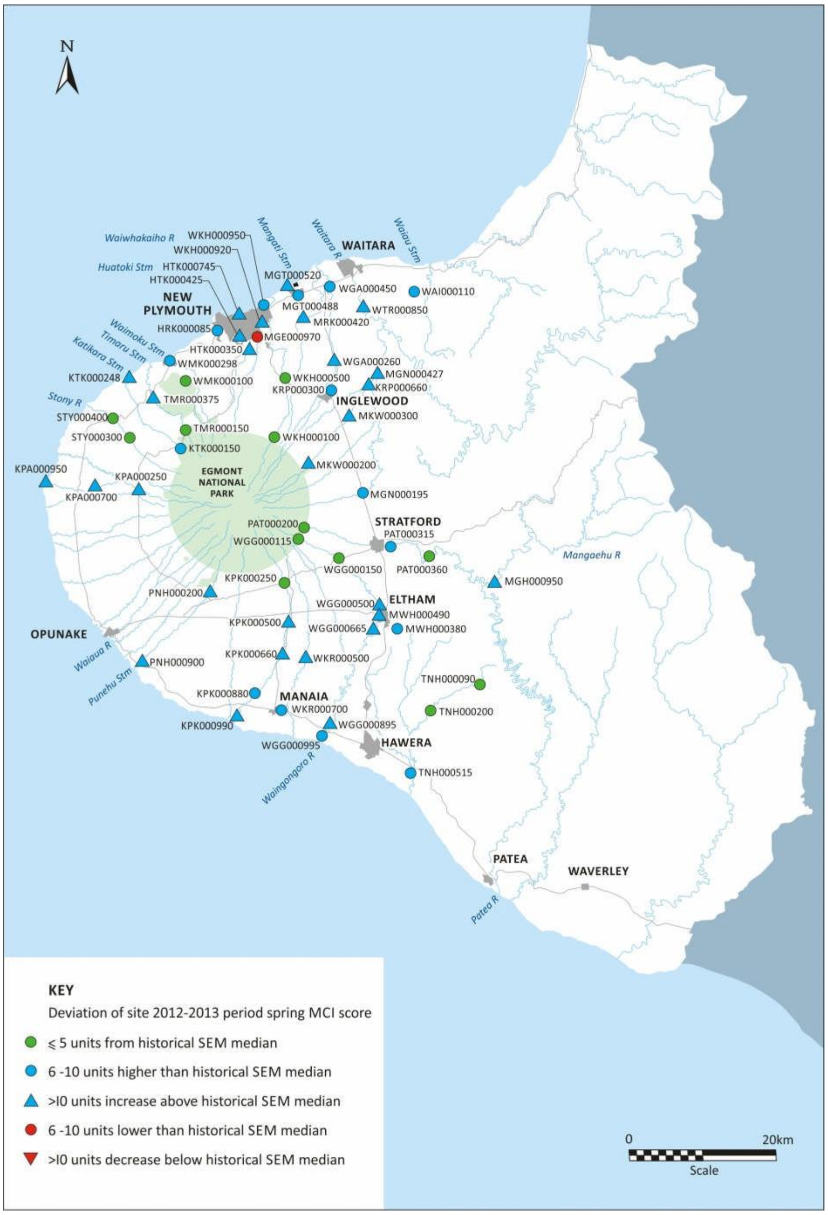


Figure 116 Spring 2012 MCI scores in relation to SEM historical median values

In summary, 49% of sites showed no significant detectable differences (Stark, 1998) between spring, 2012 MCI scores and historical median scores, while 51% of sites had significantly higher spring 2012 MCI scores. No sites had significantly lower spring 2012 scores.

4.1.1.2 Predictive TRC ringplain altitude/distance models

Predictive scores have been developed for ringplain sites in relation to altitude and distance from the National Park (Stark and Fowles, 2009). Spring scores for each ringplain site have been assessed against predicted scores for altitude in Figure 117 and against predicted scores for distance from the National Park boundary for ringplain sites with their sources inside the National Park in Figure 118.

4.1.1.2.1 Altitude

Few (two) sites had spring MCI scores more than 5 units below predicted values (Figure 117), neither of which were significantly lower than predicted. Nine sites had spring scores very similar to (within 5 units) predicted scores while the remaining 36 sites' scores were more than 5 MCI units above predicted scores for sites at equivalent altitudes. Of the latter, thirty-three sites had significantly higher MCI scores, a much higher proportion (70% of sites) than usually found.

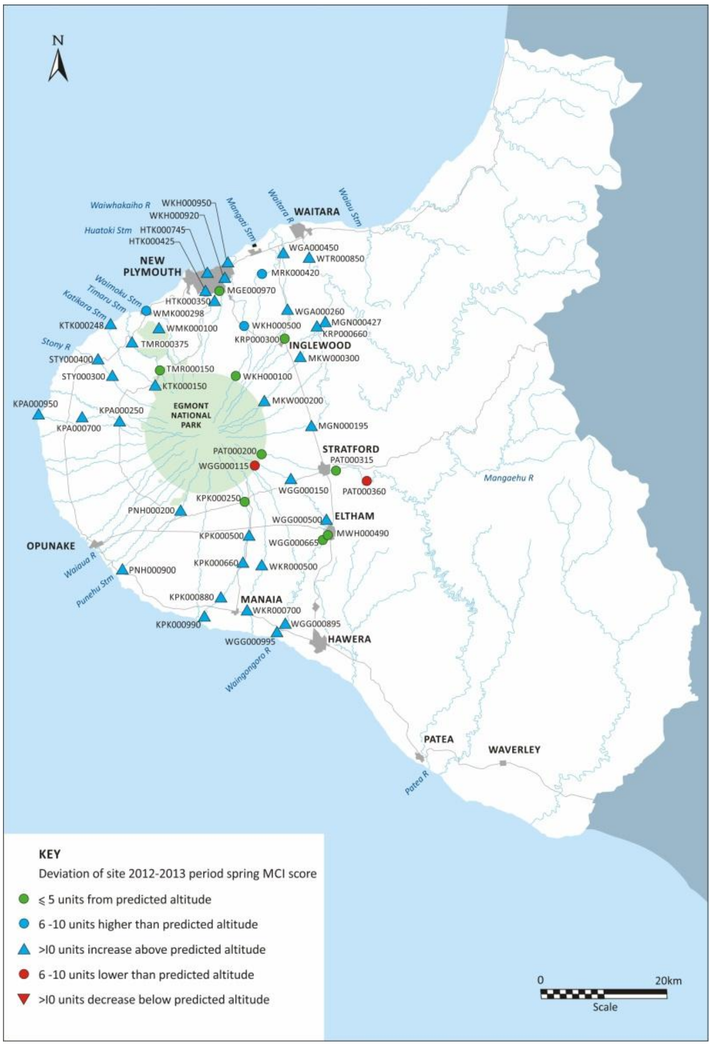


Figure 117 Spring 2012 MCI scores in relation to predicted altitude scores.

In summary, 30% of sites showed no significant detectable difference (Stark, 1998) between spring, 2012 scores and predicted altitude scores, while 70% of sites had significantly higher spring, 2012 MCI scores and no sites had significantly lower spring, 2012 scores.

4.1.1.2.2 Distance from National Park

Only one site had a spring MCI score more than 5 units below predicted values (Figure 118) and this site was significantly lower than predicted. This site was in the Waimoku Stream at the coast (due to the very short distance between the source and the coast). Ten sites had spring scores within 5 MCI units of predicted scores while twenty-six sites' scores were more than 5 units higher than predicted, a higher proportion (54%) of sites than has been typical to date.

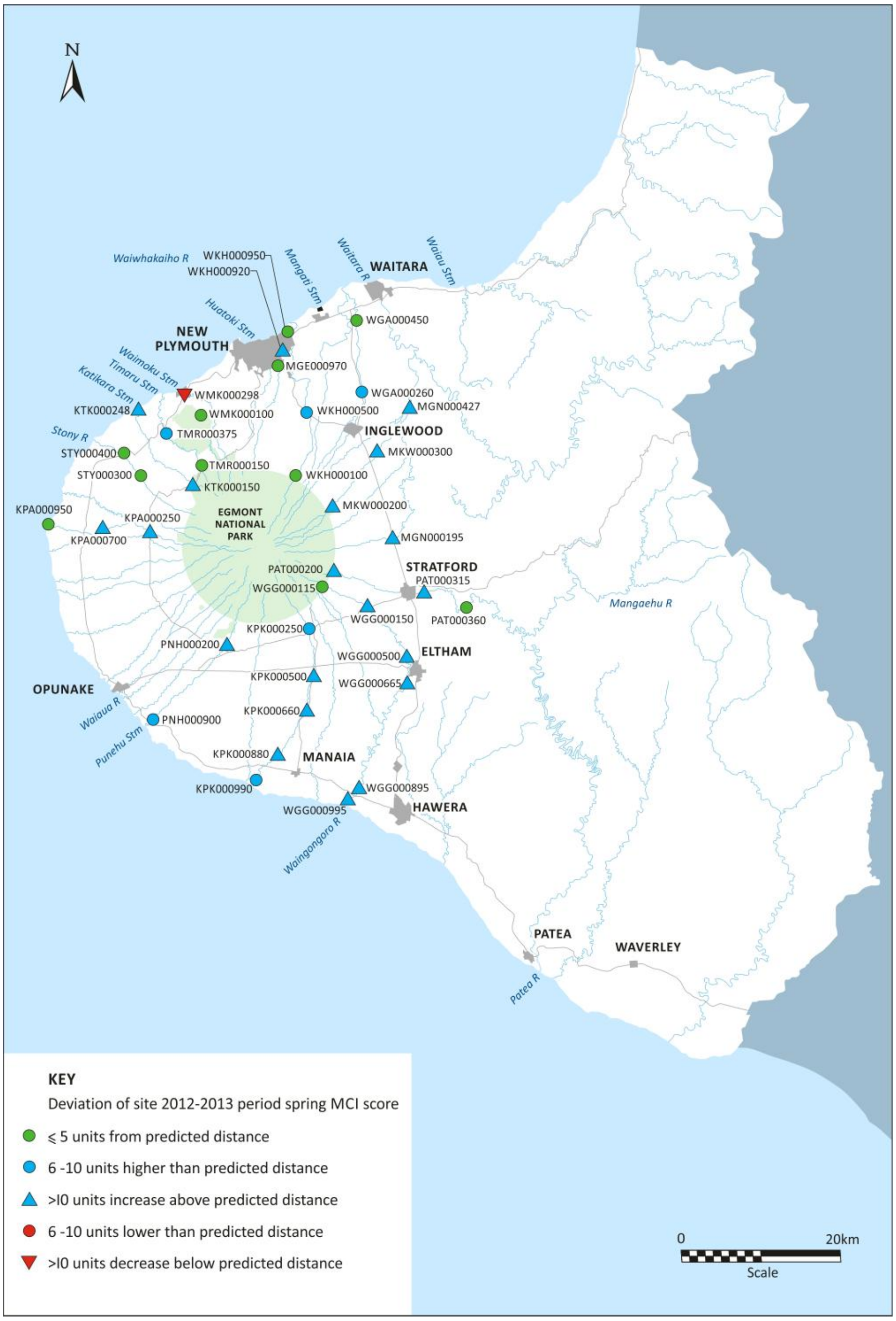


Figure 118 Spring 2012 MCI scores in relation to predicted downstream distance scores

In summary, 43% of sites showed no significant detectable difference (Stark, 1998) between spring, 2012 scores and predicted distance (from the National Park) scores, while 54% of sites had significantly higher spring, 2012 MCI scores and 3% of sites (one site) had significantly lower spring, 2012 scores.

4.1.2 Summer surveys

4.1.2.1 Historical SEM

A majority (46 of 57 sites) of sites' faunal communities' MCI scores were similar to (within 10 units) historical SEM medians for each site (Figure 119). Significantly higher scores were found at ten sites, while only one site showed a significantly lower MCI score following summer - autumn, relatively low flow conditions in the region.

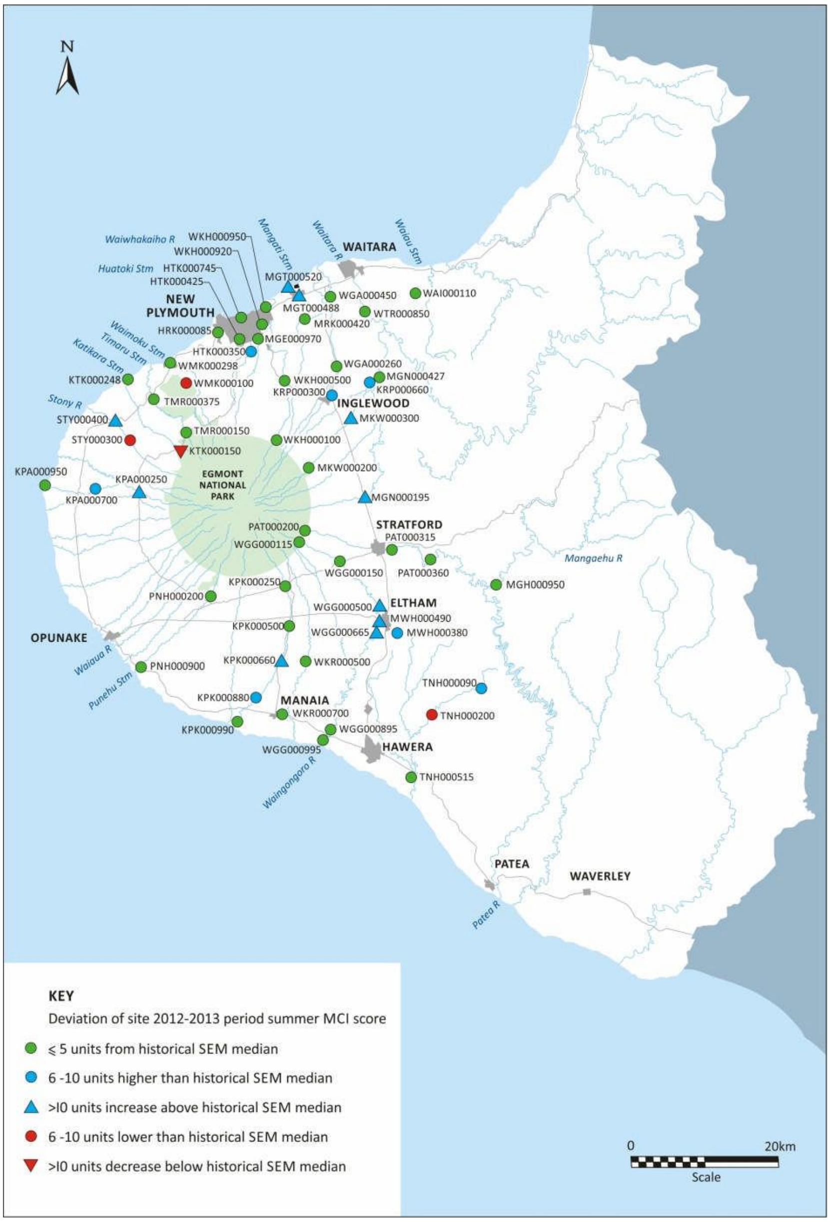


Figure 119 Summer 2013 MCI scores in relation to SEM historical median values

Significantly higher scores were found in the mid reaches of the Manganui and Waingongoro Rivers and Kaupokonui, Maketawa, Mangati, and Kapoaiaia Streams; and in the lower reaches of the Stony River and Kaupokonui, Mangawhero, and Mangati Streams. The one significantly lower score was found in the upper reaches of the Katikara Stream (see Section 3.2.20.1).

In summary, 81% of sites showed no significant detectable differences (Stark, 1998) between summer, 2013 MCI scores and historical median scores, while 18% of sites had significantly higher summer, 2013 scores.

Far fewer sites (52%) had significantly higher MCI scores (than historical medians) in summer than spring whereas only one spring and one summer sites' scores (2%) were significantly lower than historical medians. In summer, 5% of sites were 6 or more MCI units lower than historical medians compared to 2% in spring. In summer 30% of sites' scores were greater than 5 MCI units higher than historical medians compared to 63% of sites in spring, a more typical historical seasonal trend than had been recorded in the previous year.

4.1.2.2 Predictive TRC ringplain altitude/distance models

Summer scores for each ringplain site have been assessed against predicted scores (Stark and Fowles, 2009) for altitude (Figure 120) and for distance from the National Park boundary for those ringplain sites with sources inside the National Park (Figure 120).

4.1.2.2.1 Altitude

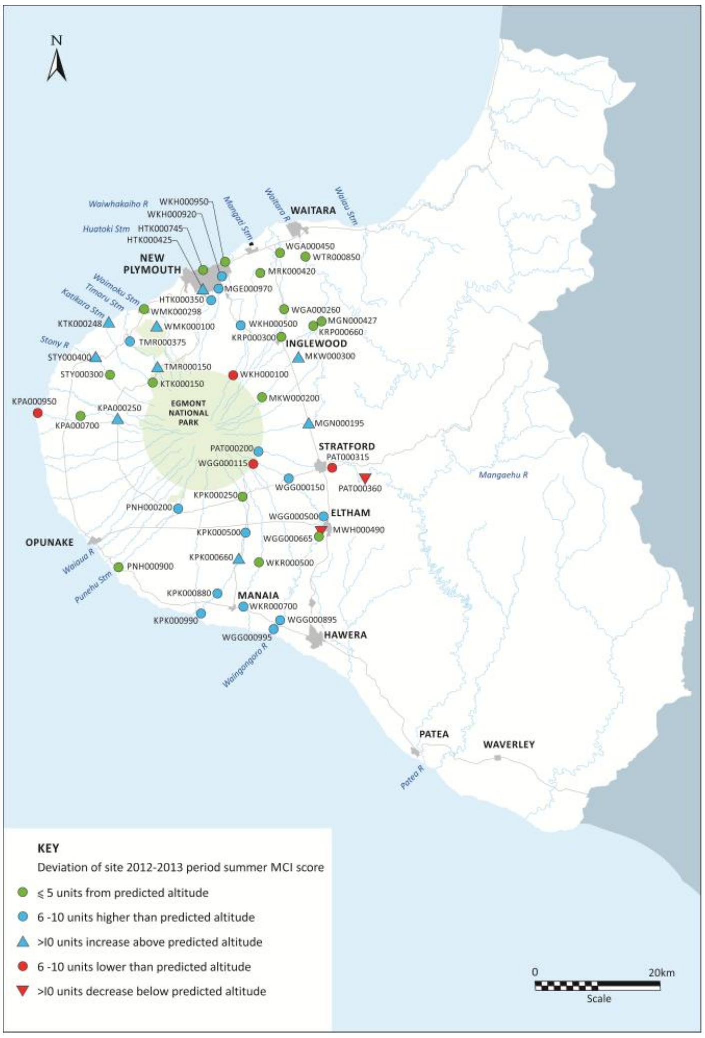


Figure 120 Summer 2013 MCI scores in relation to predicted altitude scores

Six sites had summer MCI scores greater than 5 units below predicted values, one of which (Mangawhero Stream near the Waingongoro River confluence) was downstream of the recently diverted Eltham municipal wastewater point source discharge. This site and a site in the mid reaches of the Patea River (downstream of the Stratford municipal wastewater discharge) were the only sites significantly below predictive values. Eighteen sites had scores very similar to (within 5 units) predicted scores (Figure 120), while twenty-three sites' scores were more than 5 MCI units above predicted scores for sites at equivalent altitudes. Nine sites had significantly higher MCI scores and these were situated in the upper reaches of the Waimoku and Timaru Streams; mid reaches of the Maketawa, Kapoaiaia, Kaupokonui, and Huatoki Streams and Manganui River; and in the lower reaches of the Katikara Stream and Stony River.

In summary, 77% of sites showed no significant detectable difference (Stark, 1998) between summer, 2013 scores and predicted altitude scores, while 19% of sites had significantly higher summer MCI scores and 4% of sites had significantly lower summer MCI scores. A much higher proportion (by 51%) of sites significantly exceeded the predictive scores in spring than in summer while there was a 4% difference between seasonally significant lower scores, all of which occurred in summer.

4.1.2.2.2 Distance from National Park

Seven sites (six more than in spring) had summer MCI score more than 5 units below predicted values (Figure 121) with three of these sites' scores (in the lower reaches of the Waimoku, Kapoaiaia, and Punehu Streams) significantly lower than predicted. Fifteen sites had summer scores within 5 units of predicted scores, while sixteen sites' scores (ten fewer than in spring) were more than 5 units higher than predicted. However, there were eleven sites with scores significantly higher than predicted, nine sites fewer than in spring. These sites were situated in the upper reaches of the Patea and Waingongoro Rivers; mid reaches of the Manganui and Waingongoro Rivers, and Kapoaiaia and Kaupokonui Streams; and lower reaches of the Maketawa Stream and Waingongoro and Stony Rivers (Figure 121).

In summary, 62% of sites showed no significant detectable difference (Stark, 1998) between summer, 2013 MCI scores and predicted distance (from National Park) scores, while 30% of sites had significantly higher summer scores and 8% of sites had significantly lower summer scores. A higher proportion (by 24%) of sites significantly exceeded predictive scores in spring while 5% more sites were significantly worse in summer.

Comments
The general seasonal trend in MCI scores is summarised in Table 120 which provides the percentages of sites' scores in relation to predicted scores for spring and summer surveys.

Table 120 Percentages of spring and summer MCI scores for ringplain sites in relation to Stark and Fowles (2009) predicted (altitude and distance from National Park) scores

Season	Spring 2012			Summer 2013		
Prediction	> 10 units lower	± 10 units	> 10 units higher	> 10 units lower	± 10 units	> 10 units higher
Altitude Distance	0 3	30 43	70 54	4 8	77 62	19 30

In general, while there were only small seasonal differences between seasons in sites' percentages of scores falling significantly below predicted scores (up to 5% fewer in spring), there was a much more marked decrease of 24 to 51% of sites' scores significantly exceeding predicted scores during the summer survey, the latter of which has been typical of seasonal trends reported for most SEM annual surveys to date [and far more marked than in the 2011-2012 period which, to a certain extent, was less typical of previous periods due to some spring survey delays as a result of very wet weather].

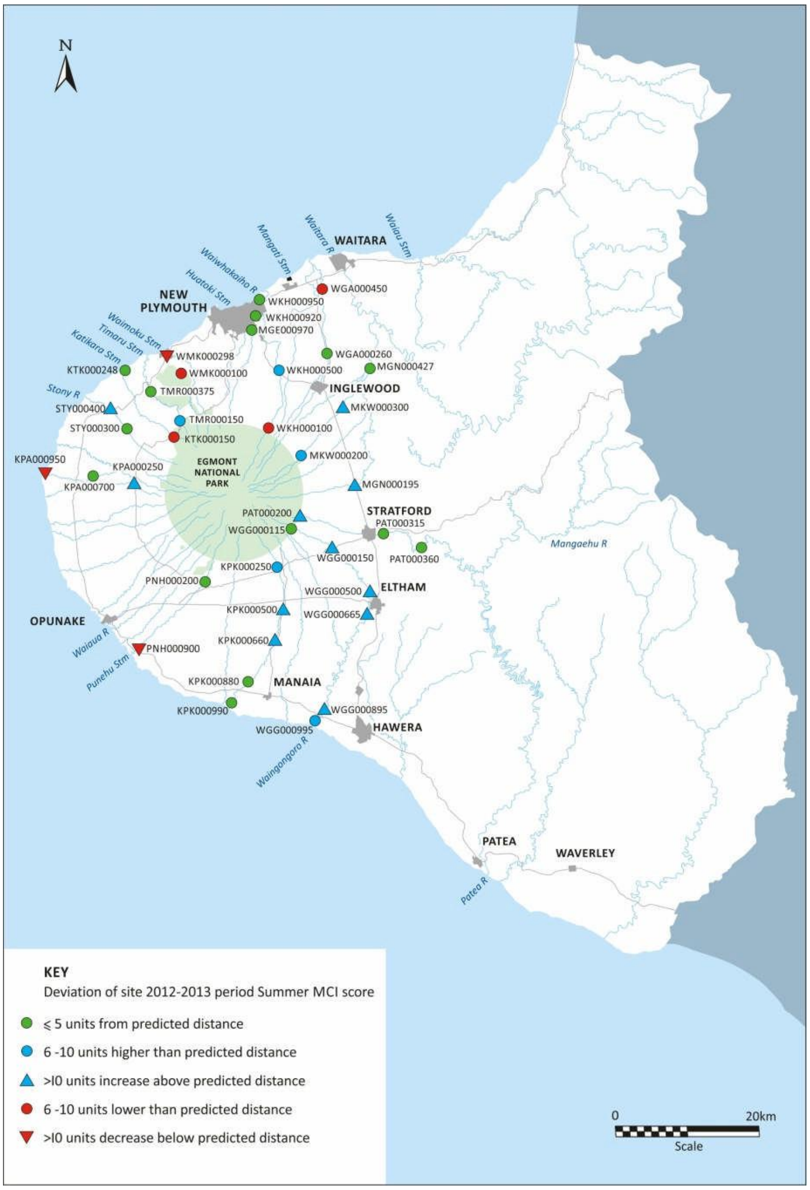


Figure 121 Summer 2013 MCI scores in relation to predicted downstream distance scores

4.1.2.2.3 General comments

Sites in the lower reaches of shorter ringplain streams (e.g. Punehu, Kapoaiaia and, in particular the Waimoku Stream), have had historical median MCI scores showing the greatest disparity between predicted scores for altitude and distance from the National Park than might be anticipated from such models (see Appendix II). These sites had wider ranges between the predicted altitude score and the predicted distance score (e.g. 31 units for the Waimoku Stream at Lucy's Gully and at the coast, 12 units for the Punehu Stream at SH45, and 10 units for the Kapoaiaia Stream at Cape Egmont).

Consideration must therefore be given to selection of the most appropriate predictive score which should be applied to a site in each case, assessed against length of catchment and site location, as the models developed by Stark and Fowles, 2009 utilised the historical macroinvertebrate 'control' sites database for the entire ringplain.

4.1.3 SEM MCI in relation to various predictive scores

In addition to the relationships established for MCI scores and ringplain streams sites' altitude and distance from the National Park by Stark and Fowles (2009), Leathwick (2009, pers comm.) has developed predictive scores based upon the River Environmental Classification (REC) system for New Zealand rivers and streams (Snelder et al, 2004). REC classifies and maps river and stream environments in a spatial framework for management purposes. It provides a context for inventories of river/stream resources and a spatial framework for effects assessment, policy development, developing monitoring programmes, and interpretations of state of the environment reporting.

Median MCI scores from the eighteen year SEM period (1995-2013) have been compared with the REC predictions for all 57 sites in Figure 122 and in Appendix II.

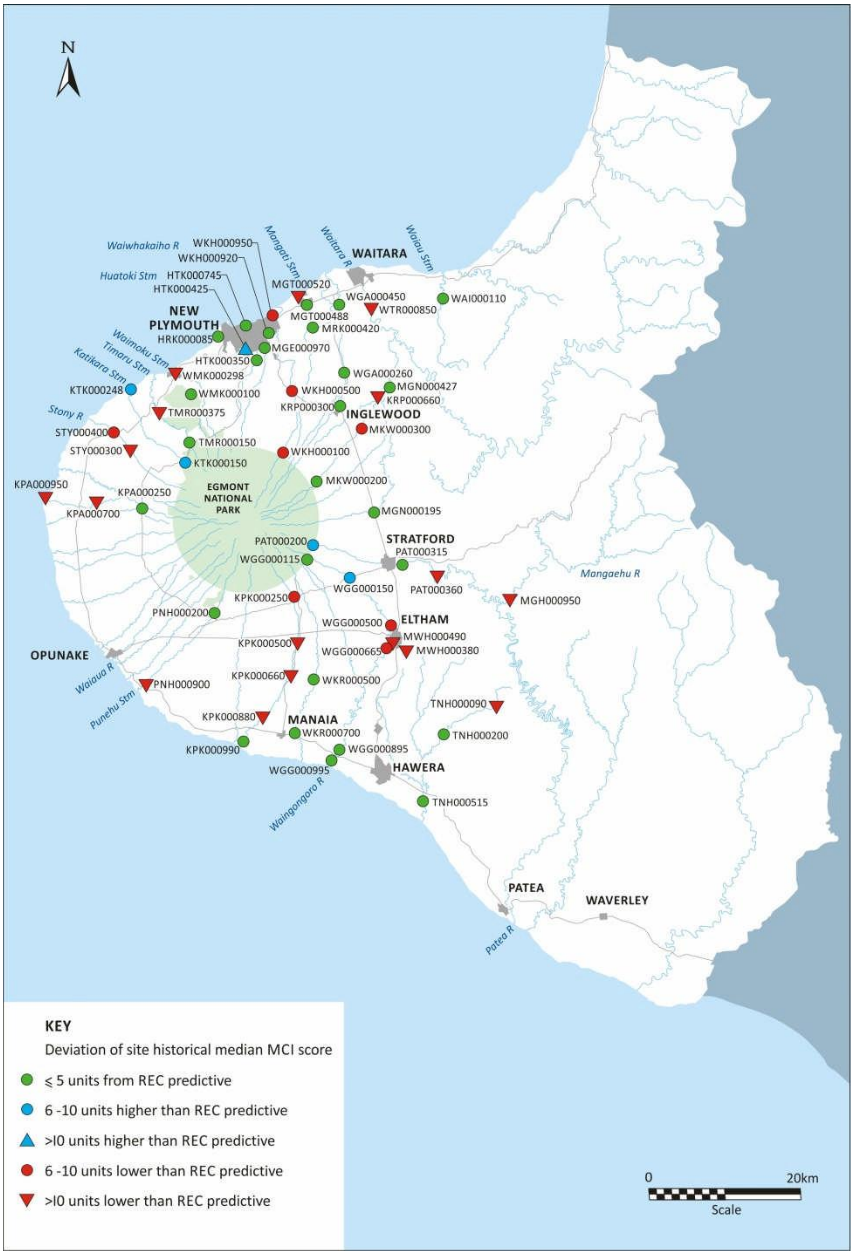


Figure 122 SEM historical (1995-2013) median values in relation to REC predictive values

Overall, this comparison indicates that only 5 sites (9%) have had median scores more than 5 units above the REC predictions, two of which are in the small Katikara Stream, and the others in the upper reaches of the Patea and Waingongoro Rivers and in the mid reaches of the Huatoki Stream (within the riparian vegetation of the Huatoki Domain, New Plymouth). Twenty-five sites (44%) were within 5 MCI units of predicted scores and 47% of sites (27) were more than 5 units below predicted REC scores. Of these lower scores, 18 sites had scores significantly lower than REC predictions with these situated in the mid reaches of the Kurapete, Timaru, Mangawhero, Kapoaiaia, and Kaupokonui Streams, and the Stony, Patea and Tangahoe Rivers; and the lower reaches of the Mangati, Waimoku, Kurapete, Kapoaiaia, Mangawhero, Punehu, and Kaupokonui Streams and the Waitara and Mangaehu Rivers. In terms of the 2012-2013 survey period; during spring, twelve sites significantly exceeded REC predictions while only four sites had significantly lower MCI scores; and during summer, six sites significantly exceeded the REC predictive scores and thirteen sites' scores were significantly lower than predicted.

The MCI scores from the eighteen year duration (1995-2013) of the SEM programme to date have been summarised in Appendix II and the median scores for all sites used to assess any deviations from those scores predicted by each of the three variables where relationships have been established (i.e. ringplain altitude and distance from the National Park, and REC [national]). Those sites' median MCI scores which deviated significantly (> 10 MCI units) from predicted scores are summarised in Table 121 and listed individually in Appendix II.

Table 121 Median SEM scores (1995-2013) showing significant differences (> 10 MCI units) from predicted scores

Sites	Deviation from predicted scores					
	Altitude[1]		Distance[1]		REC[2]	
	Lower	Higher	Lower	Higher	Lower	Higher
Upper reaches	0%	14%	0%	14%	0%	0%
Mid reaches	5%	10%	0%	19%	24%	4%
Lower reaches	5%	10%	20%	0%	40%	0%
All sites	4%	11%	8%	11%	28%	2%

[Notes: Stark and Fowles, 2009[1]; Leathwick, 2009[2]]

In summary, 15% of all sites median MCI scores differed significantly from the predictions based upon altitude on the ringplain with the majority of these higher than predicted. 19% of sites' median scores differed significantly from predictions based on distance from the National Park boundary with a slightly greater proportion higher than predicted although there was a marked downstream difference. No individual site's median MCI score differed significantly from both the predicted altitude and distance scores (Appendix II). There were no significantly lower median scores in either category situated in the upper reaches of rivers and streams on the ringplain, but a significant proportion of lower catchment sites had lower median scores than predicted by distance.

Only one median MCI score (Huatoki Stream at the Domain, New Plymouth) significantly exceeded predicted scores based upon the REC system, whereas 28% of sites' scores were significantly lower, increasing in a downstream direction from none in the upper reaches through 24% in the mid reaches to 40% of sites in the lower reaches. Interestingly, relatively few sites' median scores exceeded the REC predictions in any reaches (fourteen sites in total and most only by a few MCI units). It should be noted that SEM median MCI scores effectively incorporate equal proportions of spring and summer scores and that the maximum scores for each site (over the 1995 to 2013 period) (invariably recorded in spring) have often exceeded the REC predicted scores. Those sites where maximum scores have remained 5 or more units below REC predictions are situated in the lower reaches of the Mangati Stream, lower reaches of the Mangaehu River, and mid reaches of the Mangawhero Stream. The proportion of sites where the maximum SEM MCI scores over the eighteen years to date has significantly (11 units or more) exceeded the REC predicted scores (44%) includes 16% of sites located in the lower reaches of catchments.[Note: This exceedance increased by four sites over the 2012-2013 period].

It should be noted that the REC model predicts scores which would be expected for the best possible conditions for those locations, while Stark and Fowles' (2009) altitude and distance relationships were based on observed conditions at "control" sites.

Ranking sites, on the basis of median SEM MCI scores for the eighteen year period to date, may be attempted in terms of deviation from the predicted scores for distance from the National Park boundary (for ringplain sites) and REC predicted scores (for all sites). Table 122 provides the rankings on this basis of the best and poorest sites in the SEM programme.

Table 122 Ranking of sites' median MCI scores (1995-2013) based on deviation from predictive scores

	Distance from National Park	REC
BEST	Manganui R. SH3 (m)	Huatoki S @ Domain (m)
	Waingongoro R @ Opunake Rd (m)	Patea R @ Barclay Rd (u)
	Patea R @ Barclay Rd (u)	Katikara S @ Carrington Rd (u)
	Kaupokonui S @ Opunake Rd (u)	Katikara S @ coast (l)
	Waingongoro R @ SH45 (l)	Waingongoro R @ Opunake Rd (m)
POOREST	Waimoku S @ coast (l)	Mangaehu Rd @ Raupuha Rd (l)
	Punehu S @ SH 45 (l)	Mangati S @ Bell Block (l)
	Kapoaiaia S @ coast (l)	Kaupokonui S @ u/s Lactose (m)
	Kapoaiaia S @ Wataroa Rd (m)	Mangawhero S @ Eltham (m)
		Kaupokonui S @ Glenn Road (l)
		Mangawhero S @ d/s of Mangawharawhara S. (l)
		Stony R @ Mangatete Road (m)
		Timaru S @ SH 45 (l)

[Note: u = upper; m = middle; l = lower reaches]

The majority of the best ranked sites are located in the upper reaches and mid reaches of catchments. The Huatoki Stream in the Domain at New Plymouth has an extensive riparian cover provided by the Domain, but is excluded from the distance ranking as this stream is sourced outside of the National Park.

The majority of the poorest ranked streams are located in the lower reaches of catchments with the Kapoaiaia Stream (with very limited riparian cover) notable for its poor ranking at two sites. The Mangaehu River and the two small, non-ringplain sourced streams (Mangati and Mangawhero), which used to and/or continue to, receive significant point source discharges rank poorly in terms of the REC predictions. (Note: these streams and river sites are excluded from the distance predictive rankings as these catchments are located well away from the National Park).

4.1.4 Stream 'health' categorisation

A gradation of biological water quality conditions based upon ranges of MCI scores (see Page 3) has been used to determine the 'health' generically (Table 1) and predictively (Table 2) of each site by utilising the median score from the eighteen year period (1995-2013). These assessments are summarised in Appendix II and illustrated in Figure 123. The 'health' of streams in relation to the location of sites (upper, middle and lower reaches) in catchments is summarised in Table 123.

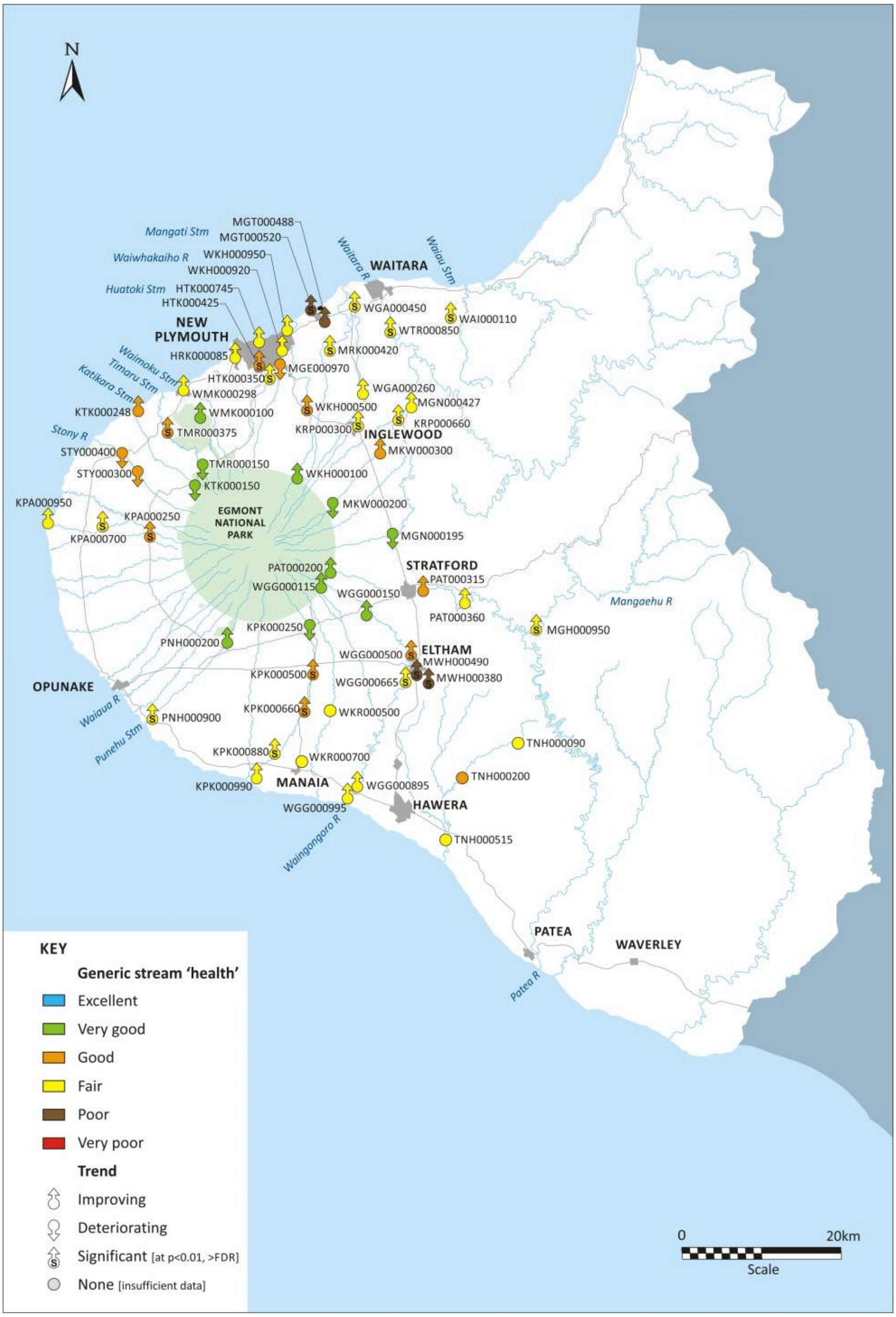


Figure 123 Generic biological 'health' (based on median MCI) and trends in biological quality for SEM sites, 1995 to 2013

Table 123 Stream 'health' assessments according to catchment reach (in terms of median MCI score)

'Health' grading	Reaches		
	Upper	Middle	Lower
Generic (Table 1)			
Excellent	0	0	0
Very good	7	4	0
Good	0	13	6
Fair	0	8	15
Poor	0	0	1
Very poor	0	0	0
Predictive (Table 2)			
Better than expected	0	2	2
Expected	7	20	17
Worse than expected	0	0	1
Median ranges (MCI units)	127-138 (11)	90-130 (40)	78-104 (26)

Typically generic 'health' (in terms of median MCI scores) decreases in a downstream direction from 'very good' in the upper reaches of catchments, through predominantly 'good-fair' in the middle reaches, to mainly 'fair' in the lower reaches toward the coast (Figure 123). In terms of predictive 'health', all but one of the gradings have varied between 'better than expected' and 'expected' through all reaches. Very few sites had 'better than expected' 'health' nor fell below 'expected' amongst the sites assessed. Each site's 'health' may vary between seasons, but seldom by no more than one category (grading) either side of this median grading in response to preceding stream flow and associated habitat (physical and physicochemical water quality) conditions. In this regard generally there has been more seasonal variability in scores at sites in the mid reaches of catchments.

4.1.5 Comments

This decreasing gradient of stream 'health', from 'very good' in the upper reaches of ringplain streams to 'fair' in the lower reaches, is indicative of a downstream progression of macroinvertebrate communities towards those that are comprised of taxa more 'tolerant' of organic enrichment and/or physical habitat deterioration in the lower reaches. These communities have become well adapted to the cumulative impacts of upstream point source discharges and non-point source diffuse run-off and are particularly resistant to further impacts (other than toxic discharges). Therefore, while some temporal trends may be detected in these lower reach communities, they are less likely to be of statistical significance and ecological significance (Figure 22). Thus, while maintenance of ('fair') stream 'health' occurs in the lower reaches of ringplain catchments (as these communities are very 'tolerant' of cumulative organic impacts), temporal trends of improvement in stream 'health' are unlikely to be statistically significant until appropriate management initiatives are substantially progressed on a catchment wide basis. Enhancement of stream health, particularly at these sites in the lower reaches of ringplain streams, is unlikely to occur until marked improvements in habitat and water quality occur. These may be implemented for instance by way of a combination of riparian fencing/planting initiatives and re-direction of dairy pond treatment system discharges from direct disposal into surface waters to irrigation to land.

4.2 Macroinvertebrate fauna MCI trends

Temporal trends measured over the monitoring period between 1995 and 2013 (Table 124, Figure 123, and Appendix II) indicated that 44 sites showed improving MCI scores during the period, 8 sites deteriorating scores, and five sites could not be trended due to the short duration of monitoring at these sites.

Table 124 Summary of Mann-Kendall test results for MCI (stream 'health') scores trended over time (1995-2013) for 52 Taranaki streams/rivers (p without FDR applied)

Site code	N	p-level	+/-(ve)	Significance
STY000300	39	0.155	-ve	N/S
STY000400	39	0.608	-ve	N/S
TMR000150	36	0.632	-ve	N/S
TMR000375	36	<0.0001	+ve	signif*
MRK000420	36	<0.0001	+ve	signif*
WGA000260	37	0.038	+ve	signif
WGA000450	36	<0.0001	+ve	signif*
WKH000100	22	0.511	+ve	N/S
WKH000500	36	0.003	+ve	signif*
WKH000920	37	0.103	+ve	N/S
WKH000950	32	0.161	+ve	N/S
MGE000970	20	0.457	-ve	N/S
MGN000195	38	0.222	-ve	N/S
MGN000427	36	0.398	+ve	N/S
MKW000200	27	0.641	-ve	N/S
MKW000300	26	0.061	+ve	N/S
WTR000850	36	0.019	+ve	signif
MGT000488	37	0.407	+ve	N/S
MGT000520	37	<0.0001	+ve	signif*
WMK000100	26	0.685	+ve	N/S
WMK000298	28	0.051	+ve	N/S
WAI000110	29	0.004	+ve	signif*
PNH000200	36	0.031	+ve	signif
PNH000900	36	0.0003	+ve	signif*
PAT000200	36	0.673	+ve	N/S
PAT000315	36	0.413	+ve	N/S
PAT000360	36	0.219	+ve	N/S
MGH000950	34	<0.0001	+ve	signif*
WGG000115	37	0. 224	+ve	N/S
WGG000150	37	0.362	+ve	N/S
WGG000500	39	0.0001	+ve	signif*
WGG000665	36	0.002	+ve	signif*
WGG000895	37	0.098	+ve	N/S
WGG000995	36	0.005	+ve	signif
MWH000380	36	0.001	+ve	signif*
MWH000490	36	0.0004	+ve	signif*
HTK000350	34	<0.0001	+ve	signif*
HTK000425	34	0.0002	+ve	signif*
HTK000745	34	0.672	+ve	N/S
KPK000250	29	0.907	- ve	N/S
KPK000500	32	0.0001	+ve	signif*
KPK000660	36	<0.0001	+ve	signif*
KPK000880	36	0.001	+ve	signif*
KPK000990	28	0.008	+ve	signif
KTK000150	28	0.432	-ve	N/S
KTK000248	27	0.024	+ve	signif
KPA000250	28	<0.0001	+ve	signif*
KPA000700	28	0.0002	+ve	signif*
KPA000950	28	0.160	+ve	N/S
KRP000300	37	<0.0001	+ve	signif*
KRP000660	37	0.0003	+ve	signif*
WKR000500	-	-	-	-
WKR000700	-	-	-	-
TNH000090	-	-	-	-
TNH000200	-	-	-	-
TNH000515	-	-	-	-
HRK000085	36	0.024	+ve	signif

[N/S = not statistically significant (ie p≥0.05),= significant before FDR (at p<0.05); = significant after FDR applied (at p < 0.05; * p< 0.01); -ve = negative trend, +ve = positive trend]

The majority of these trends were not statistically significant for the monitoring period (see also Appendix II). The following is a summary of significant trends for the SEM period to date:

– twenty-one sites with a positive very significant trend ($p \leq 0.01$ after FDR)
– five additional sites with a positive trend ($p<0.05$) but not significant ($p>0.01$ after FDR)
– no sites with a negative trend ($p<0.05$)

The sites have also been ranked in order of the significance of the strongest trends in Table 125.

Table 125 Ranking of sites in terms of significant temporal trends in MCI scores over the period 1995-2013 [significant trend at $p<0.05$ and $p<0.01$]

Site	Valid N	p-level	p-value (FDR corrected)	Trend	Ecological significance (LOWESS-smoothed range)
KPK000660	36	<<0.0001	<<0.0001	+ve	very high, 40 units
KRP000300	37	<<0.0001	<<0.0001	+ve	moderate, 16 units
MRK000420	36	<<0.0001	<<0.0001	+ve	moderate, 16 units
KPA000250	28	<<0.0001	<<0.0001	+ve	very high, 32 units
TMR000375	36	<<0.0001	<<0.0001	+ve	moderate, 19 units
MGH000950	36	<<0.0001	<<0.0001	+ve	moderate, 15 units
HTK000350	34	<<0.0001	<<0.0001	+ve	high, 20 units
MGT000520	37	<0.0001	<0.001	+ve	high, 24 units
KPK000500	32	0.0001	<0.001	+ve	high, 22 units
WGG000500	40	0.0001	<0.001	+ve	moderate, 14 units
WGA000450	36	0.0002	<0.001	+ve	moderate, 18 units
HTK000425	34	0.0002	<0.001	+ve	moderate, 15 units
KPA000700	28	0.0002	<0.001	+ve	high, 21 units
PNH000900	36	0.0004	0.001	+ve	moderate, 19 units
KRP000660	37	0.0004	0.001	+ve	moderate, 17 units
MWH000490	36	0.0004	0.001	+ve	high, 26 units
KPK000880	36	0.0013	0.004	+ve	high, 24 units
MWH000380	36	0.0015	0.004	+ve	low, 10 units
WGG000665	36	0.0018	0.005	+ve	moderate, 17 units
WKH000500	36	0.0028	0.007	+ve	high, 17 units
WAI000110	29	0.0035	0.009	+ve	moderate, 11 units
WGG000995	36	0.0047	0.011	+ve	moderate, 17 units
KPK000990	28	0.0083	0.019	+ve	moderate, 15 units

Each of these site's trends is discussed more fully earlier in the report. In general, all but three of these sites exhibited MCI score variabilities over the eighteen year SEM monitoring period which were ecologically significant, with nine sites showing variability of high ecological significance. Those sites with the strongest positive temporal improvement over the 18 year monitoring period, coupled with very significant ecological variability, have been:

– Kaupokonui Stream upstream of Fonterra, Kapuni factory
– Kapoaiaia Stream at Wiremu Road
– Mangawhero Stream upstream of Waingongoro River confluence
– Kaupokonui Stream at Glenn Road
– Huatoki Stream at Hadley Drive
– Kaupokonui Stream upstream of Kaponga WWTP
– Kapoaiaia Stream at Wataroa Road

- Mangati Stream at Bell Block

Five of these sites have illustrated particularly strong improvements over the most recent four to seven year period.

Slightly lower positive temporal improvements, but significant ecological improvement have been shown at the following sites:
- Timaru Stream at SH45
- Punehu Stream at SH45
- Waiongana Stream at SH3
- Kurapete Stream 6 km downstream of Inglewood WWTP
- Mangaoraka Stream at Corbett Road
- Kurapete Stream upstream of Inglewood WWTP
- Mangaehu Road at Raupuha Road
- Waiwhakaiho River at SH 3
- Waingongoro River at Stuart Road
- Huatoki Stream at Huatoki Domain

It is noted that although three Waingongoro River sites [at Eltham (upstream of the two former major point source discharges) and downstream at Stuart Road and at SH45] have shown significant positive trends ($p<0.01$), the two sites downstream of these former point source discharges and the lower river site have shown slightly greater ecological improvement.

5. Summary

These eighteenth spring and summer biomonitoring components of the established SEM programme were performed during the period from early October 2012 to November 2012 and February to mid March 2012 respectively. This report describes the macroinvertebrate fauna and microflora communities at 57 sites established through the Taranaki region (TRC, 1995b) including the additional riparian monitoring sites in the Katikara and Kapoaiaia Streams and the sites in the Maketawa Stream and Waiwhakaiho catchment with the two sites monitored for consent purposes in the Kurapete Stream also included. Sites in the Waiokura Stream and Tangahoe River were also added to the programme in the 2007-2008 period and a site in the lower Herekawe Stream in 2008-2009 (although this site has a lengthy historical consent monitoring record spanning the 1995 to 2008 period). Results are discussed in terms of macroinvertebrate community composition, richness and MCI scores, which are compared with prior SEM data, and stream 'health' is assessed using generic and predictive methodologies. Downstream spatial trends are also identified where possible, and results are discussed in relation to the historical Taranaki streams and river database (TRC, 1999 (updated, 2012) and TRC 2006c) where applicable and also in relation to more recently established relationships between site altitude and distance from the National Park (Stark and Fowles, 2009) and the REC system (J Leathwick, pers comm.). Discussion of temporal trends over the eighteen years of data collection is also provided for each site and causal assessments have been made where trends have been shown to be statistically significant and particularly where ecological significance has been high.

Temporal enhancement of stream 'health', particularly in the lower reaches of ringplain catchments (currently mainly in 'fair' condition), may not be expected to be significant until upstream initiatives (such as diversion to land irrigation of dairy shed wastes and riparian planting/fencing) are substantially implemented throughout catchments.

6. Recommendations from the 2011-2012 report

In the 2011-2012 report, it was recommended:-

1. THAT the freshwater biological macroinvertebrate fauna component of the SEM programme be maintained in the 2012-2013 monitoring year by means of a similar programme to that undertaken in 2011-2012.

2. THAT temporal trending of the macroinvertebrate faunal data be updated on an annual basis.

The programme followed Recommendation 1 in the 2012-2013 monitoring year and the temporal trend reporting was undertaken and included in the Annual Report.

7. Recommendations for 2013-2014

1. THAT the freshwater biological macroinvertebrate fauna component of the SEM programme be maintained in the 2013-2014 monitoring year by means of a similar programme to that undertaken in 2012-2013.
2. THAT temporal trending of the macroinvertebrate faunal data continues to be updated on an annual basis.

8. Acknowledgements

The Job Manager for the programme was Chris Fowles (Scientific Officer) who was the author of this Annual Report. Statistical analyses were provided by Alex Connolly and Fiza Hafiz (Scientific Officers) with all field sample collection performed by Chris Fowles and Bart Jansma (Scientific Officers). Macroinvertebrate sample processing was undertaken by Biosortid Ltd. (under contract to the Taranaki Regional Council) with appropriate quality control procedures initiated and reported by Taranaki Regional Council.

9. References

Biggs, BJF, 2000: New Zealand Periphyton Guideline: Detecting, Monitoring and Managing Enrichment of Streams. Prepared for Ministry for the Environment. NIWA, Christchurch, New Zealand. 122 pp.

Biggs, BJF and Kilroy C, 2000: Stream Periphyton Monitoring Manual. Published for Ministry for the Environment. NIWA, Christchurch, New Zealand. 228 pp.

Collier, KJ; Winterbourn, MJ, 2000 (eds.): New Zealand stream invertebrates: ecology and implications for management. NZ Limnological Society, Christchurch. 415pp.

Death, RG; 2000: Invertebrate-substratum relationships. In: Collier, KJ; Winterbourn, MJ eds. New Zealand stream invertebrates: ecology and implications for management. New Zealand Limnological Society, Christchurch. Pp 157-178.

Fowles, CR, 1987: Siltation of the Retaruke River 1987: Water quality aspects. Rangitikei-Wanganui Catchment Board Internal Report.

Ryan, PA, 1991: Environmental effects of sediment on New Zealand streams, a review. NZ Journal of Marine and Freshwater Research, Vol 25, 207-221.

Snelder, T, Biggs, B, Weatherhead, M, 224: New Zealand River Environment Classification User Guide. MfE publication. 145p.

Stark, JD, 1985: A macroinvertebrate community index of water quality for stony streams. Water and Soil Miscellaneous Publication No. 87.

Stark, JD, 1998: SQMCI: a biotic index for freshwater macroinvertebrate coded abundance data. *New Zealand Journal of Marine and Freshwater Research 32(1):* 55-66.

Stark, JD, 1999: An evaluation of Taranaki Regional Council's SQMCI biomonitoring index. Cawthron Report No. 472. 32pp.

Stark, JD, 2003: The water quality and biological condition of the Maketawa catchment. Cawthron Report No 742. 70pp.

Stark, JD 2000; Boothroyd, IKG, 2000: Use of invertebrates in monitoring. In Collier KJ; Winterbourn, MJ eds. New Zealand Stream Invertebrates: ecology and implications for management. NZ Limnological Society, Chch. Pp 344-373.

Stark, JD; Boothroyd, IKG; Harding, JS; Maxted JR; Scarsbrook, MR, 2001: Protocols for sampling macroinvertebrates in wadeable streams. New Zealand Macroinvertebrate Working Group Report No 1. Prepared for Ministry for the Environment. Sustainable Management Fund Project No 5103 57p.

Stark, JD and Fowles, CR, 2006: An approach to the evaluation of temporal trends in Taranaki state of the environment macroinvertebrate data. Cawthron Institute Report No 1135. 88p.

Stark, JD and Fowles, CR, 2009: Relationships between MCI, site altitude, and distance from source for Taranaki ring plain streams. Stark Environmental Report No 2009-01 47p.

Stark, JD and Maxted, JR, 2007: A user guide for the MCI. Cawthron Report No 1166. 56p.

TCC, 1984. Freshwater biology. Taranaki ring plain water resources survey. Taranaki Catchment Commission Report. 196p.

TRC, 1994: Regional Policy Statement for Taranaki. Taranaki Regional Council.

TRC, 1995a: Freshwater macroinvertebrate community data: a review of the results of biomonitoring surveys undertaken between 1980 and 1995. TRC internal report.

TRC, 1995b: Regional Monitoring Strategy for Taranaki Part II: Proposed State of the Environment Monitoring Programme. TRC internal report.

TRC, 1996a: State of the environment regional water quality monitoring network for Taranaki. Biological sampling techniques for freshwater rivers and streams. TRC internal report.

TRC, 1996b: A brief statistical summary of Taranaki freshwater macroinvertebrate surveys for the period January 1980 to July 1996. TRC internal report.

TRC, 1996c: State of the Environment - Taranaki Region 1996. Taranaki Regional Council.

TRC, 1997a: State of the Environment Procedures Document. TRC internal report.

TRC, 1997b: State of the Environment regional water quality monitoring network for Taranaki. Biological sampling techniques for freshwater rivers and streams. TRC internal report.

TRC, 1997c: Annual SEM Report 1995-96 Fresh water biological monitoring programme. Technical report 97-96.

TRC, 1998: Freshwater biological monitoring programme. Annual SEM Report 1996-97. Technical Report 97-100.

TRC, 1999a: Freshwater biological monitoring programme. Annual SEM Report 1997-98. Technical Report 99-06.

TRC, 1999b: Some statistics from the Taranaki Regional Council database (FWB) of freshwater macroinvertebrate surveys performed during the period from January 1980 to 31 December 1998 (SEM reference report). TRC Technical Report 99-17.

TRC, 2000: Fresh water biological monitoring programme Annual SEM Report 1998-99. Technical Report 99-90.

TRC, 2001: Fresh water biological monitoring programme Annual SEM Report 1999-2000, Technical Report 2000-40.

TRC, 2002: Fresh water biological monitoring programme Annual SEM Report 2000-2001, Technical Report 2001-87.

TRC, 2002: Fresh water biological monitoring programme Annual SEM Report 2001-2002, Technical Report 2002-46.

TRC, 2003: Taranaki – Our Place, Our Future, Report on the state of the environment of the Taranaki region – 2003. TRC, 206pp.

TRC, 2003: Fresh water biological monitoring programme Annual SEM Report 2002-2003, Technical Report 2003-18.

TRC, 2004: Fresh water biological monitoring programme Annual SEM Report 2003-2004, Technical Report 2004-23.

TRC, 2005: Fresh water biological monitoring programme Annual SEM Report 2004-2005, Technical Report 2005-72.

TRC, 2006a: Fresh water macroinvertebrate fauna biological monitoring programme Annual SEM Report 2005-2006, Technical Report 2006-94.

TRC, 2006b: An interpretation of the reasons for statistically significant temporal trends in macroinvertebrate (MCI) SEM data in the Taranaki region 1995-2005. TRC Internal Report.

TRC, 2006c: A review of macroinvertebrate monitoring data for large hill country catchments in the Taranaki region. TRC Internal Report.

TRC, 2007: Fresh water macroinvertebrate fauna biological monitoring programme Annual SEM Report 2006-2007, Technical Report 2007-22.

TRC, 2007a: Taranaki Regional Council freshwater biology methods manual Version 3. TRC Internal Report.

TRC, 2008: Fresh water macroinvertebrate fauna biological monitoring programme Annual SEM Report 2007-2008, Technical Report 2008-75.

TRC, 2009: Fresh water macroinvertebrate fauna biological monitoring programme Annual SEM Report: 2008–2009, Technical Report 2009-14.

TRC, 2009a: Taranaki-Where We Stand. State of the environment report. TRC, 282 p.

TRC, 2010: Fresh water macroinvertebrate fauna biological monitoring programme Annual SEM Report: 2009–2010, Technical Report 2010-16.

TRC, 2011: Fresh water macroinvertebrate fauna biological monitoring programme Annual SEM Report: 2010–2011, Technical Report 2011-38.

TRC, 2011: Freshwater physicochemical programme. State of the Environment Monitoring Annual Report 2010-2011. Technical Report 2011-47.

TRC, 2012: Fresh water macroinvertebrate fauna biological monitoring programme Annual SEM Report: 2011–2012, Technical Report 2012-18.

TRC, 2012: Freshwater physicochemical programme. State of the Environment Monitoring Annual Report 2011-2012. Technical Report 2012-27.

TRC, 2013: Freshwater physicochemical programme. State of the Environment Monitoring Annual Report 2012-2013. Technical Report 2013-49.

Wilcock RJ, Betteridge K, Shearman D, Fowles CR, Scarsbrook MR, Thorrold BS and Costall D, 2009 : Riparian protection and on-farm best management practices for restoration of a lowland stream in an intensive dairy farming catchment: a case study. NZJ of Marine and Freshwater Research 43: 803-818.

Appendix I

Macroinvertebrate faunal tables

Table 126 Macroinvertebrate fauna of the Stony River: spring SEM survey sampled on 5 October 2012

Taxa List		MCI score	1	2
	Site Number		1	2
	Site Code		STY000300	STY000400
	Sample Number		FWB12375	FWB12376
ANNELIDA (WORMS)	Lumbricidae	5	R	-
EPHEMEROPTERA (MAYFLIES)	*Coloburiscus*	7	R	-
	Deleatidium	8	A	VA
PLECOPTERA (STONEFLIES)	*Zelandoperla*	8	R	C
COLEOPTERA (BEETLES)	Elmidae	6	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	-
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	R	-
	Costachorema	7	C	C
	Hydrobiosis	5	R	-
	Pycnocentrodes	5	-	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	-	R
	Eriopterini	5	C	-
	Maoridiamesa	3	R	C
	Orthocladiinae	2	R	A
	Ephydridae	4	-	R
		No of taxa	12	8
		MCI	112	105
		SQMCIs	6.8	6.8
		EPT (taxa)	6	4
		%EPT (taxa)	50	50
'Tolerant' taxa	'Moderately sensitive' taxa		'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 127 Macroinvertebrate fauna of the Stony River: summer SEM survey sampled on 26 February 2013

Taxa List		MCI score	1	2
	Site Number		1	2
	Site Code		STY000300	STY000400
	Sample Number		FWB13110	FWB13111
CRUSTACEA	Talitridae	5	-	R
EPHEMEROPTERA (MAYFLIES)	*Deleatidium*	8	VA	XA
	Nesameletus	9	-	R
PLECOPTERA (STONEFLIES)	*Zelandoperla*	8	A	A
COLEOPTERA (BEETLES)	Elmidae	6	C	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C	A
	Costachorema	7	R	C
	Hydrobiosis	5	R	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	R	-
	Eriopterini	5	R	R
	Maoridiamesa	3	R	-
	Orthocladiinae	2	R	-
		No of taxa	10	9
		MCI	106	127
		SQMCIs	7.6	7.8
		EPT (taxa)	5	6
		%EPT (taxa)	50	67
'Tolerant' taxa	'Moderately sensitive' taxa		'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 128 Macroinvertebrate fauna of the Timaru Stream: spring SEM survey sampled on 5 October 2012

Taxa List	Site Number	MCI score	1	2
	Site Code		TMR000150	TMR000375
	Sample Number		FWB12373	FWB12374
ANNELIDA (WORMS)	Oligochaeta	1	-	R
EPHEMEROPTERA (MAYFLIES)	*Acanthophlebia*	9	R	-
	Ameletopsis	10	C	R
	Austroclima	7	R	C
	Coloburiscus	7	A	VA
	Deleatidium	8	XA	XA
	Neozephlebia	7	R	-
	Nesameletus	9	A	C
PLECOPTERA (STONEFLIES)	*Acroperla*	5	-	R
	Austroperla	9	R	R
	Stenoperla	10	C	-
	Zelandobius	5	VA	VA
	Zelandoperla	8	A	A
COLEOPTERA (BEETLES)	Elmidae	6	C	A
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	A
	Costachorema	7	C	C
	Hydrobiosis	5	C	C
	Hydrobiosella	9	R	-
	Neurochorema	6	-	C
	Orthopsyche	9	R	-
	Psilochorema	6	R	-
	Beraeoptera	8	C	VA
	Confluens	5	-	R
	Olinga	9	-	R
	Pycnocentrodes	5	-	VA
	Zelolessica	7	R	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	A
	Eriopterini	5	R	C
	Maoridiamesa	3	R	C
	Orthocladiinae	2	C	A
	Polypedilum	3	R	R
	Tanytarsini	3	-	C
	Empididae	3	R	C
	Tanyderidae	4	R	R
		No of taxa	27	28
		MCI	132	114
		SQMCIs	7.5	6.9
		EPT (taxa)	18	17
		%EPT (taxa)	67	61

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 129 Macroinvertebrate fauna of the Timaru Stream: summer SEM survey sampled on 26 February 2013

Taxa List	Site Code / Sample Number	MCI score	TMR000150 FWB13114	TMR000375 FWB13115
ANNELIDA (WORMS)	Oligochaeta	1	-	R
MOLLUSCA	*Potamopyrgus*	4	-	R
EPHEMEROPTERA (MAYFLIES)	*Ameletopsis*	10	R	-
	Austroclima	7	C	A
	Coloburiscus	7	A	A
	Deleatidium	8	VA	A
	Nesameletus	9	A	R
	Rallidens	9	-	R
PLECOPTERA (STONEFLIES)	*Austroperla*	9	C	-
	Megaleptoperla	9	C	-
	Stenoperla	10	R	-
	Zelandobius	5	A	-
	Zelandoperla	8	A	A
COLEOPTERA (BEETLES)	Elmidae	6	A	VA
	Hydraenidae	8	R	-
	Staphylinidae	5	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C	VA
	Costachorema	7	C	C
	Hydrobiosis	5	R	C
	Neurochorema	6	-	A
	Orthopsyche	9	C	-
	Psilochorema	6	R	-
	Alloecentrella	8	R	-
	Beraeoptera	8	A	C
	Olinga	9	R	-
	Oxyethira	2	-	R
	Pycnocentria	7	C	-
	Pycnocentrodes	5	R	C
	Triplectides	5	-	R
	Zelolessica	7	C	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA
	Eriopterini	5	-	R
	Maoridiamesa	3	R	A
	Orthocladiinae	2	A	A
	Tanytarsini	3	-	VA
	Empididae	3	-	C
	Muscidae	3	-	R
	Tabanidae	3	-	R
	Tanyderidae	4	-	R
		No of taxa	28	27
		MCI	138	103
		SQMCIs	7.0	4.9
		EPT (taxa)	21	13
		%EPT (taxa)	75	48
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa		

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 130 Macroinvertebrate fauna of the Mangaoraka Stream: spring SEM survey sampled on 5 October 2012

Taxa List	Site Code	MCI score	MRK000420
	Sample Number		FWB12360
ANNELIDA (WORMS)	Oligochaeta	1	A
MOLLUSCA	*Potamopyrgus*	4	C
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A
	Coloburiscus	7	C
	Deleatidium	8	VA
	Nesameletus	9	R
	Zephlebia group	7	C
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	A
COLEOPTERA (BEETLES)	Elmidae	6	VA
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A
	Costachorema	7	A
	Hydrobiosis	5	C
	Neurochorema	6	R
	Pycnocentria	7	C
	Pycnocentrodes	5	A
DIPTERA (TRUE FLIES)	*Aphrophila*	5	VA
	Maoridiamesa	3	C
	Orthocladiinae	2	A
	Polypedilum	3	R
	Tanytarsini	3	C
	Empididae	3	C
	Austrosimulium	3	C
	Tanyderidae	4	R
		No of taxa	24
		MCI	101
		SQMCIs	5.7
		EPT (taxa)	12
		%EPT (taxa)	50
'Tolerant' taxa	'Moderately sensitive' taxa		'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 131 Macroinvertebrate fauna of the Mangaoraka Stream: summer SEM survey sampled on 12 February 2013

Taxa List	Site Code	MCI score	MRK000420
	Sample Number		FWB13055
NEMERTEA	Nemertea	3	A
ANNELIDA (WORMS)	Oligochaeta	1	C
MOLLUSCA	*Potamopyrgus*	4	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	VA
	Coloburiscus	7	C
	Deleatidium	8	C
	Zephlebia group	7	R
COLEOPTERA (BEETLES)	Elmidae	6	XA
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	XA
	Hydrobiosis	5	A
	Neurochorema	6	C
	Oxyethira	2	R
	Pycnocentria	7	R
	Pycnocentrodes	5	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C
	Hexatomini	5	R
	Maoridiamesa	3	R
	Orthocladiinae	2	R
	Polypedilum	3	R
	Tanytarsini	3	A
	Empididae	3	C
	Muscidae	3	R
	Tanyderidae	4	R
		No of taxa	24
		MCI	92
		SQMCIs	5.1
		EPT (taxa)	9
		%EPT (taxa)	38
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 132 Macroinvertebrate fauna of the Waiongana Stream: spring SEM survey sampled on 5 October 2013

Taxa List	Site Code	MCI score	WGA000260	WGA000450
	Sample Number		FWB12369	FWB12370
ANNELIDA (WORMS)	Oligochaeta	1	-	C
	Lumbricidae	5	R	-
MOLLUSCA	*Potamopyrgus*	4	R	C
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R	C
	Coloburiscus	7	A	R
	Deleatidium	8	XA	VA
	Nesameletus	9	R	-
	Zephlebia group	7	R	-
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	C	C
COLEOPTERA (BEETLES)	Elmidae	6	VA	A
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	R	C
	Costachorema	7	A	C
	Hydrobiosis	5	C	C
	Neurochorema	6	R	R
	Beraeoptera	8	R	-
	Pycnocentria	7	R	-
	Pycnocentrodes	5	A	A
DIPTERA (TRUE FLIES)	*Aphrophila*	5	VA	A
	Eriopterini	5	R	R
	Maoridiamesa	3	A	A
	Orthocladiinae	2	VA	A
	Tanytarsini	3	-	R
	Empididae	3	R	R
	Austrosimulium	3	R	-
	Tanyderidae	4	R	-
		No of taxa	24	19
		MCI	110	98
		SQMCIs	6.5	5.9
		EPT (taxa)	13	9
		%EPT (taxa)	54	47

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 133 Macroinvertebrate fauna of the Waiongana Stream: summer SEM survey sampled on 6 March 2013

Taxa List	Site Code	MCI score	WGA000260	WGA000450
	Sample Number		FWB13142	FWB13143
NEMERTEA	Nemertea	3	R	R
NEMATODA	Nematoda	3	-	R
ANNELIDA (WORMS)	Oligochaeta	1	A	C
	Lumbricidae	5	-	R
MOLLUSCA	*Potamopyrgus*	4	C	XA
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	C
	Coloburiscus	7	R	R
	Deleatidium	8	XA	-
	Nesameletus	9	R	-
COLEOPTERA (BEETLES)	Elmidae	6	XA	VA
	Ptilodactylidae	8	-	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA	XA
	Costachorema	7	A	-
	Hydrobiosis	5	A	A
	Neurochorema	6	R	C
	Beraeoptera	8	R	-
	Confluens	5	R	-
	Oxyethira	2	C	R
	Pycnocentria	7	R	-
	Pycnocentrodes	5	A	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	VA	C
	Maoridiamesa	3	VA	R
	Orthocladiinae	2	A	C
	Tanytarsini	3	A	A
	Empididae	3	R	C
	Ephydridae	4	R	-
	Muscidae	3	R	R
	Austrosimulium	3	C	-
	Tanyderidae	4	-	C
		No of taxa	26	22
		MCI	98	87
		SQMCIs	6.1	4.2
		EPT (taxa)	12	6
		%EPT (taxa)	46	27

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 134 Macroinvertebrate fauna of the Waiwhakaiho River: spring SEM survey sampled on 28 November 2012

Taxa List	Site Code	MCI score	WKH000100	WKH000500	WKH000920	WKH000950
	Sample Number		FWB12460	FWB12461	FWB12462	FWB12464
NEMATODA	Nematoda	3	-	-	-	R
ANNELIDA (WORMS)	Oligochaeta	1	-	R	R	A
	Lumbricidae	5	-	-	-	R
MOLLUSCA	*Potamopyrgus*	4	-	-	-	A
CRUSTACEA	Ostracoda	1	-	-	-	R
	Paracalliope	5	-	-	-	R
	Paratya	3	-	-	-	C
EPHEMEROPTERA (MAYFLIES)	*Acanthophlebia*	9	R	-	-	-
	Austroclima	7	-	R	-	-
	Coloburiscus	7	-	A	C	C
	Deleatidium	8	XA	XA	XA	VA
	Nesameletus	9	C	-	-	R
PLECOPTERA (STONEFLIES)	*Acroperla*	5	-	R	-	-
	Austroperla	9	-	R	-	R
	Megaleptoperla	9	A	R	R	-
	Zelandobius	5	R	R	R	R
	Zelandoperla	8	VA	A	-	-
HEMIPTERA (BUGS)	*Saldula*	5	R	-	-	-
COLEOPTERA (BEETLES)	Elmidae	6	VA	A	A	A
	Hydraenidae	8	-	-	R	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	R	R	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	C	A	VA
	Costachorema	7	A	A	A	R
	Hydrobiosis	5	C	C	C	R
	Hydrochorema	9	R	-	-	-
	Orthopsyche	9	R	-	-	-
	Plectrocnemia	8	-	C	-	-
	Psilochorema	6	C	R	-	-
	Beraeoptera	8	C	-	R	-
	Olinga	9	R	-	-	R
	Oxyethira	2	-	-	-	A
	Pycnocentrodes	5	R	A	-	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	R	A	C	VA
	Eriopterini	5	C	-	-	R
	Maoridiamesa	3	C	A	A	VA
	Orthocladiinae	2	-	C	A	VA
	Polypedilum	3	R	R	-	-
	Tanytarsini	3	-	-	R	C
	Empididae	3	-	-	-	C
	Ephydridae	4	-	-	R	R
	Muscidae	3	-	-	-	R
	Austrosimulium	3	-	R	-	C
		No of taxa	21	22	17	30
		MCI	133	112	108	96
		SQMCIs	7.6	7.4	7.3	4.3
		EPT (taxa)	14	14	8	10
		%EPT (taxa)	67	64	47	33

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 135 Macroinvertebrate fauna of the Waiwhakaiho River: summer SEM survey sampled 1 February 2013

Taxa List	Site Code	MCI score	WKH000100	WKH000500	WKH000920	WKH000950
	Sample Number		FWB13033	FWB13034	FWB13035	FWB13037
NEMERTEA	Nemertea	3	-	-	R	R
NEMATODA	Nematoda	3	-	-	R	C
ANNELIDA (WORMS)	Oligochaeta	1	R	R	A	A
MOLLUSCA	*Potamopyrgus*	4	-	R	C	A
CRUSTACEA	*Paracalliope*	5	-	-	-	R
	Paratya	3	-	-	-	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	-	R	R	R
	Coloburiscus	7	R	C	-	C
	Deleatidium	8	XA	XA	VA	R
	Nesameletus	9	C	R	-	-
	Zephlebia group	7	-	-	-	R
PLECOPTERA (STONEFLIES)	*Austroperla*	9	R	R	-	-
	Megaleptoperla	9	A	-	R	-
	Stenoperla	10	-	R	-	-
	Zelandobius	5	R	-	-	-
	Zelandoperla	8	VA	R	R	R
COLEOPTERA (BEETLES)	Elmidae	6	XA	VA	VA	A
	Hydraenidae	8	R	R	-	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	-	R	C	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C	VA	VA	XA
	Costachorema	7	C	A	R	R
	Hydrobiosis	5	C	A	A	C
	Hydrobiosella	9	R	-	-	-
	Neurochorema	6	-	C	-	-
	Plectrocnemia	8	-	R	-	-
	Psilochorema	6	R	R	-	-
	Olinga	9	C	-	R	-
	Oxyethira	2	-	-	A	A
	Pycnocentria	7	R	-	-	-
	Pycnocentrodes	5	-	R	R	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	A	C	VA
	Eriopterini	5	C	-	-	-
	Maoridiamesa	3	R	VA	R	C
	Orthocladiinae	2	-	A	VA	VA
	Polypedilum	3	R	R	-	R
	Tanypodinae	5	R	-	-	-
	Tanytarsini	3	-	R	A	VA
	Empididae	3	-	R	C	R
	Ephydridae	4	-	R	C	-
	Muscidae	3	-	C	C	R
	Austrosimulium	3	-	-	-	R
	Tanyderidae	4	-	-	R	R
		No of taxa	22	27	24	26
		MCI	125	112	96	89
		SQMCIs	7.1	6.4	4.6	3.7
		EPT (taxa)	14	14	9	8
		%EPT (taxa)	64	52	38	31

'Tolerant' taxa | 'Moderately sensitive' taxa | 'Highly sensitive' taxa

R = Rare   C = Common   A = Abundant   VA = Very Abundant   XA = Extremely Abundant

Table 136 Macroinvertebrate fauna of the Mangorei Stream: spring SEM survey sampled on 28 November 2012

Taxa List	Site Code / Sample Number	MCI score	MGE000970 / FWB12459
ANNELIDA (WORMS)	Oligochaeta	1	C
MOLLUSCA	*Potamopyrgus*	4	C
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A
	Coloburiscus	7	A
	Deleatidium	8	XA
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	A
	Zelandoperla	8	C
COLEOPTERA (BEETLES)	Elmidae	6	A
	Hydraenidae	8	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA
	Costachorema	7	C
	Hydrobiosis	5	A
	Neurochorema	6	C
	Oxyethira	2	R
	Pycnocentrodes	5	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A
	Maoridiamesa	3	A
	Orthocladiinae	2	A
	Tanytarsini	3	C
	Empididae	3	C
	Muscidae	3	R
	Austrosimulium	3	A
		No of taxa	23
		MCI	97
		SQMCIs	6.6
		EPT (taxa)	10
		%EPT (taxa)	43

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 137 Macroinvertebrate fauna of the Mangorei Stream: summer SEM survey sampled on 1 February 2013

Taxa List	Site Code	MCI score	MGE000970
	Sample Number		FWB13032
ANNELIDA (WORMS)	Oligochaeta	1	C
MOLLUSCA	*Potamopyrgus*	4	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	VA
	Coloburiscus	7	A
	Deleatidium	8	A
	Ichthybotus	8	R
	Nesameletus	9	R
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	R
	Zelandoperla	8	A
COLEOPTERA (BEETLES)	Elmidae	6	A
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	XA
	Costachorema	7	C
	Hydrobiosis	5	A
	Neurochorema	6	C
	Oxyethira	2	A
	Pycnocentrodes	5	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	VA
	Harrisius	6	R
	Maoridiamesa	3	C
	Orthocladiinae	2	A
	Tanytarsini	3	A
	Empididae	3	C
	Austrosimulium	3	A
	Tanyderidae	4	R
		No of taxa	25
		MCI	102
		SQMCIs	4.7
		EPT (taxa)	12
		%EPT (taxa)	48

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 138 Macroinvertebrate fauna of the Manganui River: spring SEM survey sampled on 14 November 2012

Taxa List	Site Code	MCI score	MGN000195	MGN000427
	Sample Number		FWB12409	FWB12410
EPHEMEROPTERA (MAYFLIES)	*Ameletopsis*	10	R	-
	Austroclima	7	R	-
	Coloburiscus	7	A	A
	Deleatidium	8	XA	XA
	Nesameletus	9	A	-
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R	R
	Austroperla	9	R	-
	Megaleptoperla	9	C	R
	Stenoperla	10	R	-
	Zelandobius	5	R	R
	Zelandoperla	8	A	R
COLEOPTERA (BEETLES)	Elmidae	6	XA	C
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	-	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C	C
	Costachorema	7	R	C
	Psilochorema	6	C	-
	Beraeoptera	8	R	-
	Olinga	9	C	-
	Oxyethira	2	-	R
	Pycnocentrodes	5	-	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	C
	Eriopterini	5	C	-
	Maoridiamesa	3	-	C
	Orthocladiinae	2	R	C
	Austrosimulium	3	R	-
		No of taxa	21	15
		MCI	135	111
		SQMCIs	7.0	7.7
		EPT (taxa)	16	9
		%EPT (taxa)	76	60
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa		

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 139 Macroinvertebrate fauna of the Manganui River: summer SEM survey sampled on 28 February 2013

Taxa List	Site Code / Sample Number	MCI score	MGN000195	MGN000427
			FWB13129	FWB13130
ANNELIDA (WORMS)	Oligochaeta	1	-	C
MOLLUSCA	*Potamopyrgus*	4	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	R
	Coloburiscus	7	A	C
	Deleatidium	8	XA	VA
	Nesameletus	9	VA	-
PLECOPTERA (STONEFLIES)	*Austroperla*	9	R	-
	Megaleptoperla	9	C	-
	Stenoperla	10	R	-
	Zelandoperla	8	C	R
COLEOPTERA (BEETLES)	Elmidae	6	VA	VA
	Hydraenidae	8	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	XA
	Costachorema	7	-	C
	Hydrobiosis	5	C	A
	Neurochorema	6	-	R
	Psilochorema	6	C	-
	Olinga	9	C	-
	Oxyethira	2	-	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA
	Eriopterini	5	C	-
	Maoridiamesa	3	-	VA
	Orthocladiinae	2	-	VA
	Polypedilum	3	R	-
	Tanytarsini	3	-	C
	Empididae	3	-	R
	Muscidae	3	-	R
		No of taxa	18	19
		MCI	139	96
		SQMCIs	7.6	4.4
		EPT (taxa)	12	8
		%EPT (taxa)	67	42
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa		

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 140 Macroinvertebrate fauna of the Maketawa Stream: SEM spring survey sampled on 14 November 2012

Taxa List	Site Code / Sample Number	MCI score	MKW000200 FWB12411	MKW000300 FWB12412
EPHEMEROPTERA (MAYFLIES)	*Coloburiscus*	7	C	A
	Deleatidium	8	XA	XA
	Nesameletus	9	VA	-
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R	-
	Austroperla	9	R	-
	Megaleptoperla	9	C	R
	Zelandobius	5	-	R
	Zelandoperla	8	VA	C
COLEOPTERA (BEETLES)	Elmidae	6	VA	C
	Hydraenidae	8	R	-
	Ptilodactylidae	8	-	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	C
	Costachorema	7	C	C
	Hydrobiosis	5	C	R
	Hydrobiosella	9	R	-
	Neurochorema	6	-	R
	Plectrocnemia	8	R	-
	Psilochorema	6	C	-
	Beraeoptera	8	C	-
	Confluens	5	-	R
	Olinga	9	R	-
	Pycnocentrodes	5	R	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	C
	Eriopterini	5	C	R
	Maoridiamesa	3	-	A
	Orthocladiinae	2	-	C
	Polypedilum	3	R	-
	Tanypodinae	5	-	R
		No of taxa	21	19
		MCI	139	116
		SQMCIs	7.8	7.6
		EPT (taxa)	15	11
		%EPT (taxa)	71	58

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 141 Macroinvertebrate fauna of the Maketawa Stream: summer SEM survey sampled on 28 February 2013

Taxa List	Site Code / Sample Number	MCI score	MKW000200 / FWB13131	MKW000300 / FWB13132
MOLLUSCA	*Potamopyrgus*	4	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	-	R
	Coloburiscus	7	C	A
	Deleatidium	8	XA	XA
	Nesameletus	9	VA	R
PLECOPTERA (STONEFLIES)	*Austroperla*	9	R	R
	Megaleptoperla	9	C	R
	Stenoperla	10	R	-
	Zelandoperla	8	VA	C
COLEOPTERA (BEETLES)	Elmidae	6	VA	A
	Hydraenidae	8	-	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C	VA
	Costachorema	7	R	C
	Hydrobiosis	5	C	C
	Neurochorema	6	-	R
	Psilochorema	6	C	R
	Olinga	9	R	-
	Pycnocentria	7	-	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA
	Eriopterini	5	C	-
	Maoridiamesa	3	R	A
	Orthocladiinae	2	R	A
	Polypedilum	3	R	-
	Tanytarsini	3	-	C
	Empididae	3	-	R
	Austrosimulium	3	-	R
	Tanyderidae	4	-	R
		No of taxa	19	24
		MCI	128	117
		SQMCIs	7.7	6.7
		EPT (taxa)	12	13
		%EPT (taxa)	63	54

'Tolerant' taxa | 'Moderately sensitive' taxa | 'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 142 Macroinvertebrate fauna of the Waitara River: spring SEM survey sampled on 5 October 2012

Taxa List	Site Code	MCI score	WTR000850
	Sample Number		FWB12344
ANNELIDA (WORMS)	Oligochaeta	1	C
MOLLUSCA	*Melanopsis*	3	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R
	Coloburiscus	7	R
	Deleatidium	8	VA
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R
	Zelandobius	5	C
COLEOPTERA (BEETLES)	Elmidae	6	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C
	Costachorema	7	C
	Hydrobiosis	5	R
	Neurochorema	6	R
	Beraeoptera	8	R
	Pycnocentria	7	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A
	Maoridiamesa	3	C
	Orthocladiinae	2	A
		No of taxa	18
		MCI	107
		SQMCIs	6.3
		EPT (taxa)	11
		%EPT (taxa)	61
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa	

R = Rare   C = Common   A = Abundant   VA = Very Abundant   XA = Extremely Abundant

Table 143 Macroinvertebrate fauna of the Waitara River: summer SEM survey sampled on 28 February 2013

Taxa List	Site Code	MCI score	WTR000850
	Sample Number		FWB13133
ANNELIDA (WORMS)	Oligochaeta	1	VA
MOLLUSCA	*Potamopyrgus*	4	C
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R
COLEOPTERA (BEETLES)	Elmidae	6	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA
	Hydrobiosis	5	C
	Oxyethira	2	C
	Pycnocentrodes	5	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A
	Orthocladiinae	2	VA
	Tanytarsini	3	A
	Empididae	3	C
	Tanyderidae	4	R
		No of taxa	14
		MCI	83
		SQMCIs	2.6
		EPT (taxa)	4
		%EPT (taxa)	29
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 144 Macroinvertebrate fauna of the Mangati Stream: spring SEM survey sampled on 3 October 2012

Taxa List	Site Code	MCI score	MGT000488	MGT000520
	Sample Number		FWB12336	FWB12342
NEMATODA	Nematoda	3	R	-
ANNELIDA (WORMS)	Oligochaeta	1	C	VA
	Lumbricidae	5	R	R
MOLLUSCA	*Physa*	3	R	R
	Potamopyrgus	4	XA	XA
CRUSTACEA	*Paracalliope*	5	VA	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	R
	Zephlebia group	7	R	-
COLEOPTERA (BEETLES)	Dytiscidae	5	R	-
TRICHOPTERA (CADDISFLIES)	*Hydrobiosis*	5	C	R
	Psilochorema	6	R	-
	Oxyethira	2	-	R
	Triplectides	5	R	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	-	C
	Zelandotipula	6	R	-
	Orthocladiinae	2	A	A
	Polypedilum	3	R	C
	Empididae	3	R	C
	Austrosimulium	3	VA	-
		No of taxa	17	12
		MCI	86	75
		SQMCIs	4.0	3.5
		EPT (taxa)	5	2
		%EPT (taxa)	29	17

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 145 Macroinvertebrate fauna of the Mangati Stream: summer SEM survey sampled on 12 February 2013

Taxa List	Site Code	MCI score	MGT000488	MGT000520
	Sample Number		FWB13059	FWB13065
NEMERTEA	Nemertea	3	C	R
ANNELIDA (WORMS)	Oligochaeta	1	A	VA
	Lumbricidae	5	R	R
MOLLUSCA	*Potamopyrgus*	4	XA	XA
CRUSTACEA	Ostracoda	1	-	C
	Isopoda	5	C	R
	Paracalliope	5	XA	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	R
	Zephlebia group	7	C	-
COLEOPTERA (BEETLES)	Dytiscidae	5	R	-
TRICHOPTERA (CADDISFLIES)	*Polyplectropus*	6	R	-
	Psilochorema	6	C	-
	Oxyethira	2	-	C
	Triplectides	5	R	A
DIPTERA (TRUE FLIES)	Eriopterini	5	C	R
	Hexatomini	5	R	-
	Zelandotipula	6	-	R
	Orthocladiinae	2	R	C
	Polypedilum	3	A	-
	Empididae	3	-	R
	Sciomyzidae	3	-	R
	Austrosimulium	3	R	C
	Tanyderidae	4	-	R
ACARINA (MITES)	Acarina	5	R	-
		No of taxa	18	17
		MCI	91	75
		SQMCIs	4.4	3.5
		EPT (taxa)	5	2
		%EPT (taxa)	28	12
'Tolerant' taxa	'Moderately sensitive' taxa		'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 146 Macroinvertebrate fauna of the Waimoku Stream: spring SEM survey sampled on 5 October 2012

Taxa List	Site Code	MCI score	WMK000100	WMK000298
	Sample Number		FWB12371	FWB12372
PLATYHELMINTHES (FLATWORMS)	*Neppia*	6	R	-
ANNELIDA (WORMS)	Oligochaeta	1	R	-
	Lumbricidae	5	R	-
MOLLUSCA	*Gyraulus*	3	R	-
	Potamopyrgus	4	C	C
CRUSTACEA	Paraleptamphopidae	5	R	-
	Talitridae	5	C	-
	Paratya	3	-	R
	Paranephrops	5	R	-
EPHEMEROPTERA (MAYFLIES)	*Ameletopsis*	10	C	-
	Austroclima	7	C	C
	Coloburiscus	7	VA	A
	Deleatidium	8	A	C
	Ichthybotus	8	R	-
	Nesameletus	9	C	-
	Zephlebia group	7	VA	A
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R	R
	Austroperla	9	A	-
	Megaleptoperla	9	R	-
	Spaniocerca	8	R	-
	Stenoperla	10	C	-
	Zelandobius	5	C	C
	Zelandoperla	8	C	-
COLEOPTERA (BEETLES)	Elmidae	6	C	-
	Hydraenidae	8	R	-
	Ptilodactylidae	8	C	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	-
TRICHOPTERA (CADDISFLIES)	*Costachorema*	7	R	-
	Hydrobiosis	5	R	R
	Hydrobiosella	9	C	-
	Hydrochorema	9	R	-
	Orthopsyche	9	VA	-
	Helicopsyche	10	R	-
	Pycnocentria	7	R	-
DIPTERA (TRUE FLIES)	Hexatomini	5	R	-
	Maoridiamesa	3	-	R
	Orthocladiinae	2	R	C
	Polypedilum	3	R	R
	Nothodixa	4	R	-
	Empididae	3	R	-
	Austrosimulium	3	-	C
		No of taxa	38	13
		MCI	129	95
		SQMCIs	7.6	5.9
		EPT (taxa)	21	7
		%EPT (taxa)	55	54
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa		

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 147 Macroinvertebrate fauna of the Waimoku Stream: summer SEM survey sampled on 26 February 2013

Taxa List	Site Code	MCI score	WMK000100	WMK000298
	Sample Number		FWB13112	FWB13113
PLATYHELMINTHES (FLATWORMS)	*Neppia*	6	R	-
NEMERTEA	Nemertea	3	-	C
ANNELIDA (WORMS)	Oligochaeta	1	R	A
	Lumbricidae	5	-	R
MOLLUSCA	*Gyraulus*	3	R	-
	Potamopyrgus	4	C	XA
CRUSTACEA	Ostracoda	1	-	R
	Talitridae	5	C	-
	Paratya	3	-	R
	Paranephrops	5	R	-
EPHEMEROPTERA (MAYFLIES)	*Ameletopsis*	10	R	-
	Austroclima	7	C	C
	Coloburiscus	7	VA	C
	Deleatidium	8	R	-
	Ichthybotus	8	C	-
	Nesameletus	9	C	-
	Zephlebia group	7	VA	C
PLECOPTERA (STONEFLIES)	*Acroperla*	5	-	R
	Austroperla	9	A	-
	Megaleptoperla	9	R	-
	Stenoperla	10	C	-
	Zelandobius	5	R	-
	Zelandoperla	8	C	-
COLEOPTERA (BEETLES)	Elmidae	6	A	C
	Hydraenidae	8	C	-
	Hydrophilidae	5	R	-
	Ptilodactylidae	8	R	-
	Staphylinidae	5	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	C
	Costachorema	7	R	R
	Hydrobiosis	5	R	A
	Hydrobiosella	9	C	-
	Hydrochorema	9	R	-
	Orthopsyche	9	VA	C
	Oxyethira	2	-	C
	Pycnocentrodes	5	-	C
	Triplectides	5	R	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	R	A
	Hexatomini	5	R	-
	Harrisius	6	R	-
	Maoridiamesa	3	R	C
	Orthocladiinae	2	C	A
	Polypedilum	3	C	C
	Tanytarsini	3	-	C
	Empididae	3	R	A
	Ephydridae	4	-	R
	Austrosimulium	3	-	R
	Tanyderidae	4	-	C
ACARINA (MITES)	Acarina	5	R	-
		No of taxa	38	27
		MCI	124	87
		SQMCIs	7.4	4.0
		EPT (taxa)	18	9
		%EPT (taxa)	47	33

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 148 Macroinvertebrate fauna of the Waiau Stream: spring SEM survey sampled on 5 October 2012

Taxa List	Site Code / Sample Number	MCI score	WAI000110 / FWB12361
ANNELIDA (WORMS)	Oligochaeta	1	C
	Lumbricidae	5	R
MOLLUSCA	*Latia*	5	C
	Potamopyrgus	4	C
CRUSTACEA	*Paracalliope*	5	R
	Paraleptamphopidae	5	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	XA
	Zephlebia group	7	R
PLECOPTERA (STONEFLIES)	*Acroperla*	5	C
	Zelandobius	5	R
COLEOPTERA (BEETLES)	Elmidae	6	VA
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA
	Costachorema	7	R
	Hydrobiosis	5	A
	Pycnocentria	7	VA
	Pycnocentrodes	5	VA
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A
	Maoridiamesa	3	R
	Orthocladiinae	2	R
		No of taxa	20
		MCI	100
		SQMCIs	6.2
		EPT (taxa)	9
		%EPT (taxa)	45

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare   C = Common   A = Abundant   VA = Very Abundant   XA = Extremely Abundant

Table 149 Macroinvertebrate fauna of the Waiau Stream: summer SEM survey sampled on 12 February 2013

Taxa List	Site Code	MCI score	WAI000110
	Sample Number		FWB13056
COELENTERATA	Coelenterata	3	R
PLATYHELMINTHES (FLATWORMS)	*Cura*	3	R
NEMERTEA	Nemertea	3	A
ANNELIDA (WORMS)	Oligochaeta	1	VA
	Lumbricidae	5	R
MOLLUSCA	*Latia*	5	A
	Physa	3	R
	Potamopyrgus	4	VA
CRUSTACEA	Ostracoda	1	R
	Paracalliope	5	C
	Talitridae	5	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A
	Coloburiscus	7	C
	Zephlebia group	7	C
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	R
COLEOPTERA (BEETLES)	Elmidae	6	XA
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA
	Costachorema	7	R
	Hydrobiosis	5	C
	Neurochorema	6	R
	Hudsonema	6	A
	Oxyethira	2	C
	Pycnocentria	7	A
	Pycnocentrodes	5	A
DIPTERA (TRUE FLIES)	*Aphrophila*	5	R
	Maoridiamesa	3	R
	Orthocladiinae	2	VA
	Tanytarsini	3	C
	Austrosimulium	3	C
		No of taxa	30
		MCI	90
		SQMCIs	4.7
		EPT (taxa)	11
		%EPT (taxa)	37
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 150 Macroinvertebrate fauna of the Punehu Stream: spring SEM survey sampled on 4 October 2012

Taxa List	Site Code	MCI score	PNH000200	PNH000900
	Sample Number		FWB12355	FWB12356
ANNELIDA (WORMS)	Oligochaeta	1	-	C
MOLLUSCA	*Potamopyrgus*	4	-	C
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	R
	Coloburiscus	7	A	A
	Deleatidium	8	VA	XA
	Nesameletus	9	VA	C
PLECOPTERA (STONEFLIES)	*Acroperla*	5	-	R
	Austroperla	9	R	-
	Megaleptoperla	9	C	-
	Stenoperla	10	R	-
	Zelandobius	5	-	C
	Zelandoperla	8	VA	-
COLEOPTERA (BEETLES)	Elmidae	6	A	A
	Hydraenidae	8	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C	A
	Costachorema	7	C	R
	Hydrobiosis	5	R	C
	Psilochorema	6	C	-
	Beraeoptera	8	VA	C
	Helicopsyche	10	R	-
	Olinga	9	R	-
	Pycnocentria	7	R	-
	Pycnocentrodes	5	A	VA
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	R
	Eriopterini	5	C	-
	Maoridiamesa	3	R	R
	Orthocladiinae	2	R	C
	Polypedilum	3	R	-
	Tanytarsini	3	-	R
		No of taxa	25	19
		MCI	134	106
		SQMCIs	7.8	7.2
		EPT (taxa)	17	11
		%EPT (taxa)	68	58
'Tolerant' taxa	'Moderately sensitive' taxa		'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 151 Macroinvertebrate fauna of the Punehu Stream: summer SEM survey sampled on 10 February 2013

Taxa List	Site Code	MCI score	PNH000200	PNH000900
	Sample Number		FWB13074	FWB13075
PLATYHELMINTHES (FLATWORMS)	*Cura*	3	-	R
NEMERTEA	Nemertea	3	-	C
ANNELIDA (WORMS)	Oligochaeta	1	R	C
MOLLUSCA	*Potamopyrgus*	4	R	A
CRUSTACEA	Ostracoda	1	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	C
	Coloburiscus	7	A	C
	Deleatidium	8	XA	VA
	Nesameletus	9	VA	-
PLECOPTERA (STONEFLIES)	*Austroperla*	9	R	-
	Megaleptoperla	9	C	-
	Stenoperla	10	R	-
	Zelandoperla	8	A	-
COLEOPTERA (BEETLES)	Elmidae	6	VA	A
	Hydraenidae	8	C	-
	Staphylinidae	5	-	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA	VA
	Costachorema	7	C	C
	Hydrobiosis	5	C	-
	Psilochorema	6	R	-
	Beraeoptera	8	A	-
	Olinga	9	R	-
	Oxyethira	2	R	R
	Pycnocentrodes	5	C	A
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	C
	Eriopterini	5	R	-
	Maoridiamesa	3	A	C
	Orthocladiinae	2	R	A
	Polypedilum	3	R	R
	Tanytarsini	3	-	C
	Austrosimulium	3	-	R
	Tabanidae	3	R	-
	Tanyderidae	4	-	C
		No of taxa	27	22
		MCI	119	85
		SQMCIs	7.2	5.4
		EPT (taxa)	15	6
		%EPT (taxa)	56	27

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare   C = Common   A = Abundant   VA = Very Abundant   XA = Extremely Abundant

Table 152 Macroinvertebrate fauna of the Patea River: spring SEM survey sampled on 1 October 2012

Taxa List	Site Code	MCI score	PAT000200	PAT000315	PAT000360
	Sample Number		FWB12316	FWB12317	FWB12320
ANNELIDA (WORMS)	Oligochaeta	1	-	-	C
MOLLUSCA	*Potamopyrgus*	4	-	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R	C	-
	Coloburiscus	7	A	VA	A
	Deleatidium	8	XA	XA	XA
	Neozephlebia	7	R	-	-
	Nesameletus	9	R	A	-
PLECOPTERA (STONEFLIES)	*Acroperla*	5	C	C	-
	Megaleptoperla	9	C	R	-
	Spaniocerca	8	R	-	-
	Stenoperla	10	-	R	-
	Zelandobius	5	C	R	C
	Zelandoperla	8	A	C	R
COLEOPTERA (BEETLES)	Elmidae	6	C	C	A
	Hydraenidae	8	R	-	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	-	C	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	C	A
	Costachorema	7	VA	C	C
	Hydrobiosis	5	R	R	R
	Hydrobiosella	9	R	-	-
	Orthopsyche	9	C	-	-
	Psilochorema	6	R	-	-
	Beraeoptera	8	A	C	-
	Confluens	5	R	R	R
	Helicopsyche	10	C	-	-
	Pycnocentrodes	5	-	C	C
	Zelolessica	7	C	-	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	A	A
	Eriopterini	5	-	R	R
	Limonia	6	R	-	-
	Maoridiamesa	3	-	C	C
	Orthocladiinae	2	C	C	A
	Tanytarsini	3	-	R	C
	Empididae	3	-	R	-
		No of taxa	24	23	18
		MCI	138	118	100
		SQMCIs	7.7	7.6	7.3
		EPT (taxa)	19	15	9
		%EPT (taxa)	79	65	50
'Tolerant' taxa		'Moderately sensitive' taxa		'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 153 Macroinvertebrate fauna of the Patea River: summer SEM survey sampled on 28 February 2013

Taxa List	Site Code	MCI score	PAT000200	PAT000315	PAT000360
	Sample Number		FWB13119	FWB13120	FWB13125
NEMERTEA	Nemertea	3	-	-	R
ANNELIDA (WORMS)	Oligochaeta	1	-	C	C
MOLLUSCA	*Potamopyrgus*	4	-	-	C
CRUSTACEA	Isopoda	5	R	-	-
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	-	R
	Coloburiscus	7	VA	VA	A
	Deleatidium	8	VA	XA	VA
	Nesameletus	9	A	A	-
	Zephlebia group	7	R	R	-
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R	-	-
	Austroperla	9	R	-	-
	Megaleptoperla	9	A	R	-
	Stenoperla	10	R	-	-
	Taraperla	10	R	-	-
	Zelandobius	5	R	-	-
	Zelandoperla	8	A	C	-
HEMIPTERA (BUGS)	*Saldula*	5	-	R	-
COLEOPTERA (BEETLES)	Elmidae	6	A	VA	VA
	Hydraenidae	8	A	A	C
	Staphylinidae	5	R	-	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	A	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	VA	XA
	Costachorema	7	C	C	C
	Hydrobiosis	5	C	C	A
	Hydrobiosella	9	R	-	-
	Neurochorema	6	-	C	R
	Orthopsyche	9	A	-	-
	Psilochorema	6	R	-	-
	Alloecentrella	8	R	-	-
	Beraeoptera	8	C	C	-
	Confluens	5	C	-	R
	Helicopsyche	10	C	-	-
	Olinga	9	R	R	-
	Oxyethira	2	-	R	R
	Pycnocentria	7	R	-	-
	Pycnocentrodes	5	-	C	R
	Zelolessica	7	R	-	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA	VA
	Eriopterini	5	-	R	-
	Maoridiamesa	3	-	R	VA
	Orthocladiinae	2	-	A	A
	Polypedilum	3	C	C	-
	Tanytarsini	3	-	C	VA
	Empididae	3	-	-	C
	Ephydridae	4	-	C	-
	Muscidae	3	R	-	C
	Psychodidae	1	-	R	-
	Austrosimulium	3	-	C	R
	Tanyderidae	4	-	R	-
ACARINA (MITES)	Acarina	5	-	R	-
		No of taxa	32	30	25
		MCI	141	106	95
		SQMCIs	7.5	6.7	4.6
		EPT (taxa)	24	13	10
		%EPT (taxa)	75	43	40

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 154 Macroinvertebrate fauna of the Mangaehu River: spring SEM survey sampled on 1 October 2012

Taxa List	Site Code / Sample Number	MCI score	MGH000950 / FWB12323
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C
	Coloburiscus	7	C
	Deleatidium	8	VA
	Zephlebia group	7	A
PLECOPTERA (STONEFLIES)	*Acroperla*	5	C
	Zelandobius	5	C
	Zelandoperla	8	R
COLEOPTERA (BEETLES)	Elmidae	6	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	R
	Costachorema	7	A
	Hydrobiosis	5	R
	Neurochorema	6	R
	Oxyethira	2	R
	Pycnocentrodes	5	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A
	Maoridiamesa	3	VA
	Orthocladiinae	2	VA
	Tanytarsini	3	C
	Empididae	3	R
	Austrosimulium	3	R
		No of taxa	20
		MCI	101
		SQMCIs	4.7
		EPT (taxa)	12
		%EPT (taxa)	60

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 155 Macroinvertebrate fauna of the Mangaehu River: summer SEM survey sampled on 28 February 2013

Taxa List	Site Code	MCI score	MGH000950
	Sample Number		FWB13128
NEMERTEA	Nemertea	3	R
ANNELIDA (WORMS)	Oligochaeta	1	R
MOLLUSCA	*Potamopyrgus*	4	C
CRUSTACEA	*Paracalliope*	5	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	VA
	Coloburiscus	7	A
	Deleatidium	8	R
	Zephlebia group	7	C
COLEOPTERA (BEETLES)	Elmidae	6	C
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA
	Hydrobiosis	5	VA
	Neurochorema	6	R
	Oxyethira	2	R
	Pycnocentria	7	R
	Pycnocentrodes	5	A
DIPTERA (TRUE FLIES)	*Aphrophila*	5	VA
	Maoridiamesa	3	A
	Orthocladiinae	2	VA
	Polypedilum	3	C
	Tanytarsini	3	VA
	Empididae	3	C
	Muscidae	3	C
	Austrosimulium	3	R
		No of taxa	24
		MCI	91
		SQMCIs	4.4
		EPT (taxa)	9
		%EPT (taxa)	38
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 156 Macroinvertebrate fauna of the Waingongoro River: spring SEM survey sampled on 31 October 2012

Taxa List	Site Code	MCI score	WGG000115	WGG000150	WGG000500	WGG000540	WGG000895	WGG000995
	Sample Number		FWB12392	FWB12393	FWB12394	FWB12395	FWB12398	FWB12399
NEMATOMORPHA	Nematomorpha	3	-	-	-	R	-	-
ANNELIDA (WORMS)	Oligochaeta	1	-	-	-	A	VA	C
	Lumbricidae	5	-	-	-	-	R	R
MOLLUSCA	*Latia*	5	-	-	-	-	-	R
	Potamopyrgus	4	-	-	-	C	C	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	A	R	-	C	C
	Coloburiscus	7	XA	XA	VA	VA	C	R
	Deleatidium	8	XA	XA	XA	XA	XA	VA
	Nesameletus	9	C	A	C	A	-	-
PLECOPTERA (STONEFLIES)	*Acroperla*	5	C	-	R	-	-	-
	Austroperla	9	A	-	-	-	-	-
	Megaleptoperla	9	A	C	-	-	-	-
	Spaniocerca	8	R	-	-	-	-	-
	Stenoperla	10	C	-	-	-	-	-
	Zelandobius	5	C	-	A	A	R	R
	Zelandoperla	8	VA	XA	C	R	-	-
COLEOPTERA (BEETLES)	Elmidae	6	A	A	A	VA	C	R
	Hydraenidae	8	A	R	R	R	-	-
	Ptilodactylidae	8	-	R	-	-	-	-
	Staphylinidae	5	R	-	-	-	-	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	A	C	C	C	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	A	C	A	A	VA
	Costachorema	7	C	R	C	C	R	C
	Hydrobiosis	5	R	R	C	C	C	-
	Hydrobiosella	9	R	-	-	-	-	-
	Neurochorema	6	-	-	R	R	-	R
	Orthopsyche	9	R	R	-	-	-	-
	Psilochorema	6	-	-	R	-	-	-
	Beraeoptera	8	A	VA	C	A	-	-
	Confluens	5	R	C	-	-	-	-
	Helicopsyche	10	A	C	-	-	-	-
	Olinga	9	C	C	-	R	-	-
	Pycnocentria	7	R	-	-	-	R	-
	Pycnocentrodes	5	C	C	A	VA	VA	XA
	Zelolessica	7	C	-	-	-	-	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	VA	A	C	C	-	A
	Eriopterini	5	-	C	-	R	-	-
	Harrisius	6	-	R	-	-	-	-
	Maoridiamesa	3	R	R	-	-	-	VA
	Orthocladiinae	2	C	-	R	R	-	XA
	Polypedilum	3	C	-	-	R	-	-
	Tanytarsini	3	C	-	-	-	-	-
	Empididae	3	R	C	-	-	-	-
	Austrosimulium	3	-	-	-	-	R	R
	Tanyderidae	4	-	-	-	-	R	-
		No of taxa	33	24	19	22	16	18
		MCI	130	134	124	114	106	100
		SQMCIs	7.3	7.6	7.5	7.0	6.5	3.9
		EPT (taxa)	23	15	14	12	9	8
		%EPT (taxa)	70	63	74	55	56	44

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 157 Macroinvertebrate fauna of the Waingongoro River: summer SEM survey sampled on 25 February 2013

Taxa List	Site Code	MCI score	WGG000115	WGG000150	WGG000500	WGG000665	WGG000895	WGG000995
	Sample Number		FWB13098	FWB13099	FWB13100	FWB13105	FWB13106	FWB13107
PLATYHELMINTHES (FLATWORMS)	*Cura*	3	-	-	-	-	R	-
	Neppia	6	R	-	-	-	-	-
NEMERTEA	Nemertea	3	-	-	-	-	C	R
ANNELIDA (WORMS)	Oligochaeta	1	R	-	C	C	C	VA
	Lumbricidae	5	-	-	-	-	R	R
MOLLUSCA	*Potamopyrgus*	4	R	R	C	C	VA	VA
CRUSTACEA	*Paracalliope*	5	-	-	-	R	-	-
	Talitridae	5	-	-	-	-	R	R
EPHEMEROPTERA (MAYFLIES)	*Acanthophlebia*	9	-	R	-	-	-	-
	Ameletopsis	10	R	-	-	-	-	-
	Austroclima	7	C	VA	A	A	VA	C
	Coloburiscus	7	XA	XA	VA	A	C	R
	Deleatidium	8	XA	XA	XA	XA	VA	C
	Nesameletus	9	VA	XA	C	-	-	-
	Zephlebia group	7	-	-	R	R	R	R
PLECOPTERA (STONEFLIES)	*Austroperla*	9	R	R	-	-	-	-
	Megaleptoperla	9	A	C	-	-	-	-
	Stenoperla	10	C	-	-	-	-	-
	Taraperla	10	R	-	-	-	-	-
	Zelandobius	5	R	-	-	-	-	-
	Zelandoperla	8	XA	VA	R	R	-	-
HEMIPTERA (BUGS)	*Saldula*	5	-	-	-	-	-	R
COLEOPTERA (BEETLES)	Elmidae	6	VA	VA	VA	VA	R	-
	Hydraenidae	8	VA	A	A	R	-	-
	Staphylinidae	5	R	-	-	-	-	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	VA	A	A	A	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	VA	XA	VA	VA	XA
	Costachorema	7	C	R	R	C	R	-
	Hydrobiosis	5	R	A	A	A	C	R
	Neurochorema	6	-	-	R	-	-	-
	Orthopsyche	9	R	-	-	-	-	-
	Psilochorema	6	R	-	-	-	-	-
	Beraeoptera	8	A	VA	R	R	-	-
	Confluens	5	R	C	-	-	-	-
	Helicopsyche	10	A	C	-	-	-	-
	Olinga	9	A	A	-	-	-	-
	Pycnocentria	7	C	-	R	R	C	-
	Pycnocentrodes	5	-	R	R	C	VA	VA
	Zelolessica	7	R	-	-	-	-	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	VA	VA	A	A	C	A
	Eriopterini	5	R	C	R	C	-	-
	Harrisius	6	-	-	R	-	-	R
	Maoridiamesa	3	-	-	R	A	R	C
	Orthocladiinae	2	C	R	C	A	C	VA
	Polypedilum	3	C	R	-	R	C	R
	Tanytarsini	3	-	-	C	A	R	A
	Dolichopodidae	3	R	-	-	-	-	-
	Empididae	3	-	R	-	-	-	-
	Ephydridae	4	-	-	-	-	-	C
	Austrosimulium	3	-	-	C	R	C	R
	Tabanidae	3	-	-	-	R	-	-
	Tanyderidae	4	-	-	R	-	-	-
		No of taxa	34	25	26	25	23	22
		MCI	132	130	112	105	96	93
		SQMCIs	7.5	7.5	6.1	6.6	5.5	3.6
		EPT (taxa)	22	16	13	11	9	7
		%EPT (taxa)	65	64	50	44	39	32

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 158 Macroinvertebrate fauna of the Mangawhero Stream: spring SEM survey sampled on 31 October 2012

Taxa List	Site Code / Sample Number	MCI score	MWH000380 / FWB12400	MWH000490 / FWB12401
ANNELIDA (WORMS)	Oligochaeta	1	VA	VA
	Lumbricidae	5	-	R
MOLLUSCA	*Potamopyrgus*	4	C	A
CRUSTACEA	*Paracalliope*	5	C	R
	Talitridae	5	R	C
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	VA	C
	Coloburiscus	7	-	R
	Deleatidium	8	-	XA
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	-	A
	Zelandoperla	8	-	R
COLEOPTERA (BEETLES)	Elmidae	6	-	A
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	-	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	VA
	Costachorema	7	R	C
	Hydrobiosis	5	C	C
	Pycnocentria	7	-	C
	Pycnocentrodes	5	-	XA
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA
	Maoridiamesa	3	C	C
	Orthocladiinae	2	A	A
	Muscidae	3	-	R
	Austrosimulium	3	R	R
		No of taxa	12	22
		MCI	85	102
		SQMCIs	4.0	5.7
		EPT (taxa)	4	10
		%EPT (taxa)	33	45

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 159 Macroinvertebrate fauna of the Mangawhero Stream: summer SEM survey sampled on 25 February 2013

Taxa List	Site Code / Sample Number	MCI score	MWH000380 FWB13108	MWH000490 FWB13109
NEMERTEA	Nemertea	3	R	C
ANNELIDA (WORMS)	Oligochaeta	1	R	A
	Lumbricidae	5	-	R
MOLLUSCA	*Potamopyrgus*	4	C	R
CRUSTACEA	Ostracoda	1	VA	R
	Paracalliope	5	VA	C
	Talitridae	5	-	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	A
	Coloburiscus	7	-	C
	Deleatidium	8	-	XA
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	-	R
COLEOPTERA (BEETLES)	Elmidae	6	R	XA
	Dytiscidae	5	R	-
	Hydrophilidae	5	C	-
	Staphylinidae	5	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	VA
	Costachorema	7	-	R
	Hydrobiosis	5	R	A
	Neurochorema	6	-	R
	Polyplectropus	6	R	-
	Oxyethira	2	R	R
	Pycnocentria	7	-	A
	Pycnocentrodes	5	-	A
	Triplectides	5	R	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	C
	Eriopterini	5	-	R
	Limonia	6	R	-
	Chironomus	1	R	-
	Maoridiamesa	3	-	C
	Orthocladiinae	2	A	A
	Tanytarsini	3	-	A
	Dolichopodidae	3	-	R
	Paradixa	4	R	-
	Ephydridae	4	-	R
	Muscidae	3	-	C
	Austrosimulium	3	A	C
	Stratiomyidae	5	R	-
ACARINA (MITES)	Acarina	5	R	-
		No of taxa	24	29
		MCI	85	90
		SQMCIs	3.5	6.4
		EPT (taxa)	5	10
		%EPT (taxa)	21	34

'Tolerant' taxa | 'Moderately sensitive' taxa | 'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 160 Macroinvertebrate fauna of the Huatoki Stream: spring SEM survey sampled on 16 November 2012

Taxa List	Site Code	MCI score	HTK000350	HTK000425	HTK000745
	Sample Number		FWB12421	FWB12422	FWB12423
ANNELIDA (WORMS)	Oligochaeta	1	R	C	A
	Lumbricidae	5	R	-	-
MOLLUSCA	*Latia*	5	R	R	-
	Potamopyrgus	4	C	A	VA
	Sphaeriidae	3	-	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	A	-
	Coloburiscus	7	VA	VA	-
	Deleatidium	8	XA	VA	R
	Ichthybotus	8	-	R	-
	Nesameletus	9	VA	C	-
	Zephlebia group	7	VA	A	R
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	A	A	C
	Zelandoperla	8	C	C	-
COLEOPTERA (BEETLES)	Elmidae	6	VA	VA	VA
	Ptilodactylidae	8	R	R	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA	A	R
	Costachorema	7	C	C	-
	Hydrobiosis	5	R	R	-
	Neurochorema	6	-	-	R
	Polyplectropus	6	-	-	R
	Psilochorema	6	-	R	-
	Beraeoptera	8	R	-	-
	Pycnocentria	7	-	R	-
	Pycnocentrodes	5	R	C	C
	Triplectides	5	-	-	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	C	-
	Eriopterini	5	-	-	R
	Orthocladiinae	2	R	R	R
	Austrosimulium	3	R	C	-
	Tanyderidae	4	-	C	C
ACARINA (MITES)	Acarina	5	R	-	-
		No of taxa	23	24	17
		MCI	114	115	101
		SQMCIs	7.2	6.5	4.7
		EPT (taxa)	12	14	8
		%EPT (taxa)	52	58	47

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 161 Macroinvertebrate fauna of the Huatoki Stream: summer SEM survey sampled on 4 March 2013

Taxa List	Site Code	MCI score	HTK000350	HTK000425	HTK000745
	Sample Number		FWB13134	FWB13135	FWB13136
NEMERTEA	Nemertea	3	R	C	C
ANNELIDA (WORMS)	Oligochaeta	1	C	A	VA
MOLLUSCA	*Latia*	5	C	C	R
	Potamopyrgus	4	A	A	XA
	Sphaeriidae	3	-	-	R
CRUSTACEA	Ostracoda	1	-	-	R
	Paratya	3	-	R	VA
	Paranephrops	5	R	-	-
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	VA	-
	Coloburiscus	7	XA	XA	R
	Deleatidium	8	XA	A	R
	Ichthybotus	8	R	-	-
	Nesameletus	9	VA	A	-
	Zephlebia group	7	A	A	R
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R	-	-
	Zelandobius	5	R	R	-
	Zelandoperla	8	R	R	-
COLEOPTERA (BEETLES)	Elmidae	6	VA	XA	XA
	Ptilodactylidae	8	R	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA	VA	-
	Costachorema	7	C	R	-
	Hydrobiosis	5	A	C	-
	Neurochorema	6	-	R	-
	Psilochorema	6	R	-	-
	Confluens	5	-	R	-
	Oeconesidae	5	-	-	R
	Oxyethira	2	R	-	R
	Pycnocentria	7	-	C	R
	Pycnocentrodes	5	-	C	R
	Triplectides	5	R	C	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	C	-
	Eriopterini	5	-	-	R
	Harrisius	6	R	-	R
	Maoridiamesa	3	R	-	-
	Orthocladiinae	2	C	C	C
	Polypedilum	3	R	C	-
	Tanypodinae	5	R	-	-
	Tanytarsini	3	C	-	-
	Culicidae	3	-	-	R
	Empididae	3	R	R	R
	Muscidae	3	R	-	-
	Austrosimulium	3	R	A	R
	Tanyderidae	4	R	C	C
ACARINA (MITES)	Acarina	5	-	-	R
		No of taxa	34	29	25
		MCI	101	105	88
		SQMCIs	7.0	6.2	4.5
		EPT (taxa)	14	15	7
		%EPT (taxa)	41	52	28

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 162 Macroinvertebrate fauna of the Kaupokonui River: spring SEM survey sampled on 4 October 2012

Taxa List	Site Code	MCI score	KPK000250	KPK000500	KPK000660	KPK000880	KPK000990
	Sample Number		FWB12352	FWB12353	FWB12346	FWB12349	FWB12354
NEMATODA	Nematoda	3	-	-	-	R	R
ANNELIDA (WORMS)	Oligochaeta	1	R	-	-	R	C
MOLLUSCA	*Potamopyrgus*	4	-	-	C	R	C
EPHEMEROPTERA (MAYFLIES)	*Ameletopsis*	10	-	-	-	R	-
	Austroclima	7	-	C	C	R	R
	Coloburiscus	7	VA	XA	VA	A	A
	Deleatidium	8	XA	XA	XA	XA	XA
	Nesameletus	9	C	VA	A	R	-
PLECOPTERA (STONEFLIES)	*Acroperla*	5	-	-	R	R	-
	Austroperla	9	R	-	R	-	-
	Megaleptoperla	9	C	R	-	-	-
	Stenoperla	10	C	-	-	-	-
	Zelandobius	5	C	R	R	R	R
	Zelandoperla	8	VA	A	R	-	-
COLEOPTERA (BEETLES)	Elmidae	6	VA	A	VA	C	R
	Hydraenidae	8	-	R	C	-	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	-	A	A	A	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	A	XA	A	VA
	Costachorema	7	R	-	R	C	C
	Hydrobiosis	5	R	C	C	C	R
	Hydrobiosella	9	C	-	-	-	-
	Neurochorema	6	-	-	R	-	-
	Plectrocnemia	8	-	-	R	-	-
	Psilochorema	6	R	R	R	-	-
	Beraeoptera	8	A	VA	VA	R	R
	Helicopsyche	10	R	R	R	-	-
	Olinga	9	A	A	A	-	-
	Pycnocentria	7	R	-	-	-	R
	Pycnocentrodes	5	-	A	VA	A	VA
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA	VA	R	A
	Eriopterini	5	C	A	C	-	R
	Harrisius	6	-	-	-	-	R
	Maoridiamesa	3	-	R	-	A	VA
	Orthocladiinae	2	R	C	R	A	A
	Tanypodinae	5	-	R	-	-	-
	Dolichopodidae	3	-	-	R	-	-
	Empididae	3	R	-	-	-	-
	Psychodidae	1	R	-	-	-	-
	Tabanidae	3	-	-	-	-	R
ACARINA (MITES)	Acarina	5	-	-	-	R	-
		No of taxa	24	22	26	21	21
		MCI	128	128	128	110	103
		SQMCIs	7.5	7.3	6.1	7.3	6.3
		EPT (taxa)	17	14	18	12	10
		%EPT (taxa)	71	64	69	57	48

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 163 Macroinvertebrate fauna of the Kaupokonui Stream: summer SEM survey sampled on 15 February 2013

Taxa List	Site Code	MCI score	KPK000250	KPK000500	KPK000660	KPK000880	KPK000990
	Sample Number		FWB13069	FWB13070	FWB13080	FWB13083	FWB13073
NEMERTEA	Nemertea	3	-	-	A	A	C
ANNELIDA (WORMS)	Oligochaeta	1	-	-	C	C	VA
	Lumbricidae	5	-	-	R	-	R
MOLLUSCA	*Potamopyrgus*	4	-	R	VA	A	A
CRUSTACEA	*Paracalliope*	5	-	-	R	-	-
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	C	A	A	A
	Coloburiscus	7	VA	XA	A	C	-
	Deleatidium	8	XA	XA	XA	XA	R
	Nesameletus	9	VA	VA	C	R	-
PLECOPTERA (STONEFLIES)	*Austroperla*	9	R	-	-	-	-
	Megaleptoperla	9	A	R	R	-	-
	Zelandoperla	8	VA	-	R	-	-
COLEOPTERA (BEETLES)	Elmidae	6	A	VA	VA	A	C
	Hydraenidae	8	R	C	C	-	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	A	VA	A	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	VA	XA	XA	XA
	Costachorema	7	-	-	R	R	R
	Hydrobiosis	5	R	C	C	A	A
	Neurochorema	6	R	R	R	-	-
	Psilochorema	6	-	-	R	-	-
	Beraeoptera	8	C	A	A	-	-
	Helicopsyche	10	R	-	-	-	-
	Olinga	9	VA	C	C	-	-
	Pycnocentrodes	5	R	A	C	A	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA	A	R	R
	Eriopterini	5	R	R	R	-	-
	Maoridiamesa	3	-	R	-	C	C
	Orthocladiinae	2	C	C	R	VA	VA
	Polypedilum	3	R	-	-	-	-
	Tanytarsini	3	-	-	-	R	R
	Muscidae	3	R	-	-	-	R
	Psychodidae	1	R	-	-	-	-
	Austrosimulium	3	-	R	R	-	-
	Tabanidae	3	-	-	R	-	-
	Tanyderidae	4	-	-	-	R	-
		No of taxa	23	20	27	18	17
		MCI	125	120	115	100	92
		SQMCIs	7.9	7.0	5.9	5.6	3.5
		EPT (taxa)	14	11	14	8	6
		%EPT (taxa)	61	55	52	44	35

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 164 Macroinvertebrate fauna of the Katikara Stream: spring SEM survey sampled on 16 November 2012

Taxa List	Site Code / Sample Number	MCI score	KTK000150 / FWB12424	KTK000248 / FWB12425
MOLLUSCA	*Potamopyrgus*	4	-	C
EPHEMEROPTERA (MAYFLIES)	*Ameletopsis*	10	R	-
	Austroclima	7	-	C
	Coloburiscus	7	C	A
	Deleatidium	8	A	XA
	Nesameletus	9	C	C
	Zephlebia group	7	R	R
PLECOPTERA (STONEFLIES)	*Austroperla*	9	C	-
	Megaleptoperla	9	-	R
	Stenoperla	10	R	-
	Zelandobius	5	C	C
	Zelandobius illiesi	10	R	-
	Zelandoperla	8	A	R
COLEOPTERA (BEETLES)	Elmidae	6	R	VA
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	-	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	VA
	Costachorema	7	C	C
	Hydrobiosis	5	-	A
	Hydrobiosella	9	R	-
	Neurochorema	6	-	R
	Orthopsyche	9	R	-
	Beraeoptera	8	-	R
	Pycnocentrodes	5	-	A
DIPTERA (TRUE FLIES)	*Aphrophila*	5	-	A
	Eriopterini	5	R	R
	Orthocladiinae	2	C	R
	Polypedilum	3	-	R
	Tanypodinae	5	R	-
	Empididae	3	R	-
	Austrosimulium	3	-	R
		No of taxa	18	22
		MCI	143	118
		SQMCIs	7.4	6.9
		EPT (taxa)	13	14
		%EPT (taxa)	72	64

'Tolerant' taxa | 'Moderately sensitive' taxa | 'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 165 Macroinvertebrate fauna of the Katikara Stream: summer SEM survey sampled on 4 March 2013

Taxa List	Site Code / Sample Number	MCI score	KTK000150 / FWB13137	KTK000248 / FWB13138
PLATYHELMINTHES (FLATWORMS)	*Cura*	3	-	R
ANNELIDA (WORMS)	Oligochaeta	1	-	C
MOLLUSCA	*Potamopyrgus*	4	-	VA
CRUSTACEA	*Paratya*	3	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R	C
	Coloburiscus	7	A	A
	Deleatidium	8	A	VA
	Nesameletus	9	A	R
	Rallidens	9	-	R
	Zephlebia group	7	R	R
PLECOPTERA (STONEFLIES)	*Austroperla*	9	C	-
	Spaniocerca	8	R	-
	Zelandobius	5	C	R
	Zelandobius illiesi	10	R	-
	Zelandoperla	8	C	-
COLEOPTERA (BEETLES)	Elmidae	6	R	VA
	Hydraenidae	8	C	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	VA
	Hydrobiosis	5	-	VA
	Neurochorema	6	-	C
	Orthopsyche	9	R	-
	Psilochorema	6	R	-
	Beraeoptera	8	-	R
	Oxyethira	2	A	C
	Pycnocentria	7	C	-
	Pycnocentrodes	5	-	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	-	VA
	Eriopterini	5	R	-
	Maoridiamesa	3	R	C
	Orthocladiinae	2	C	VA
	Polypedilum	3	R	-
	Tanytarsini	3	-	C
	Nothodixa	4	R	-
	Empididae	3	R	R
	Ephydridae	4	-	R
	Austrosimulium	3	-	VA
		No of taxa	23	26
		MCI	124	99
		SQMCIs	6.5	4.7
		EPT (taxa)	13	12
		%EPT (taxa)	57	46

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 166 Macroinvertebrate fauna of the Kapoaiaia Stream: spring SEM survey sampled on 4 October 2012

Taxa List	Site Code / Sample Number	MCI score	KPA000250 / FWB12357	KPA000700 / FWB12358	KPA000950 / FWB12359
ANNELIDA (WORMS)	Oligochaeta	1	-	-	A
	Lumbricidae	5	-	-	C
MOLLUSCA	*Potamopyrgus*	4	-	-	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	-	C
	Coloburiscus	7	VA	VA	-
	Deleatidium	8	XA	XA	A
	Nesameletus	9	A	A	R
PLECOPTERA (STONEFLIES)	*Acroperla*	5	A	C	-
	Austroperla	9	R	-	-
	Stenoperla	10	R	-	-
	Zelandobius	5	C	C	C
	Zelandoperla	8	VA	R	-
COLEOPTERA (BEETLES)	Elmidae	6	A	VA	C
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	C	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	R	C	VA
	Costachorema	7	C	C	A
	Hydrobiosis	5	C	R	C
	Psilochorema	6	R	-	-
	Beraeoptera	8	VA	C	-
	Confluens	5	R	-	-
	Helicopsyche	10	A	-	-
	Olinga	9	C	-	-
	Pycnocentria	7	R	-	-
	Pycnocentrodes	5	VA	A	VA
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	R	A
	Eriopterini	5	C	-	-
	Maoridiamesa	3	-	A	VA
	Orthocladiinae	2	R	R	A
	Polypedilum	3	C	-	-
	Empididae	3	-	-	R
		No of taxa	25	16	17
		MCI	130	118	101
		SQMCIs	7.5	7.4	4.3
		EPT (taxa)	19	11	8
		%EPT (taxa)	76	69	47

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare   C = Common   A = Abundant   VA = Very Abundant   XA = Extremely Abundant

Table 167 Macroinvertebrate fauna of the Kapoaiaia Stream: summer SEM survey sampled on 4 March 2013

Taxa List	Site Code / Sample Number	MCI score	KPA000250 / FWB13139	KPA000700 / FWB13140	KPA000950 / FWB13141
NEMERTEA	Nemertea	3	-	-	C
ANNELIDA (WORMS)	Oligochaeta	1	R	R	A
	Lumbricidae	5	-	-	C
MOLLUSCA	*Potamopyrgus*	4	-	R	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	C	A	-
	Coloburiscus	7	A	VA	-
	Deleatidium	8	XA	VA	-
	Nesameletus	9	A	C	-
	Zephlebia group	7	R	-	-
PLECOPTERA (STONEFLIES)	*Austroperla*	9	R	-	-
	Megaleptoperla	9	R	-	-
	Stenoperla	10	R	-	-
	Zelandoperla	8	C	R	-
COLEOPTERA (BEETLES)	Elmidae	6	XA	VA	C
	Hydraenidae	8	R	R	-
	Ptilodactylidae	8	-	-	R
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A	C
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA	VA	VA
	Costachorema	7	C	C	-
	Hydrobiosis	5	C	A	A
	Psilochorema	6	C	-	-
	Beraeoptera	8	C	R	-
	Confluens	5	R	-	-
	Helicopsyche	10	C	-	-
	Olinga	9	R	-	-
	Oxyethira	2	-	R	R
	Pycnocentrodes	5	C	C	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	C	C
	Eriopterini	5	C	-	-
	Maoridiamesa	3	A	VA	C
	Orthocladiinae	2	A	VA	VA
	Polypedilum	3	R	R	-
	Tanytarsini	3	-	A	A
	Empididae	3	-	C	C
	Muscidae	3	R	A	C
	Austrosimulium	3	-	R	C
	Tanyderidae	4	-	R	-
		No of taxa	28	25	18
		MCI	126	100	80
		SQMCIs	6.6	5.0	3.3
		EPT (taxa)	18	10	3
		%EPT (taxa)	64	40	17

'Tolerant' taxa | 'Moderately sensitive' taxa | 'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 168 Macroinvertebrate fauna of the Kurapete Stream: spring SEM survey sampled on 5 October 2012

Taxa List	Site Code	MCI score	KRP000300	KRP000660
	Sample Number		FWB12364	FWB12365
NEMATODA	Nematoda	3	R	-
ANNELIDA (WORMS)	Oligochaeta	1	A	A
	Lumbricidae	5	R	-
MOLLUSCA	*Potamopyrgus*	4	VA	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R	A
	Coloburiscus	7	-	VA
	Deleatidium	8	VA	VA
	Nesameletus	9	-	R
	Zephlebia group	7	VA	A
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R	C
	Zelandobius	5	-	VA
	Zelandoperla	8	-	R
COLEOPTERA (BEETLES)	Elmidae	6	A	A
	Ptilodactylidae	8	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	C	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	C	VA
	Costachorema	7	R	C
	Hydrobiosis	5	C	C
	Hydrobiosella	9	-	C
	Neurochorema	6	-	C
	Orthopsyche	9	R	-
	Beraeoptera	8	R	-
	Oxyethira	2	R	-
	Pycnocentria	7	-	R
	Pycnocentrodes	5	-	A
	Triplectides	5	-	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	A	VA
	Eriopterini	5	R	R
	Maoridiamesa	3	R	C
	Orthocladiinae	2	A	A
	Polypedilum	3	R	-
	Tanypodinae	5	C	-
	Austrosimulium	3	A	R
	Tanyderidae	4	R	-
		No of taxa	25	25
		MCI	101	112
		SQMCIs	5.6	5.6
		EPT (taxa)	9	16
		%EPT (taxa)	36	64

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 169 Macroinvertebrate fauna of the Kurapete Stream: summer SEM survey sampled on 1 February 2013

Taxa List	Site Code	MCI score	KRP000300	KRP000660
	Sample Number		FWB13028	FWB13029
PLATYHELMINTHES (FLATWORMS)	*Cura*	3	R	R
ANNELIDA (WORMS)	Oligochaeta	1	A	A
	Lumbricidae	5	R	-
MOLLUSCA	*Potamopyrgus*	4	VA	C
CRUSTACEA	Paraleptamphopidae	5	R	-
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	C
	Coloburiscus	7	-	A
	Deleatidium	8	VA	VA
	Nesameletus	9	-	R
	Rallidens	9	-	R
	Zephlebia group	7	VA	C
COLEOPTERA (BEETLES)	Elmidae	6	VA	VA
	Hydraenidae	8	R	-
	Ptilodactylidae	8	R	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA	VA
	Costachorema	7	-	C
	Hydrobiosis	5	C	A
	Neurochorema	6	-	C
	Psilochorema	6	R	-
	Oxyethira	2	-	C
	Pycnocentrodes	5	-	C
DIPTERA (TRUE FLIES)	*Aphrophila*	5	R	A
	Maoridiamesa	3	-	R
	Orthocladiinae	2	R	A
	Polypedilum	3	-	R
	Tanypodinae	5	A	C
	Tanytarsini	3	-	C
	Empididae	3	R	C
	Muscidae	3	-	C
	Austrosimulium	3	R	C
	Tanyderidae	4	C	R
		No of taxa	21	27
		MCI	101	97
		SQMCIs	5.7	5.4
		EPT (taxa)	6	11
		%EPT (taxa)	29	41
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa		

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 170 Macroinvertebrate fauna of the Waiokura Stream:spring SEM survey sampled on 4 October 2012

Taxa List	Site Code	MCI score	WKR000500	WKR000700
	Sample Number		FWB12350	FWB12351
PLATYHELMINTHES (FLATWORMS)	*Cura*	3	-	R
ANNELIDA (WORMS)	Oligochaeta	1	C	A
	Lumbricidae	5	R	R
MOLLUSCA	*Potamopyrgus*	4	R	C
CRUSTACEA	*Paracalliope*	5	-	R
	Paraleptamphopidae	5	R	C
	Talitridae	5	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	VA	VA
	Coloburiscus	7	A	C
	Deleatidium	8	A	C
	Zephlebia group	7	C	A
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	C	C
COLEOPTERA (BEETLES)	Elmidae	6	VA	VA
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	A	C
	Costachorema	7	-	C
	Hydrobiosis	5	R	-
	Psilochorema	6	-	R
	Confluens	5	C	R
	Helicopsyche	10	R	-
	Pycnocentria	7	R	C
	Pycnocentrodes	5	VA	VA
DIPTERA (TRUE FLIES)	*Aphrophila*	5	R	R
	Harrisius	6	-	R
	Orthocladiinae	2	-	R
	Tanyderidae	4	-	R
		No of taxa	18	24
		MCI	114	105
		SQMCIs	6.0	5.8
		EPT (taxa)	11	11
		%EPT (taxa)	61	46
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa		

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 171 Macroinvertebrate fauna of the Waiokura Stream: summer SEM survey sampled on 15 February 2013

Taxa List	Site Code	MCI score	WKR000500	WKR000700
	Sample Number		FWB13084	FWB13086
PLATYHELMINTHES (FLATWORMS)	*Cura*	3	-	R
NEMERTEA	Nemertea	3	R	C
NEMATODA	Nematoda	3	-	R
ANNELIDA (WORMS)	Oligochaeta	1	R	A
MOLLUSCA	*Potamopyrgus*	4	VA	R
CRUSTACEA	*Paracalliope*	5	-	C
	Paraleptamphopidae	5	-	R
	Paranephrops	5	R	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	XA	VA
	Coloburiscus	7	A	VA
	Deleatidium	8	VA	R
	Nesameletus	9	R	C
	Zephlebia group	7	VA	XA
PLECOPTERA (STONEFLIES)	*Zelandobius*	5	C	-
COLEOPTERA (BEETLES)	Elmidae	6	XA	VA
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	A
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	VA	VA
	Costachorema	7	-	R
	Hydrobiosis	5	C	C
	Psilochorema	6	R	C
	Beraeoptera	8	R	-
	Confluens	5	-	R
	Hudsonema	6	-	R
	Oeconesidae	5	R	-
	Oxyethira	2	R	-
	Pycnocentria	7	R	A
	Pycnocentrodes	5	C	R
	Zelolessica	7	R	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C	-
	Chironomus	1	-	R
	Harrisius	6	R	-
	Maoridiamesa	3	R	-
	Orthocladiinae	2	R	R
	Polypedilum	3	C	-
	Tanytarsini	3	C	-
	Austrosimulium	3	A	R
	Tanyderidae	4	R	R
		No of taxa	29	27
		MCI	101	100
		SQMCIs	6.2	6.4
		EPT (taxa)	14	13
		%EPT (taxa)	48	48

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 172 Macroinvertebrate fauna of the Tangahoe River: spring SEM survey sampled on 30 October 2012

Taxa List	Site Code	MCI score	TNH000090	TNH000200	TNH000515
	Sample Number		FWB12389	FWB12390	FWB12391
ANNELIDA (WORMS)	Oligochaeta	1	C	C	R
MOLLUSCA	*Potamopyrgus*	4	C	C	C
CRUSTACEA	*Paracalliope*	5	-	-	R
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	A	C
	Coloburiscus	7	-	C	R
	Deleatidium	8	XA	VA	C
	Oniscigaster	10	-	R	-
	Zephlebia group	7	C	C	R
PLECOPTERA (STONEFLIES)	*Acroperla*	5	C	C	R
	Zelandobius	5	R	A	C
COLEOPTERA (BEETLES)	Elmidae	6	A	A	A
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	R	C	-
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	-	R	VA
	Costachorema	7	-	R	C
	Hydrobiosis	5	C	C	C
	Neurochorema	6	-	R	-
	Oxyethira	2	R	-	-
	Pycnocentrodes	5	-	-	A
DIPTERA (TRUE FLIES)	*Aphrophila*	5	R	A	A
	Eriopterini	5	R	R	-
	Hexatomini	5	R	-	-
	Paralimnophila	6	R	-	-
	Harrisius	6	-	R	-
	Maoridiamesa	3	-	R	A
	Orthocladiinae	2	A	A	A
	Polypedilum	3	C	-	-
	Tanytarsini	3	-	C	-
	Empididae	3	R	-	-
	Austrosimulium	3	VA	C	-
	Tanyderidae	4	R	R	R
ACARINA (MITES)	Acarina	5	-	-	R
		No of taxa	20	23	19
		MCI	93	104	100
		SQMCIs	6.8	6.1	4.4
		EPT (taxa)	6	11	10
		%EPT (taxa)	30	48	53

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 173 Macroinvertebrate fauna of the Tangahoe River: summer SEM survey sampled on 7 March 2013

Taxa List	Site Code	MCI score	TNH000090	TNH000200	TNH000515
	Sample Number		FWB13145	FWB13146	FWB13147
NEMERTEA	Nemertea	3	R	-	-
ANNELIDA (WORMS)	Oligochaeta	1	A	C	A
MOLLUSCA	*Latia*	5	-	R	C
	Potamopyrgus	4	VA	VA	C
CRUSTACEA	Ostracoda	1	R	-	-
	Paracalliope	5	-	-	C
	Paranephrops	5	R	-	-
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	A	VA	R
	Coloburiscus	7	-	A	-
	Deleatidium	8	XA	VA	-
	Nesameletus	9	R	-	-
	Oniscigaster	10	R	-	-
	Zephlebia group	7	C	C	-
PLECOPTERA (STONEFLIES)	*Acroperla*	5	-	R	-
	Megaleptoperla	9	A	-	-
	Zelandobius	5	-	C	-
COLEOPTERA (BEETLES)	Elmidae	6	VA	VA	A
	Hydraenidae	8	R	R	-
	Scirtidae	8	R	-	-
	Staphylinidae	5	R	-	-
MEGALOPTERA (DOBSONFLIES)	*Archichauliodes*	7	A	C	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	R	VA	XA
	Costachorema	7	-	R	-
	Hydrobiosis	5	C	C	C
	Psilochorema	6	R	R	-
	Oxyethira	2	-	R	-
	Pycnocentrodes	5	-	R	C
	Triplectides	5	R	-	-
DIPTERA (TRUE FLIES)	*Aphrophila*	5	-	VA	A
	Eriopterini	5	C	-	-
	Hexatomini	5	C	-	-
	Harrisius	6	-	C	-
	Maoridiamesa	3	-	-	R
	Orthocladiinae	2	R	C	VA
	Polypedilum	3	-	R	R
	Tanypodinae	5	R	-	-
	Tanytarsini	3	R	A	C
	Paradixa	4	-	R	-
	Empididae	3	-	C	-
	Muscidae	3	-	R	-
	Austrosimulium	3	C	R	-
	Tanyderidae	4	C	-	-
		No of taxa	27	27	15
		MCI	107	97	87
		SQMCIs	6.9	5.5	3.7
		EPT (taxa)	10	11	4
		%EPT (taxa)	37	41	27

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 174 Macroinvertebrate fauna of the Herekawe Stream: spring SEM survey sampled 5 October 2012

Taxa List	Site Code	MCI score	HRK000085
	Sample Number		FWB12362
ANNELIDA (WORMS)	Oligochaeta	1	A
MOLLUSCA	*Potamopyrgus*	4	VA
CRUSTACEA	*Paracalliope*	5	A
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R
	Coloburiscus	7	A
	Zephlebia group	7	C
PLECOPTERA (STONEFLIES)	*Acroperla*	5	R
	Zelandobius	5	R
COLEOPTERA (BEETLES)	Elmidae	6	R
TRICHOPTERA (CADDISFLIES)	*Hydrobiosis*	5	R
	Orthopsyche	9	R
	Hudsonema	6	R
	Pycnocentrodes	5	R
	Triplectides	5	R
DIPTERA (TRUE FLIES)	*Aphrophila*	5	C
	Eriopterini	5	R
	Orthocladiinae	2	C
	Polypedilum	3	R
	Empididae	3	R
	Tanyderidae	4	R
ACARINA (MITES)	Acarina	5	R
		No of taxa	21
		MCI	99
		SQMCIs	4.2
		EPT (taxa)	10
		%EPT (taxa)	48

'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Table 175 Macroinvertebrate fauna of the Herekawe Stream: summer SEM survey sampled on 12 February 2013

Taxa List	Site Code	MCI score	HRK000085
	Sample Number		FWB13057
ANNELIDA (WORMS)	Oligochaeta	1	A
	Lumbricidae	5	R
MOLLUSCA	Lymnaeidae	3	R
	Potamopyrgus	4	XA
CRUSTACEA	*Paracalliope*	5	XA
EPHEMEROPTERA (MAYFLIES)	*Austroclima*	7	R
	Coloburiscus	7	C
PLECOPTERA (STONEFLIES)	*Megaleptoperla*	9	R
	Zelandobius	5	R
COLEOPTERA (BEETLES)	Elmidae	6	R
TRICHOPTERA (CADDISFLIES)	*Aoteapsyche*	4	R
	Hydrobiosis	5	C
	Orthopsyche	9	R
	Psilochorema	6	R
	Oxyethira	2	VA
	Triplectides	5	R
DIPTERA (TRUE FLIES)	Eriopterini	5	R
	Paralimnophila	6	R
	Chironomus	1	C
	Orthocladiinae	2	C
	Empididae	3	R
	Austrosimulium	3	R
	Tanyderidae	4	R
		No of taxa	23
		MCI	93
		SQMCIs	4.2
		EPT (taxa)	9
		%EPT (taxa)	39
'Tolerant' taxa	'Moderately sensitive' taxa	'Highly sensitive' taxa	

R = Rare C = Common A = Abundant VA = Very Abundant XA = Extremely Abundant

Appendix II

Summary of SEM sites' information, 2012-2013 and historical MCI scores, predicted scores and 1995-2013 trends

Summary of MCI scores at all SEM sites: significance in relation to various predictive methodologies (Stark and Fowles, 2009[1]; Leathwick, 2008[2]), and trends over the SEM period 1995 to 2013

Site code	River Environment Classification (REC)	Altitude (masl)	Distance from National Park (km)	MCI Values											Trends (1995-2013)			Ecological significance
				SEM 1995 to 2013								Predicted			p	p>FDR	+ / -	
				Spring 2012	Summer 2013	1995-2013				Median stream 'health' category		Altitude[1]	Distance[1]	REC[2]				
						Range	Medians Spring	Medians Summer	Medians Overall	Generic[3]	Predicted [4]							
STY000300	CX/H/VA/S/MO/MG	160	7.3	(112)	(106)	64-160	112	112	113	Good	Expected	101[+]	109[0]	128[-]	0.155	0.239	-ve	-
STY000400	CX/H/VA/S/MO/MG	70	12.5	(105)	(127)	0-160	107	109	108	Good	Better than	92[+]	103[0]	115[0]	0.608	0.702	-ve	-
TMR000150	CX/H/VA/IF/LO/HG	420	0	132	138	119-144	135	137	136	Very good	Expected	127[0]	132[0]	141[0]	0.632	0.709	-ve	-
TMR000375	CX/L/VA/P/MO/MG	100	10.9	114	103	89-120	103	102	103	Good	Expected	95[0]	105[0]	117[-]	<0.0001	<0.001	+ve	Yes
MRK000420	WW/L/VA/P/MO/LG	60	N/A	101	92	75-105	93	89	90	Fair	Expected	91[0]	N/A	92[0]	<0.0001	<0.001	+ve	Yes
WGA000260	CX/L/VA/P/MO/LG	140	16.1	110	98	82-112	97	94	96	Fair	Expected	99[0]	100[0]	99[0]	0.037	0.071	+ve	(Yes)
WGA000450	WW/L/VA/P/MO/LG	20	31.2	98	87	72-102	93	84	88	Fair	Expected	86[0]	93[0]	88[0]	0.0002	<0.001	+ve	Yes
WKH000100	CX/H/VA/IF/LO/HG	460	0	133	125	115-147	133	125	129	Very good	Expected	131[0]	132[0]	137[0]	0.511	0.604	+ve	-
WKH000500	CX/H/VA/P/MO/MG	175	10.6	112	112	87-122	111	106	109	Good	Expected	102[0]	105[0]	115[0]	0.003	0.007	+ve	Yes
WKH000920	CX/H/VA/P/HO/LG	20	26.6	108	96	71-110	101	92	95	Fair	Expected	86[0]	95[0]	97[0]	0.103	0.168	+ve	-
WKH000950	CX/H/VA/P/HO/LG	2	28.4	96	89	70-111	92	84	88	Fair	Expected	85[0]	94[0]	97[0]	0.161	0.239	+ve	-
MGE000970	CX/L/VA/P/MO/LG	90	15.6	97	102	86-113	106	99	103	Good	Expected	94[0]	101(0)	101[0]	0.457	0.552	-ve	-
MGN000195	CX/H/VA/P/MO/LG	330	8.7	135	139	113-143	130	123	126	Very good	Better than	118[0]	107[+]	124[0]	0.222	0.306	-ve	-
MGN000427	CX/L/VA/P/HO/MG	140	37.9	111	96	77-115	102	96	98	Fair	Expected	99[0]	91[0]	103[0]	0.398	0.524	+ve	-
MKW000200	CX/H/VA/IF/MO/MG	380	2.3	139	128	100-141	132	125	128	Very good	Expected	123[0]	121[0]	130[0]	0.641	0.709	-ve	-
MKW000300	CX/H/VA/P/MO/LG	150	15.5	116	117	90-115	107	100	104	Good	Expected	100[0]	101[0]	111[0]	0.061	0.106	+ve	-
WTR000850	WX/L/SS/P/HO/LG	15	N/A	107	83	64-107	91	78	86	Fair	Expected	85[0]	N/A	98[-]	0.019	0.041	+ve	Yes
MGT000488	WN/L/VA/P/LO/LG	30	N/A	86	91	56-91	77	79	78	N/A	N/A[s]	N/A	N/A	80[0]	0.407	0.527	+ve	-
MGT000520	WW/L/VA/U/LO/LG	20	N/A	75	75	44-78	63	65	64	N/A	N/A[s]	N/A	N/A	88[-]	<0.0001	<0.001	+ve	Yes
WMK000100	WW/L/VA/P/LO/HG	160	0	129	124	121-141	133	130	131	Very good	Expected	101[+]	132[0]	128[0]	0.685	0.712	+ve	-
WMK000298	WW/L/VA/P/MO/MG	1	4.0	95	87	75-101	93	88	89	Fair	Expected	85[0]	116[-]	103[-]	0.051	0.092	+ve	-
WAI000110	WW/L/VA/P/MO/LG	50	N/A	100	90	80-100	92	88	90	Fair	N/A	N/A	N/A	91[0]	0.004	0.009	+ve	Yes
PNH000200	CX/H/YA/IF/MO/MG	270	4.4	134	119	104-134	127	119	122	Very good	Expected	112[0]	115[0]	121[0]	0.031	0.060	+ve	(Yes)
PNH000900	CW/L/VA/P/MO/LG	20	20.9	106	85	70-106	91	83	87	Fair	Expected	86[0]	98[-]	100[-]	0.0004	0.001	+ve	Yes
PAT000200	CX/H/VA/IF/MO/MG	500	1.9	138	141	127-145	138	137	138	Very good	Expected	135[0]	125[+]	129[0]	0.673	0.712	+ve	-
PAT000315	CX/H/VA/P/MO/LG	300	12.9	118	106	99-130	115	106	110	Good	Expected	115[0]	103[0]	112[0]	0.413	0.524	+ve	-
PAT000360	CW/L/VA/P/HO/LG	240	19.2	100	95	86-105	100	97	98	Fair	Expected	109[-]	99[0]	109[-]	0.219	0.306	+ve	-
MGH000950	CW/L/SS/P/HO/LG	120	N/A	101	91	77-104	93	88	90	Fair	N/A	N/A	N/A	117[-]	<0.0001	<0.0001	+ve	Yes
WGG000115	CX/H/VA/IF/LO/MG	540	0.7	130	132	122-139	132	134	132	Very good	Expected	140[0]	130[0]	131[0]	0.224	0.306	+ve	-
WGG000150	CX/H/VA/P/LO/MG	380	7.2	134	130	119-139	132	128	130	Very good	Better than	123[0]	110[+]	124[0]	0.717	0.731	+ve	-
WGG000500	CW/L/VA/P/MO/LG	200	23.0	124	112	91-124	103	101	102	Good	Expected	105[0]	97[0]	110[0]	0.0001	<0.001	+ve	Yes
WGG000665	CW/L/VA/P/HO/MG	180	29.6	108	105	77-111	98	93	94	Fair	Expected	103[0]	94[0]	102[0]	0.002	0.005	+ve	Yes
WGG000895	CW/L/VA/P/HO/LG	40	63.0	106	96	73-106	96	92	95	Fair	Expected	89[0]	85[0]	92[0]	0.198	0.164	+ve	-
WGG000995	CW/L/VA/P/HO/MG	5	66.6	100	83	69-100	93	85	91	Fair	Expected	85[0]	85[0]	95[0]	0.005	0.011	+ve	Yes
MWH000380	WW/L/M/P/MO/LG	200	N/A	85	85	58-85	76	73	75	N/A	N/A[s]	N/A	N/A	92[-]	0.002	0.004	+ve	No
MWH000490	CN/L/VA/P/MO/LG	190	N/A	102	90	63-102	80	76	78	Poor	Worse than	104[-]	N/A	93[-]	<0.0005	0.001	+ve	Yes
HTK000350	WX/L/VA/P/MO/LG	60	N/A	114	101	79-114	96	94	95	Fair	Expected	91[0]	N/A	95[0]	<0.0001	<0.001	+ve	Yes
HTK000425	WW/L/VA/P/MO/LG	30	N/A	115	105	91-115	105	102	103	Good	Better than	87[+]	N/A	92[+]	0.0002	<0.001	+ve	Yes
HTK000745	WW/L/VA/U/MO/MG	5	N/A	101	88	69-101	85	86	86	Fair	Expected	85[0]	N/A	93[0]	0.672	0.712	+ve	-
KPK000250	CX/H/VA/IF/MO/MG	380	3.3	128	125	125-138	130	128	129	Very good	Expected	123[0]	118[+]	137[0]	0.907	0.907	-ve	-
KPK000500	CX/H/VA/P/MO/MG	260	9.2	128	120	98-133	121	112	116	Good	Expected	111[0]	107[0]	127[-]	0.0001	<0.001	+ve	Yes
KPK000660	CX/H/VA/P/MO/LG	170	15.5	128	115	71-128	104	102	103	Good	Expected	102[0]	101[0]	122[-]	<0.0001	<0.0001	+ve	Yes
KPK000880	CW/H/VA/P/MO/LG	60	25.7	110	100	66-110	92	90	91	Fair	Expected	91[0]	95[0]	106[-]	0.001	0.004	+ve	Yes
KPK000990	CW/L/VA/P/HO/LG	5	31.1	103	92	69-98	93	88	91	Fair	Expected	85[0]	93[0]	96[0]	0.008	0.019	+ve	Yes
KTK000150	CW/L/VA/P/HO/LG	420	0	143	124	112-148	138	136	137	Very good	Expected	127[0]	132[0]	131[0]	0.432	0.535	-ve	-
KTK000248	WX/L/VA/P/MO/LG	5	18.1	118	99	81-118	102	103	102	Good	Expected	85[+]	99[0]	96[0]	0.024	0.048	+ve	Yes
KPA000250	CX/H/VA/P/MO/MG	240	5.7	130	126	83-130	111	109	110	Good	Expected	109[0]	112[0]	111[0]	<0.0001	<0.0001	+ve	Yes
KPA000700	CX/H/VA/P/MO/MG	140	13.5	118	100	78-118	95	93	94	Fair	Expected	99[0]	103[0]	105[-]	0.0002	<0.001	+ve	Yes
KPA000950	CX/L/VA/P/MO/LG	20	25.2	101	80	76-101	89	80	85	Fair	Expected	86[0]	96[-]	99[-]	0.160	0.289	+ve	-
KRP000300	WX/L/VA/P/LO/LG	180	N/A	101	101	80-103	93	95	94	Fair	Expected	103[0]	N/A	92[0]	<0.0001	<0.0001	+ve	Yes
KRP000660	WW/L/VA/P/LO/LG	120	N/A	112	97	70-112	95	91	91	Fair	Expected	97[0]	N/A	102[-]	<0.001	0.001	+ve	Yes
WKR000500	WW/L/VA/P/MO/LG	150	N/A	114	101	88-114	102	95	98	Fair	Expected	100[0]	N/A	97[0]	N/A	N/A	+ve	N/A
WKR000700	WW/L/VA/P/MO/LG	70	N/A	105	100	92-105	99	100	100	Good	Expected	92[0]	N/A	95[0]	N/A	N/A	+ve	N/A
TNH000090	WW/L/SS/P/MO/LG	85	N/A	93	107	90-107	95	99	97	Fair	N/A	N/A	N/A	110[0]	N/A	N/A	-ve	N/A
TNH000200	WW/L/SS/P/HO/LG	65	N/A	104	97	92-108	105	100	104	Good	N/A	N/A	N/A	108[0]	N/A	N/A	-ve	N/A
TNH000515	WW/L/SS/P/HO/LG	15	N/A	100	87	84-104	97	87	92	Fair	N/A	N/A	N/A	95[0]	N/A	N/A	+ve	N/A
HRK000085	WW/L/VA/U/MO/MG	5	N/A	99	93	68-99	89	86	89	Fair	N/A	N/A	N/A	89[0]	0.024	0.048	+ve	Yes

Notes: ( ) = affected by headwater erosion events; [+/-] = median score ecologically significant deviation from predicted scores; Trend significant/not significant at p<0.05; N/A = non-ringplain sites; N/A[s] = soft-bedded sites
3 =TRC generic health categories (Table 1), 4 = TRC predictive ringplain health categories (Table 2).